Consortium Members































D4.1: Product Validation and Intercomparison Report

Reference

CCI-LAKES-0009-PVIR

Issue

2.1

Date

6 July 2022





Version h	istory:		
Issue:	Date:	Reason for change:	Author
1.0	30 March 2020	Initial Version	PML: Xiaohan Liu, Stefan Simis
		CLS: Beatriz Calmettes	
			H2O: Claude Duguay
			SERTIT: Hervé Yésou
1.1	30 April 2020	Adding Chapter LSWT and minor	UoR: Chris Merchant
		corrections	CLS: B Calmettes / B Coulon
1.2	30 Sept. 2020	Revision following ESA Review	
			ICUBE-SERTIT : H. Yésou
		Adding results and minor corrections in LWE part	NORCE : E. Malmes
			TRE-Altamira : P.Blanco
		Legos : JF Crétaux	
			CLS: B Calmettes / B Coulon
1.3	4 Oct. 2021	Accompanying CRDP v1.1 (direct product validation) and evaluating CRDP v1.0 and v1.1	C. Duguay
1.4	11 Nov. 2021	Revision following ESA Review	B. Calmettes
2.0	22 Apr. 2022	Accompanying CRDP v2.0 (direct product validation) and evaluating	PML: Xiaohan Liu, Stefan Simis
		CRDP v1.0 and v1.1 (user feedback, use cases)	CLS: Beatriz Calmettes
			H2O: Claude Duguay
			UoR: Chris Merchant/Laura Carrea
			Legos: JF Crétaux
2.1	6 July 2022	Revision following ESA Review	B. Calmettes

People involv	ed in this issue:	Signature		
Authors:	Stefan Simis,	Plymouth	Marine	
	Xiaohan Liu	Laboratory		
	B. Calmettes	CLS		
	Hervé Yésou	SERTIT		
	Chris Merchant, Laura Carrea	University Reading	of	
	Claude Duguay Yuhao Wu	H20 Geomat	ics	
Internal review:	Stefan Simis	Plymouth Laboratory	Marine	
	B. Calmettes	CLS		
Approved by:	B. Coulon	CLS		
Authorized by:	C. Albergel	ESA		

Distribution				
Company	Names	Contact Details		
ESA	C. Albergel	Clement.Albergel@esa.int		
ВС	K. Stelzer	kerstin.stelzer@brockmann-consult.de		
	J. Scholze	jorrit.sholze@brockmann-consult.de		
CLS	B. Coulon	bcoulon@groupcls.com		
	B. Calmettes	bcalmettes@groupcls.com		
	P. Thibaut	pthibaut@groupcls.com		
	A . Mangilli	amangilli@groupcls.com		
CNR	C. Giardino	giardino.c@irea.cnr.it		
	G. Free	free.g@irea.cnr.it		
	M. Pinardi	pinardi.m@irea.cnr.it		
	M. Bresciani	bresciani.m@irea.cnr.it		
GeoEcoMar	A. Scrieciu	albert.scrieciu@geoecomar.ro		
	A. Constantinescu	adriana.c@geoecomar.ro		
H2OG	C. Duguay	claude.duguay@h2ogeomatics.com		
LEGOS	J.F. Crétaux	jean-francois.cretaux@legos.obs-mip.fr		
	A Kouraev	alexei.kouraev@legos.obs-mip.fr		
NORCE	E. Malnes	eirik.malnes@norcesearch.no		
PML	S. Simis	stsi@pml.ac.uk		
	X. Liu	liux@pml.ac.uk		
SERTIT	H. Yésou	herve.yesou@unsitra.fr		
TRE-ALTAMIRA	P. Blanco	Pablo.blanco@tre-altamira.com		
UoR	C. Merchant	c.j.merchant@reading.ac.uk		
	L. Carrea	l.carrea@reading.ac.uk		
UoS	A. Tyler	a.n.tyler@stir.ac.uk		
	E. Spyrakos	evangelos.spyrakos@stir.ac.uk		
	D. Jiang	dalin.jiang@stir.ac.uk		
EOLA	E. Zakharova	zavocado@gmail.com		

List of Contents

1.	Overview	8
2.	Lake Water Level (LWL)	9
	2.1. Data description	
	2.2. Comparison methods	
	2.2.1. Comparison with in situ data	9
	2.2.2. Dedicated field work	. 10
	2.3. Description of work	10
	2.4. Result analysis	11
	2.4.1. Comparison with Hydrolare	. 11
	2.4.2. Comparison to Hidricos Argentina	. 13
	2.4.3. Comparison to USGS	. 15
	2.4.4. Comparison to Canadian Water office	. 17
	2.4.5. Comparison to Office Fédéral de l'Environnement (OFE/Switzerland)	. 18
	2.4.6. Field work experiments	. 19
	2.5. Conclusions and recommendations	20
3.	Lake Water Extent (LWE)	22
	3.1. Data description	22
	3.1.1. Optical Data	. 22
	3.1.2. SAR data	. 24
	3.1.3. Exogeneous database exploited as inputs	. 24
	3.2. Comparison of methods	26
	3.2.1. Comparison of methods for optical sensors (VIS_NIR_SWIR)	. 26
	3.2.2. Validation of LWE derived from HR optical sensors based on VHR sensors	. 28
	3.3. CCI test sites result analysis	33
	3.3.1. Results analysis for optical sensors-based approaches	. 33
	3.3.2. Results analysis for SAR sensors-based approaches	. 59
	3.3.3. Comparison of optical and SAR sensors-based approaches	. 71
	3.3.4. Results analysis for hypsometric approach	. 86
	3.4. Conclusions and recommendations	94
4.	Lake Surface Water Temperature (LSWT)	96
	4.1. Data description	96

	4.1.1. Satellite data	96
	4.1.2. In situ data	97
	4.2. Comparison methods	104
	4.2.1. Generation of the L2 matchups	104
	4.2.2. Validation of the L3S CCI LSWT v4.3	105
	4.2.3. Validation of the LSWT uncertainty	105
	4.2.4. Number of CCI lakes with LSWT	105
	4.3. Description of work	105
	4.4. Result analysis	107
	4.4.1. Validation of LSWTs	107
	4.4.2. Validation of the uncertainty LSWT v4.3	117
	4.4.3. Lakes with no LSWT	119
	4.5. Conclusions and recommendations	120
	4.6. Acknowledgement	120
5.	. Lake Ice Cover (LIC)	122
	5.1 Data description	
	5.1.1. MODIS Terra/Aqua Calibrated Radiances Level 1B product (MOD02/MYD02).	122
	5.1.2. MODIS Snow Cover product (MOD10/MYD10)	122
	5.2 Comparison methods	123
	5.3 Description of work	123
	5.4 Result analysis	125
	5.5 Conclusions and recommendations	131
6.	. Lake Water-Leaving Reflectance (LWLR)	133
	6.1. Data description	133
	6.1.1. In situ data	133
	6.1.2. Satellite data	133
	6.2. Comparison methods	134
	6.3. Description of work	134
	6.3.1. Lake Water-Leaving Reflectance	134
	6.3.2. Chlorophyll-a and TSM	134
	6.3.3. Inter-sensor consistency evaluations	138
	6.4. Validation results	139
	6.4.1. Validation results for MERIS	139
	6.4.1.1. LWLR from MERIS	139
	6.4.2. Validation results for MODIS	144
	6.4.3. Inter-sensor consistency evaluation	152

6.5. Conclusions and recommendations	155
7. References	157
Appendix A. Hydrolare LWL Comparison	161
Appendix B. Water Office Canada Comparison	166

Reference documents

- RD- 1: Algorithm Theoretical Basis Document CCI-LAKES-0024-ATBD_v3.1
- RD- 2: Product Validation Plan CCI-LAKES-0030-PVP_V1.3
- RD- 3: Product User Guide CCI-LAKES-0029-PUG_v2.1

1. Overview

This document contains product validation results accompanying the second major release of the the Lakes_cci dataset (V2.0). The major change in this release is the inclusion of new Lake Ice Cover algorithms, a partial filling of the 2012-2016 temporal gap in LWRL, and the expansion of the dataset to 2000+ waterbodies. The Lakes_cci project has three stages of product validation:

- Validation of individual thematic variables based on direct comparison between remote sensing products and in situ data or other remoting sensing datasets
- Consistency between these variables through five use cases
- Feedback from users of the data set

The purpose of this document is to summarize the results of the different thematic products (LWL, LWE, LSWT, LIC and LWLR) under the first activity, as described in the Product Validation Plan (RD- 2). Feedback from users and use cases from earlier data releases (v1.0 and v1.1) will have been adopted where possible, and remains very welcome.

2. Lake Water Level (LWL)

2.1. Data description

Lake Water Level is the measure of the absolute height of the reflecting water surface beneath the satellite with respect to a vertical datum (geoid) and expressed in metres. The time series has been computed from multiple altimetry satellites since late 1992 to 2018 inclusive. The time periods used for each satellite/instrument are provided in Table 1 but may vary from one lake to the other, depending on the orbits of the satellites with respect to the location of the lake All missions has been used when the data were available and valid, however during tandem overlapping phase (TOPEX/Poseidon/Jason1, Jason1/jason2, Sentinel3A/Sentinel3B) we always test what is the most precise solution before choosing the data used in the time series. Therefore, from one lake to another one, the month of the year used for a given mission may change.

Table 1. Time periods for the satellite/instrument used to generate the lake product

Satellite	Instrument	Time Period
TOPEX/Poseidon (T/P)	Poseidon-1	08/1992 - 01/2002
ERS-2	RA	04/1995-06/2003
Jason-1	Poseidon-2	12/2001 - 07/2013
Jason-2	Poseidon-3	06/2008 - 10/2016
Jason-3	Poseidon-3B	01/2016 - present
Envisat	Radar Altimeter (RA-2)	Mission: 03/2002 - 10/2010
Cryosat-2	SAR interferometric Radar Altimeter (Siral)	04/2010-2015
SARAL	AltiKa	02/2013 - 07/2016
Sentinel-3a	SRAL	02/2016 - present
Sentinel-3b	SRAL	01/2019 - present

A detailed description of the product generation is provided in the Algorithm Theoretical Basis Document (ATBD) with further information on the product given in the Product User Guide (PUG)

2.2. Comparison methods

2.2.1. Comparison with in situ data

External in situ data are useful to assess the quality of the LWL products. The comparison with these products, using different datums and different dates, is not straight forward.

However, it provides information on the product precision and accuracy. The list of datasets used is provided in Table 2

Table 2. In situ databases for LWL validation

Dataset name	Description
<u>Hydrolare</u>	The International Data Centre on Hydrology of Lakes And Reservoirs provides data on mean monthly water level of nearly 1200 water bodies. The Centre operates under the auspices of WMO and a detailed protocol developed by the International Steering Committee of the Centre and agreed by WMO.
Hidricos Argentina	The database base of Hidricos Argentina provides in-situ data on national rivers and lakes.
<u>USGS</u>	The database of US Geological Survey provides in situ data
US Army	The Army Corps of Engineer provides in-situ data on Great Lakes. All levels are referenced to the International Great Lakes Datum of 1985 (IGLD 85
Water Office Canada	This database contains in situ historical hydrometric data in Canada.
Federal Office for the Environment (FOEN)	This database contains in situ historical hydrometric data in Switzerland.

2.2.2. Dedicated field work

The second comparison method is the base on 18 years of the dedicated field work in the framework of satellite altimetry cal/val programmes over lake Issykkul.

2.3. Description of work

About the situ comparison method, interpolation of LWL product to the dates of in situ measurements are first performed, then the mean bias between in situ and satellite timeseries is calculated. A bias is always detected since satellite time series and in situ measurements are never given using the same geodetic reference frame. Some results of these comparison are given in Cretaux et al. (2016) and Ričko et al. (2012).

If in such analysis the in-situ data are considered as ground truth, there is still difficulties to directly compared them since physical properties and scaling of the measurements are fully different. Satellite measurements are collected along the track of the satellite with a footprint of several square kilometres, while the in-situ measurements are generally done on the lakeshore very often far from the satellite track by several kilometres. The reference system of each of the measurements are moreover different, and uncertainty on the geoid over the lake adds a source of bias between both types of measurement. Moreover, technical issues with gauges, data gaps, human error in collecting the data, can

easily increase the gauge error level to few centimeters, making them an unreliable source of ground truth for small variations of lake level.

The inter-comparison between the two types of dataset is also complicated because the in-situ measurements have generally discreteness to monthly average values, while the frequency of the satellite flight over the lake is fully determined by its orbit. This leads us to interpolate, which can also be a source of error, especially when the lake water level variations are brutal, or in presence of seiche for example

The errors of satellite altimetry over lakes depend on several factors. Depending on the size and shape of lakeshore, the altimetry telemetry waveform which is analysed for the calculation of water level over the footprint can be more complex than the usual shape of such signal from the open surface over the oceans, or over very large lakes. Footprint over narrow reservoirs for example, covers inhomogeneous surface over the lake, with non-water surface such as vegetation, bare soil, sandbanks, ice. This explains the variation from few centimetres to few decimetres in altimeter performance over a large set of lakes. Moreover, evolution of the altimetry technique from the Low-Resolution-Mode altimeters on the Jason/OSTM series to the SAR altimetry used with the sentinel-3 series generates improvement of accuracy from the oldest time series to the newest one. For example, over the Issykkul or the Illmen lakes, we have obtained a gain of a factor 2 once the sentinel-3 data were used.

Drift can subsequently be adjusted if it is observed. Root-mean-square differences of unbiased time series are calculated, for the complete time series and for the Jason 3 and Sentinel 3A missions.

Concerning the dedicated field work, it is based on 15 years experiments over the lake Issykkul in Centreal Asia. This large lake (6000 km²) was selected in 2004 to serve as a dedicated calibration / validation site for satellite altimetry over lakes. It has the advantage of overpasses by all past, present and future altimetry missions. The instrumental concept for the field work is widely described in several publications (Cretaux et al. 2009, 2011, 2013, 2018, Bonnefond et al. 2018). In brief, the field work is organised yearly or bi-yearly after consulting the ephemerides of the satellites. GPS levelling of the lake surface is performed along the satellite tracks using a GPS system. In situ fixed instrumentation allows to assess the stability of the LWL product, and also to validate the atmospheric and geodetic corrections. The main purpose is to perform full error budget analysis including the range measurements using different retracking algorithms (so called ice-1, Ice-2, ocean) and also the different corrections (ionosphere, troposphere, geoid).

2.4. Result analysis

2.4.1. Comparison with Hydrolare

Thanks to the collaboration with the International Data Centre on Hydrology of Lakes and Reservoirs, Hydrolare, the information on ten lakes was provided in a monthly time step. Two indicators were estimated: Bias and RMS (Table 3). These values were estimated for each lake for the full time series of nearly 30 years including data from all missions. Since 2016, the indicators were separately evaluated for Jason 3 and Sentinel 3A missions for certain lakes. In some cases (Issykkul for example), the bias is high (up to 2.3m) which is

due to dery different reference system.. We are therefore more interested in the variability (Appendix A) and the correlation indicated by the Pearson coefficient. Appendix A contains the figures corresponding to some of these lakes.

Table 3. Hydrolare LWL comparison

Lake Name Multi satellite Jason 3 Sentinel 3A						A	
			(since20	16)	(since 201		
	Time period	Bias (cm)	RMS(m)	Bias (cm)	RMS(m)	Bias (cm)	RMS (m)
Baikal	1992/09 - 2019/12	0,975	0,11	-	-	-	-
Bratskoye	1992/09 - 2019/12	74,302	0,30	-	-	-	-
Caspian	1992/09 - 2021/09	30.845	0.062	36.020	0.052	34.700	0.050
Issykkul	1992/09 - 2017/12	-232.984	0.045	-231.098	0.017	-231.252	0.017
Khanka	2000/01 - 2018/12	101.989	0.19	-	-	-	-
Kuybyshevs koye	1992/09 - 2018/12	26.550	0.228	29.802	0.082	23.132	0.207
Ladoga	1992/09 - 2018/12	-3.727	0.053	-6.217	0.026	-6.387	0.025
Onega	1992/10 - 2018/12	39.7111	0.059	34.536	0.038	34.837	0.042
Rybinskoye	1992/09 - 2014/12	10.122	0.178	-	-	-	-
Illmen	1996/01- 2019/12	-	35	-	-	-	15
Krasnoyarsk	2003/01- 2008/12	-	37	-	-	-	-
Kubensk	2016/01- 2019/12	-	27	-	-	-	-
Kumsk	2009/01- 2019/12	-	12	-	-	-	-
Lacha	2009/01- 2019/12	-	20	-	-	-	-
Segozersk	2009/01- 2019/12		12	-	-	-	-
Syamozero	2016/01- 2019/12		10	-	-	-	-
Tsimlyansk	1993/01- 2019/12		43	-	-	-	-

Lake Name	e Name Multi satellite		Jason 3		Sentinel 3A		
	l l		(since2016)		(since 2016)		
	Time period	Bias (cm)	RMS(m)	Bias (cm)	RMS(m)	Bias (cm)	RMS (m)
Ust-Ilimsk	2016/01- 2019/12		15	-	-	-	-
Verkhnee- Kuyto	2009/01- 2019/12		16	-	-	-	-
vodlozrero	2019/01- 2019/12		8	-	-	1	1
Superior	1992/09 - 2017/12	-59.093	0.042	-62.369	0.011	-62.495	0.013

The rms value for both missions Jason 3 and Sentinel 3B is very similar. This value is lower for the last two missions than for the overall period, indicating a better estimation of the LWL for the current missions

2.4.2. Comparison to Hidricos Argentina

The information concerning the variation on the Water Lever for lake Argentino and lake General Carrera were obtained online from the Base de datos Hidrologica integrada (BDHI): bdhi.hidricosargentina.gob.ar. For those lakes two indicators were evaluated (Table 4): the RMS of the variations and Pearson coefficient, indicating the correlation between time series.

Table 4. Hidricos Argentina LWL Comparison

Lake Name	Time period	RMS(m)	Pearson
Argentino	1992/10 - 2019/12	0.159	0.969
General Carrera	2008/09 - 2014/10	0.387	0.582

Since for General Carrea the comparison period ends before 2016, analysis of performance in current missions is not possible. Concerning Argentino lake, the Pearson coefficient is 0.995 and 0.991 for Jason 3 and Sentinel 3A missions respectively showing a very strong correlation between the time series. This is also indicated by the low value of the RMS for missions: 0.063m and 0.079m for Jason 3 and Sentinel 3A

Figure 1 and Figure 2 show the comparison of the LWL variation from CCI lakes and Hidricos Argentina for lakes Argentino and General Carrera respectively. For Argentino lake, there is a very good correlation between time series, also indicated by a Pearson coefficient near to 1. However, for General Carrera lake, this correlation is less strong. The General Carrea shows a high variability.

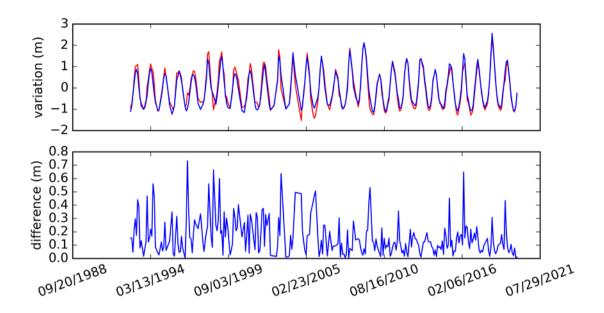


Figure 1. Hidricos Argentina comparison for the lake Argentino (red: Lakes_cci, blue: Hidricos Argentina)

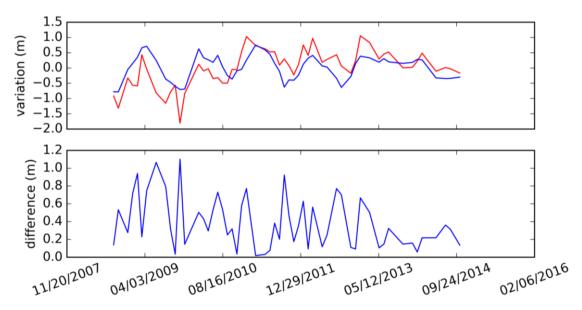


Figure 2. Hidricos Argentina comparison for the lake General Carrera ((red: Lakes_cci, blue: Hidricos Argentina))

2.4.3. Comparison to USGS

The information concerning the variation on the Water Lever for lakes Michigan and Des Bois were obtained online from the US Geological Survey database (USGS). For those lakes, as for the previous comparisons the RMS of the variations and Pearson coefficient were evaluated (Table 5). Figure 3 and Figure 4 show the comparison of the LWL variation from CCI lakes and UGS.

Table 5. USGS LWL Comparison

Lake Name	Time period	RMS(m)	Pearson
Des Bois	1992/09 - 2019/02	0.217	0.668
Michigan	1997/09 - 2019/10	0.084	0.965

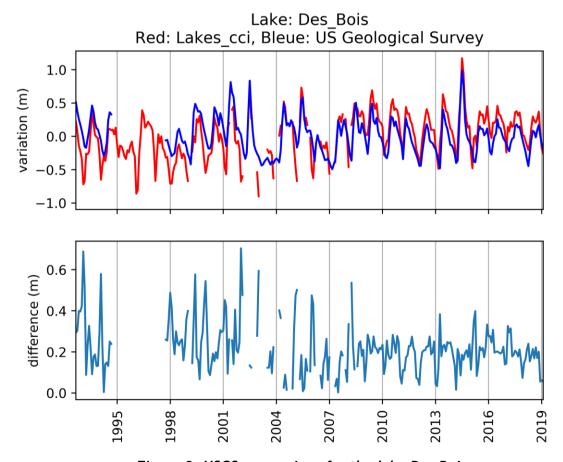


Figure 3. USGS comparison for the lake Des Bois

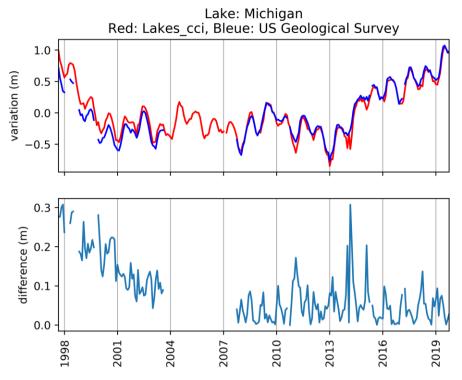


Figure 4. USGS comparison for the lake Michigan

It should be noticed that although in situ data are generally considered as the truth, which is valid most of the time, they may also present severe limitations. Some periods of time are not covered at all with in situ data. Some human errors in the data collection are also happening sometimes as we see it with the lake Onega. In other case like for the lake Michigan or the Caspian Sea, it exists several in situ instrumentations that provide different values of LWL. Sometimes this can be easily explained by local effect at high frequency (like the Seiche effect) sometimes it is less understandable. For example, we see that with two sources for the lake Michigan, the US army corps (Figure 5), and the USGS, the LWL present drifts and systematic disagreements. We can see with the USGS data between 1998 and 2004 when compared to the data of US army. In the first case the comparison with the altimetry shows big disagreements while in the second case the correlation and the RMS are much better.

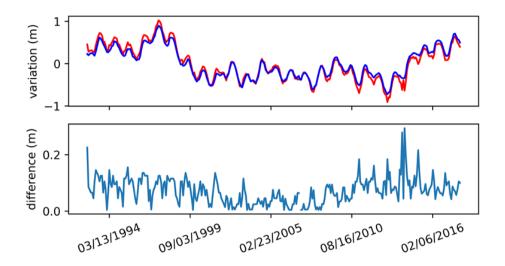


Figure 5. US Army comparison for the lake Michigan (red: Lakes_cci, blue: Hidricos Argentina)

2.4.4. Comparison to Canadian Water office

The Water Office of Canada contains information on the water levels and flood for multiple Canadian lakes and rivers at several time resolutions. Monthly historical data for 19 lakes included in first version of the CCI lakes dataset were obtained on-line: https://wateroffice.ec.gc.ca/mainmenu/historical data index e.html.

The three indicators used with the previous in situ datasets: Bias, RMSE and Pearson Coefficient were evaluated. Figure 6 and Figure 7 show respectively the Pearson coefficient and RMS value for the 19 lakes compared. Most of them have a high Pearson coefficient showing a good time series correlation. Appendix B contains the figures of time series, variation and unbiased absolute difference for each lake. For the lakes with low value of Pearson coefficient, there a variety of reasons for this:

- In some cases, as for the lakes Aylmer or Caribou, there is small amount of in situ data
- In other case, as for the lakes Great Slave or Williston, the altimetric level value couldn't be estimated
- Some outliers, as for the Ontario lake, will affect the correlation between time series

In these cases, the low Person coefficient does not actually represent a poor correlation between time series.

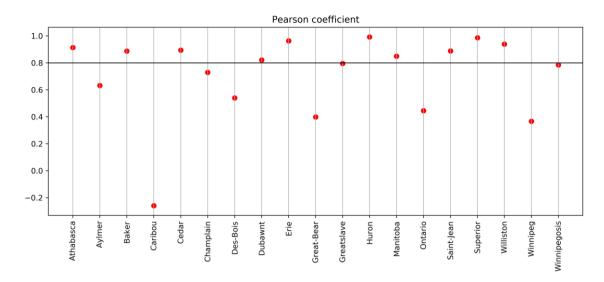


Figure 6. Pearson coefficient CCI Lakes - Water Office Canada.

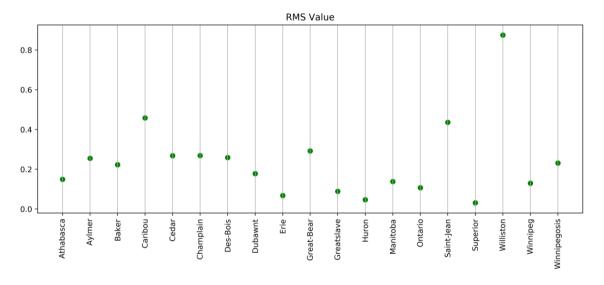


Figure 7. RMS Value. CCI Lakes - Water Office of Canada

2.4.5. Comparison to Office Fédéral de l'Environnement (OFE/Switzerland)

A series of 5 years (2016-2020) of in situ LWL over a set of mountain lakes located in Switzerland has been acquired from the web site: https://www.hydrodaten.admin.ch/. The results are summarized in Table 6. It shows that only using sentinel-3A and sentinel-3B (which was done for these lakes) the LWL product is very accurate. Except fo the Bodensee, RMS are always below 10 cm.

Table 6. FOEN LWL Compa

Lake Name	Time period	RMS(m)	Pearson
Geneva	2016-04/2020/07	4	0.998
Neuchatel	2016-04/2020/07	6.4	0.914
Bodensee	2016-04/2020/07	18	0.945
Luzern	2016-04/2020/07	8	
Zürich	2016-04/2020/07	7.5	0.874

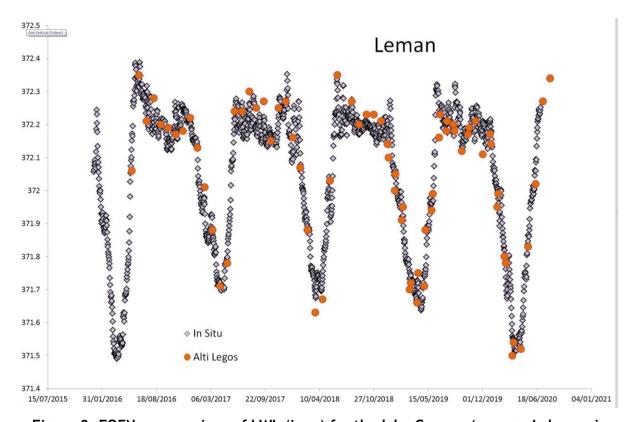


Figure 8. FOEN, comparison of LWL (in m) for the lake Geneva (orange: Lakes_cci, grey: FOEN)

2.4.6. Field work experiments

Figure 9 shows an example of LWL altimetry measurements with Sentinel 3 along two tracks (666 and 707) against in-situ measurements over Issykkul lake. There is an excellent correlation between both series (99%) and a low value of RMS (Figure 10)

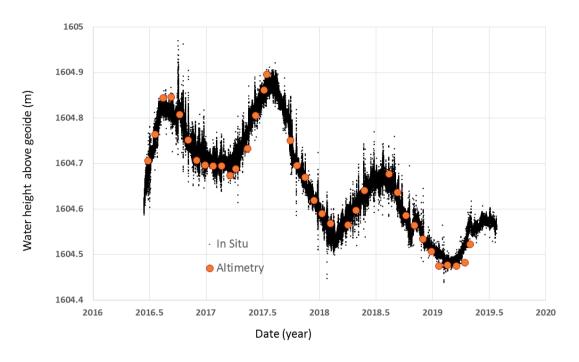


Figure 9. Sentinel 3 vs In-situ measurements for Issykkul Lake

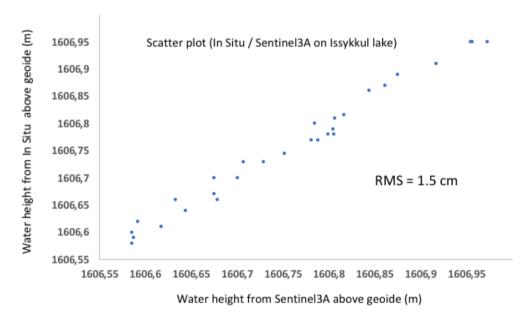


Figure 10. Correlation between Sentinel 3 vs In-situ measurements for Issykkul Lake

2.5. Conclusions and recommendations

We have validated the Lake Water Level, one of the lakes_cci products by comparing the time series to multiple external datasets of in situ measurements. Globally, there is a

good correlation with external datasets from different organisations with data in different regions of the world.

Validation of the Lake water level highlights that, given that LWL being a multi-mission product, the quality of the results may differ over time. In this context a reprocessing of past missions can be very beneficial and is recommended in future generation of datasets

The different comparisons will help us to determine where improvements in the data processing are still needed: better identification of outliers, re-analysis of past missions like Topex / Poseidon, ERS2 or Envisat, and better calculation of some inter-satellite biases.

We would like to thank Prof. Valery Vuglinskiy (State Hydrological Institute, 23, 2nd Line, Vassilievsky Island, 199053, St Petersburg, Russia) for having provided us with in-situ data from Hydrolare lakes.

3. Lake Water Extent (LWE)

3.1. Data description

3.1.1. Optical Data

Landsat 5 and 8 images and Sentinel2 time series have been exploited to derived LWE over a set of test sites.

3.1.1.1. Landsat 5 TM TOA Reflectance:

Landsat 5 TM calibrated top-of-atmosphere (TOA) reflectance were exploited. Calibration coefficients are extracted from the image metadata. See Chander et al. (2009) for details on the TOA computation. (Earth Engine Data Catalog).

Table 7: Landsat 5 wavelengths and resolution

В	а	I	lC	S

Name	Resolution	Wavelength	Description
B1	30 meters	0.45 - 0.52 μm	Blue
B2	30 meters	0.52 - 0.60 μm	Green
B3	30 meters	0.63 - 0.69 μm	Red
B4	30 meters	0.76 - 0.90 μm	Near infrared
B5	30 meters	1.55 - 1.75 µm	Shortwave infrared 1
B6	30 meters	10.40 - 12.50 μm	Thermal Infrared 1. Resampled from 60m to 30m.
B7	30 meters	2.08 - 2.35 μm	Shortwave infrared 2

3.1.1.2. Landsat 8 TOA Reflectance

Landsat 8 calibrated top-of-atmosphere (TOA) reflectance. Calibration coefficients are extracted from the image metadata. See <u>Chander et al. (2009) for details on the TOA computation. (Earth Engine Data Catalog).</u>

Table 8: Landsat 8 wavelength and resolution

Bands Description Resolution Wavelength B1 30 meters 0.43 - 0.45 μm Coastal aerosol 0.45 - 0.51 μm **B2** 30 meters Blue В3 0.53 - 0.59 μm 30 meters Green R4 30 meters 0.64 - 0.67 um Red B5 0.85 - 0.88 μm 30 meters Near infrared R6 30 meters 1.57 - 1.65 µm Shortwave infrared 1 В7 30 meters 2.11 - 2.29 µm Shortwave infrared 2 R8 15 meters 0.52 - 0.90 μm Band 8 Panchromatic В9 1.36 - 1.38 µm 15 meters R10 30 meters 10.60 - 11.19 μm Thermal infrared 1, resampled from 100m to 30m B11 30 meters 11.50 - 12.51 um Thermal infrared 2, resampled from 100m to 30m

3.1.1.3. Sentinel-2 L1C/L2A

3.1.1.3.1. Sentinel 2 Radiometric Resolutions

Table 9: Landsat 8 wavelength and resolution

Bands							
Name	Units	Min	Max	Scale	Resolution	Wavelength	Description
B1				0.0001	60 meters	443.9nm (S2A) / 442.3nm (S2B)	Aerosols
B2				0.0001	10 meters	496.6nm (S2A) / 492.1nm (S2B)	Blue
B3				0.0001	10 meters	560nm (S2A) / 559nm (S2B)	Green
B4				0.0001	10 meters	664.5nm (S2A) / 665nm (S2B)	Red
B5				0.0001	20 meters	703.9nm (S2A) / 703.8nm (S2B)	Red Edge 1
B6				0.0001	20 meters	740.2nm (S2A) / 739.1nm (S2B)	Red Edge 2
B7				0.0001	20 meters	782.5nm (S2A) / 779.7nm (S2B)	Red Edge 3
B8				0.0001	10 meters	835.1nm (S2A) / 833nm (S2B)	NIR
B8A				0.0001	20 meters	864.8nm (S2A) / 864nm (S2B)	Red Edge 4
B9				0.0001	60 meters	945nm (S2A) / 943.2nm (S2B)	Water vapor
B11				0.0001	20 meters	1613.7nm (S2A) / 1610.4nm (S2B)	SWIR 1
B12				0.0001	20 meters	2202.4nm (S2A) / 2185.7nm	SWIR 2

More details can be obtained at **ESA**: Radiometric.

3.1.1.3.2. Level-1C

The Level 2 Sentinel 2 images is not systematically produced all over the world. By the way it is Level 1C data that have been proceed. Level-1C product provides orthorectified Top-Of-Atmosphere (TOA) reflectance with sub-pixel multispectral registration. Cloud and land/water masks are included in the product.

More details can be obtained at ESA: Level-1C Processing for details.

3.1.1.3.3. Level-2A

Level-2A product provides orthorectified Bottom-Of-Atmosphere (BOA) reflectance with sub-pixel multispectral registration. A Scene Classification map (cloud. cloud shadows. vegetation. soils/deserts. water. snow. etc.) is included in the product.

More details can be obtained at ESA: Level-2A product for details

3.1.2. SAR data

The SAR data used to calculate the LWE are Sentinel-1 images acquired in the Level-1 Interferometric Wide Swath (IWS) mode. NORCE has employed these data in Ground Range Detected (GRD) format while TRE-Altamira has employed Single Look Complex (SLC). GRD images contain the detected amplitude and multi-looked to reduce the impact of speckle. SLC images preserve phase information and are processed at the natural pixel spacing. IW mode is a dual-pol acquisition mode. In this case, images are acquired in both VH and VV polarization. Except for some particular cases, the SRTM DEM has been used for geocoding purposes.

Table 10: Sentinel-1 employed data spatial resolution

Mode	Resolution rg x az	Pixel spacing rg x az	Number of looks	ENL
GRD IW	20x22 m	10x10 m	5x1	4.4
SLC IW	2.7x22 m to 3.5x22 m	2.3x14.1 m	1x1	1

A database of Envisat ASAR WSM data is available which has been also used for some lakes. ASAR WSM data has in general coarser spatial resolution, and only one polarization so the quality is in general poorer than for S1 data.

3.1.3. Exogeneous database exploited as inputs

3.1.3.1. Global Surface Water database

The European Commission's Joint Research Centre developed this new water dataset in the framework of the Copernicus Programme. This maps the location and temporal distribution of water surfaces at the global scale over the past 32 years and provides statistics on the extent and change of those water surfaces. The dataset produced from Landsat imagery (courtesy USGS and NASA) will support applications including water resource management, climate modelling, biodiversity conservation and food security. (EU Open Data Portal).

3.1.3.2. Lakes contours database

The analysis is done at a given lake scale. To that matter a precise contour of the lake shore is requested, allowing to decrease processing time but more important to limit artefacts related to the lake's environment.

Whereas exploitation of Medium or low resolution satellite imagery for LWST or LWSR, is based on the analysis of "pure" water bodies, considering that parameters retrieval is done based on the lake AOIs plus a kilometric buffer, for the lake water extent, the investigation is focused on the much precise as possible shore line. This is not a simple limit; for lots of areas, related to water level increase, the water surface's expansion is observed on shoreline, but also on bordering wetlands.

So, for MR and LR satellite imagery exploitation, a relative rough definition of the Area of interest is sufficient, and the buffer application would correct some potential mistake. Then for exploitation of HR satellite imagery, such as Sentinel2, is requiring a precise definition of the AOIs, this can be done exploiting the CCI lakes AOIs database, or the Hydrolakes database derived from the SRTM mission and containing 1,4 millions of lakes larger than 10 ha, (https://www.hydrosheds.org/pages/hydrolakes) but more often in the WP6, the AOIs were at least validated based on Sentinel 2 imagery acquired at different hydrological period, and when requested modified.



Orange: CCI lake AOI



Blue: Hydrolake AOI



Sentinel 2 image acquired on the 2019-05-14

Presenting water inundating the North Western and South Western branch



Red: AOIs defined and exploited in order to include the two "'arms" that correspond to wetlands more of less inundated due to water level increase.

Figure 11. Comparison of AOIs contours: case of Bosten lake (PR China).

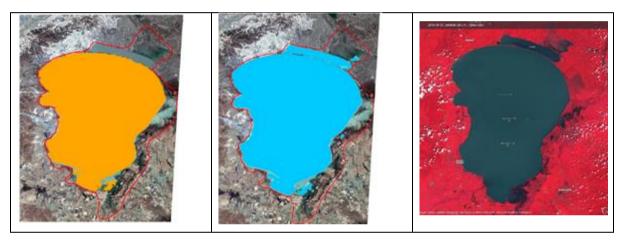


Figure 12. Comparison of AOIs contours: case of Khanka Lake (PR China-Russia) with in orange CCI Lakes AOIs, in blue, Hydrolake and in red exploited AOI for Sentinel2 exploitation.

When comparing, with a Sentinel2, 2019-08-03, it is well noticeable that the AOIs proposed by both Hydrolakes and CCI lakes database are too restrictive, as excluding the wetlands in the North and South East parts of the region.

3.2. Comparison of methods

3.2.1. Comparison of methods for optical sensors (VIS NIR SWIR)

The aim of this part is to compare lakes area and lakes vector extracted from sentinel-2 and Landsat images with different classifiers. Initially two none supervised approach: OTSU and K-Means and two supervised ones. SVM and Random forest have been applied on the data set over the lakes' test. Finally, for this comparison of the results from the optical processing approaches only OTSU SVM and Random forest are discussed.

3.2.1.1. Description of work

A selection of lakes with various characteristic, i.e. spectral behaviours of shallow waters, environment more or less arid, relief, presence of ice and/or snow etc.. (Table 11 and Table 12) have been done. The results obtained over these lakes based on different processing approaches have been compared.

More precisely, this analysis was done in three steps. In a first one, a relative long time series of images was selected for a first set of lakes. In a second time, a specific inter comparison was done on a short dense time series over Chad lake covering a field period survey. At least a third time, for consolidation of the approach, as well as investigated a better parametrization of the process, such as the Random Forest (RF) approaches, an additional set of lakes, with a relative low amount of images, 6 to 8, was selected in order to present different levels' state of theses lakes.

Table 11: Major characteristics of the Chad lake case study and the second set of the analysed lakes

	Altevatnet Norway	Colhue Argentin a	Namts o China	Alakol Kazakhst an	Sassykkol Kazakhst an	Chilwa Malawi	Al Hamar Irak
Shadow water		XXX				Х	XXX
ice	XXX		XXX	XXX	XXX		
snow	XXX		XXX	XXX	XXX		
Topograp hic position	xxx						
Local environm ent	XXX	XXX	XXX	XX	XX	XXX	XX
sunglint			XXX				
Floating vegetatio n						XXX	XXX
Lake dynamic	Х	XXX	Х	Х	Х	XXX	xxx

Table 12: Major characteristics of the Chad lake case study and the second set of the analysed lakes

	Chad	Argentino Argentina	Bosten China	Khanka China Russia	Illmen Russia	Sary kamysh Turkménistan - Ouzbékistan
Shadow water	Х		Х	Х	Х	
ice		XXX	XX	XXX	XX	
snow		XXX		XXX		
Topographi c position		XXX				
Local environme nt	XX			Х		Х
sunglint						
Floating vegetation	XX		Х	XX	х	
Lake dynamic	Х		XX		XX	Х

3.2.2. Validation of LWE derived from HR optical sensors based on VHR sensors

Validation of water extent is pure and great challenge by itself. Few methods can be investigated:

- Comparison of LWE with databases. There are lot of limitations, genes of the database, such as the resolution of the input data, the date/period, a LWE can change greatly from date to date, a year to another.
- Field comparison by surveying the water bodies limits walking along the shore with a GPS tracking, or using a boat or kayak to follow the shore. It is not always possible to walk around lake, all there are the question of the accessibility of the lake. In case of shallow water, what is the exact distance between the boat and the shore, a tens of meters (or more) that represents two to five/ten Sentinel pixel.
- Extract LWE from satellite of high and very high resolution.

The last approach is according to us the most promising. Therefore, it is not so convenient to be implemented, it requested pair of HR and VHR images acquired within a very short time. The data have to cover if possible, the targeted lake as a whole, englobing the surrounding areas. Most of the time, and it is particularly the case with CCI lakes that are large lakes. So on in most of case only a part of the lake is covered by the two sensors and so what is the representativity of the covered area. And, of course, when VHR data have to be ordered it can be a costly approach. For this reason, an agreement with the CNES, French Space Agency was initiated in order to order a low coast VHR SPOT 6-7 or Pleaides imagery and share these data with the WP6 team. Therefore, the analysis of the

catalogues was not so successful and this approach was abandoned. Hopefully in parallel, CNES was able to order VHR Pleiades images over large reservoirs in France, the Der and Orient lakes, and this during a dynamic period of infilling.

3.2.2.1. Location of the test areas

The Der and Orient lakes, located in the East of Paris, within the Champagne area region, are part of the Seine River flood management systems.

The Der lake is the largest artificial reservoir in France, with a surface around 48km2, for a maximal depth of 18m, the Orient water surface is 22 km2 for a maximal depth of 22 m. The functioning of the reservoir is the following:

Water is taken from the rivers, i.e. Marne River for Der lake, la Seine River for Orient lake, from November / December to June, thus filling the reservoir. From July to October, water is released to support the flow of rivers. As a result, water surfaces change considerably during the year, for the Der lake going from around 40 square kilometers during the high season, to less than ten square kilometers during the very low water period.



Figure 13. Location of the Der lake and Orient Lake.

3.2.2.2. Exploited data

The lakes are located within an Overlapping part of Sentinel 2 tracks, allowing up to 14 acquisitions by month. So, it was an ideal case to order VHR Pleaides imagery, knowing that the acquisition will by the way have at maximal one day of delay between the VHR and HR data.

Finally, two pairs of Pleiades HR data, 70 cm of spatial resolution, a panchromatic channel and 4 visible ones from blue to near Infrared channels, were acquired on the 30 of

December, with a delay of one day with Sentinel2 and 6 of January 2020, same date as a Sentinel2 acquisition.



Figure 14. The Der Lake; as viewed by Pleiades on the 30 of December 2019, and on the 1 of January 2020

From the Pleiades and Sentinel2 data, LWE were extracted for each date and each reservoir based on a SVM approach.

3.2.2.3. Results

When comparing the LWE derived with Pleaides and Sentinel 2 data acquired within 24 hours, the difference in term of surface are very low, i.e. one 29.07 km2 for Sentinel 2, and 30.58 km2 for Pleiades. 95% of Pleiades Water is recognized by Sentinel 2. There is a very low level of commission, $0.05 \, \mathrm{km}^2$.

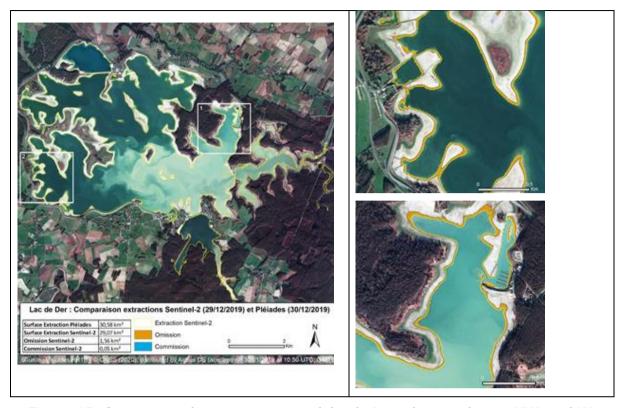


Figure 15. Omission and commission over lake de Der when exploiting VHR and HR images acquired with a 24h of delay

When the acquisition of the VHR and HR images was the same day, 95,5% of Pleiades Water is recognized by Sentinel 2. Of course, there is an effect of resolution i.e. 10 m versus 0.70m. The shoreline is off course much finer on the VHR derived LWE. Therefore, the omission is relatively low, and the commission very low.

This case of study allows also to evaluate, in the context of infilling reservoir, the part of the 24h of delay between the two acquisitions.

When data are acquired the same day, the space occupied along the shore of the omission is very narrow, in fact the shoreline corresponds to a staircase, of swatooth's effects, alternating omission and commission pixel. An effect that is related in fact to the difference of spatial resolution.

Where, the LWE represent two stages of infilling, we observe a large omission belt around the lake shore. This belt in fact corresponds to the increase of the surface of water within one day. So, of course, what is seen as water on the Pleiades image, cannot be described as water on the Sentinel2 image acquired a day before.



Figure 16. Comparison of LWE accuracy derived with one day of delay between the Sentinel2 and Pleaides acquisitions and the same day (right).



Figure 17. Omission and commission over lake de Der when exploiting VHR and HR images acquired quasi simultaneously.

3.3. CCI test sites result analysis

3.3.1. Results analysis for optical sensors-based approaches

As indicated, a first analysis was conducted lake by lake. An analysis of each important difference/gap from an approach to another one was done and commented.

3.3.1.1. Altevatnet Lake

Altevanet Lake is a narrow long lake, i.e. 2*38 km2, within an incisive valley is located rather north of Norway: 68°N. This Northern location induces the presence of snow/ice on the lake shores as well as relatively low solar position. The analysis of the observed gaps can be related to these characteristics.

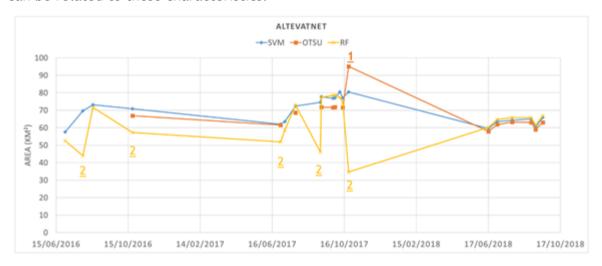


Figure 18. LWE derived from optical imagery based on different approaches over Altevatnet lake.

Observations based on the graphic and vector extractions analyse are the following:

- On the S2 image of the 23/10/2017. OTSU's overestimation caused by snow.
- On the images of the 30/07/2016. 22/10/2016. 30/06/2017. 05/09/2017 & 23/10/2017 RF's generates an underestimation caused by low reflectance (lower than training image) and wisp of cloud.
- In some case, SVM classifies shadows into water's class. Same problem can also be observed when applying RF.

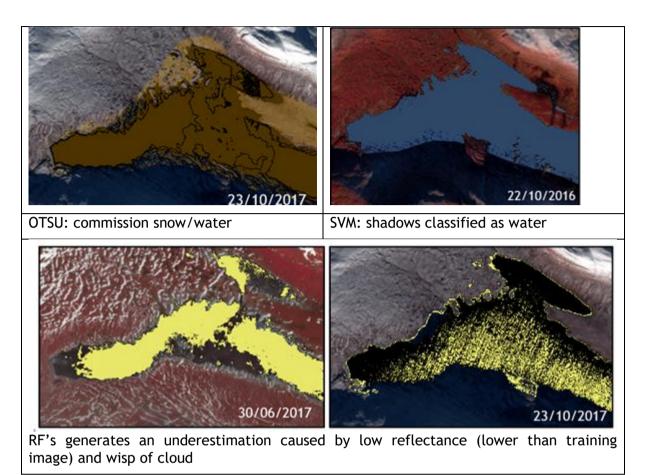
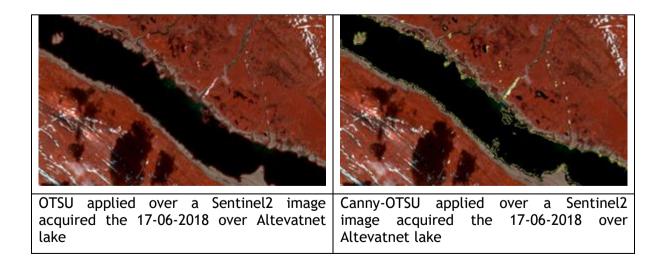


Figure 19. Omission and commission observed over Altevatnet lake

In addition, over the Altevatnet site, Otsu and Otsu Canny approaches were tested. It is appearing that the OTSU Canny is more restrictive on water, therefore it is also OTSU Canny that is more strongly influenced by shadows, snow, clouds.



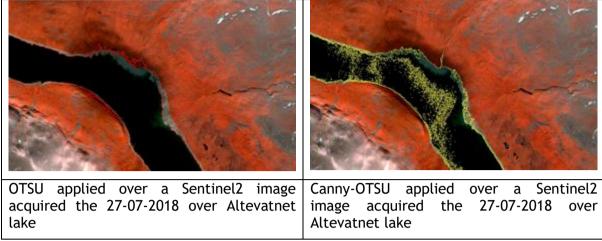


Figure 20. Comparison of Otsu and Canny-Otsu results.

3.3.1.2. Colhue Lake

Colhue Lake. It is located in a tectonic depression reshaped depression by wind and fluvial activities on the Western margin of the Patagonian plains and it is protected from Western rains by a North South a ridge of 600m. It is a shallow lake that knows fluctuations in terms of precipitations and rivers discharges resulting on LWE variations at least from 1998 to 2015. The analysis of the Sentinel 2 time series indicates that for the bear period, i.e. 2015-2017, the lake known a period of total dryness and of rapid infilling.

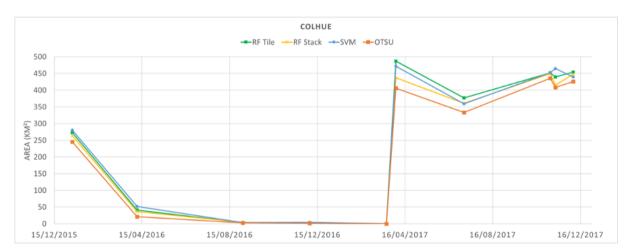
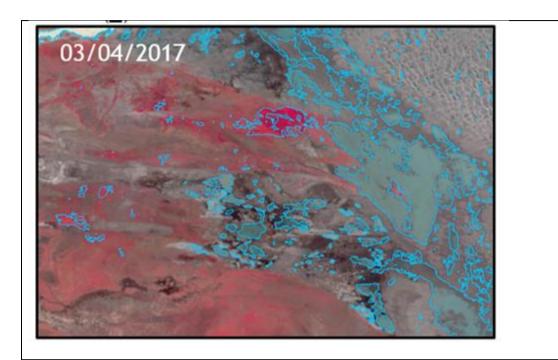


Figure 21. LWE derived from optical imagery based on different approaches over Colhue Lake.

Over Colhue Lake, the difference processing's approaches provided very similar results Observations based on the graphic and vector extractions are the following:

 Main differences between methods are caused by the misidentification of the water/non-water limits. SVM and RF classify muddy part of the lake/trickles of water. SVM classify also some vegetation as water surface. Not detected by RF.



Commission between vegetation and water, when applying SVM on Sentinel2 image

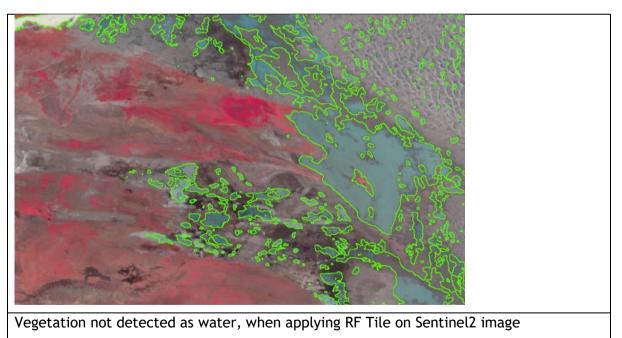


Figure 22. Analysis and comparison between SVM and RF Tile approaches over vegetated areas bordering the Colhue Lake.

In addition, over this test site, Otsu and Otsu Canny approaches were tested. The obtained results are for this case very similar.



Figure 23. OTSU (left) and Canny-OTSU (rigth) applied on a Sentinel2 image acquired the 27/02/2019

3.3.1.3. Namtso Lake

Namtso Lake is a large water bodies of 1900 km2 located on the Tibetan Plateau. This lake knows a relative increase of water height and water extent (+600km2) from 1994 to 2000 and would have been relatively stable since then. It is located at a very high elevation, ie 4720 m, by the way the lake is covered by ice a long part of the year, from November to April, and the atmospheric effect are important.



Figure 24. LWE derived from optical imagery based on different approaches over Namsto Lake.

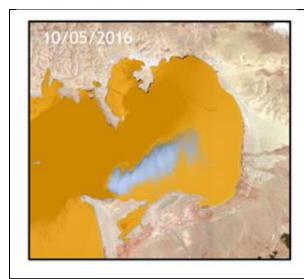




Figure 25. OTSU processing: under estimation du to sunglint and over estimation du to snow.

Observations based on the graphic and vector extractions analyse:

- On the images of the 10/05/2016 and of the 24/07/2018 OTSU's underestimation caused by sunlight.
- On the S2 image of the 21/11/2018, there are OTSU's overestimation caused by snow.
- SVM and RF have pretty much the same good results. Differences between methods are lower than 5%.

3.3.1.4. Sassykol Lake

Sassykol Lake is located at an altitude of 350 m in the southeast of the eastern province of Kazakhstan in the Balkhash-Alakol lowland between mountain systems of Zhetysu Alatau in the south. Tarbagatai in the north, and Barlyk in the east. It is the part of a complex hydro system formed by three successive lakes: Sassykol on the upstream part, the Kosharkol and in downstream the Alakol Lake. Sassykol lake communicates with the Alakol Lake through an extensive wetland surrounding the intermediate lake named the Kosharko Lake Sassykol is a shallow fresh-water lake which shores are gently sloping and densely covered with reeds. From November to the end of March-April, the lakes are covered by ice.

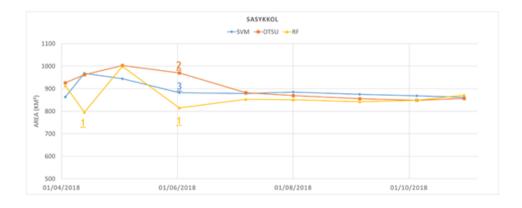


Figure 26. LWE derived from optical imagery based on different approaches over Sassykol lake.

- 1. Random Forest approaches (RF) involve an underestimation that could be caused by low reflectance (lower than training image) and wisp of clouds.
- 2. OTSU's approach generates an overestimation: it classifies vegetation and wisp of the cloud.
- 3. SVM: it shows a good classification. The boundaries of the water bodies are well marked, the vegetation is well distinguished.

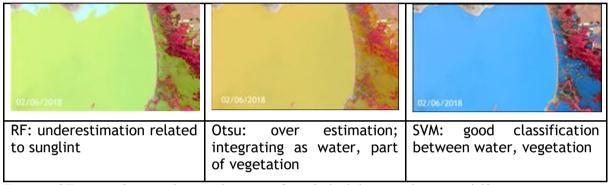


Figure 27. Artefacts observed over Sassykol lake exploiting different processing approaches.

3.3.1.5. Alakol Lake

Alakol Lake is an endorheic salted lake relatively deep, 45m, and large, more than near 3000 km2. The shores of Alakol Lake are rugged and an unstable coastal zone, with also large islands. As already indicated, from November to end of March-April, the lake can be covered by ice.



Figure 28. LWE derived from optical imagery based on different approaches over Alakol lake.

Observations based on the graphic and vector extractions analysis are the following:

- RF-underestimation caused by a wisp of cloud for the images from 28/08/2018, 22/10/2018, 06/11/2018 & 24/07/2018, indicated with (1) on the Figure 29.
- OTSU's approach presents an overestimation caused by snow, on the Sentinel2 image acquired on 26/11/2018, indicated with (2) on the Figure 29.

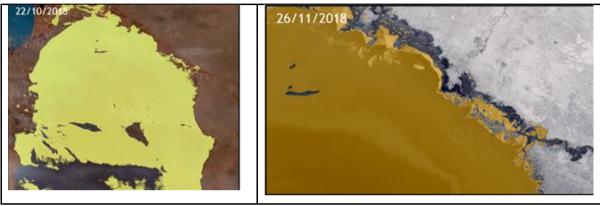


Figure 29. Artefacts observed over Alakol Lake when applying, left; RF: underestimation related to clouds right, OTSU; commission with snow.

3.3.1.6. Chilwa Lake

Lake Chilwa is the second-largest lake in Malawi after Lake Malawi. It is in eastern Zomba District, near the border with Mozambique. Lake Chilwa is a shallow, i.e. 3m, enclosed saline lake located along the East African Rift Valley in southern Malawi near its border with Mozambique.

Approximately 60 km long and 40 km wide. The lake is surrounded by extensive wetlands. There is a large island in the middle of the lake called Chisi Island. The lake has no outlet and the level of water is greatly affected by seasonal rains and summer evaporation. In recent years, Lake Chilwa has been shrinking.



Figure 30. LWE derived from optical imagery based on different approaches over Chilwa Lake

Observations based on the graphic and vector extractions analysis are the following:

- OTSU: underestimates the water surface indicate by (1) on Figure 31.
- RF: underestimates. don't classify water with vegetation
- SVM: overestimates LWE as classifying burned areas as water surfaces

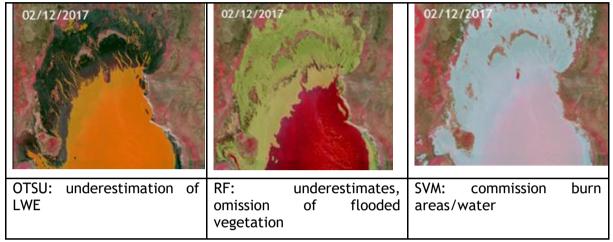


Figure 31. Artefacts observed over Chilwa Lake when applying OTSU, RF Tile and SWM on Sentinel2 imagery.

3.3.1.7. Al Hamar Wetland

Al Hammar wetlands are a large <u>wetland</u> complex in South Eastern <u>Iraq</u> that are part of the <u>Mesopotamian Marshes</u> in the <u>Tigris–Euphrates river system</u>. Historically, the Hammar Marshes extended up to 4,500 km2 during seasonal floods. They were destroyed during the 1990s by large-scale drainage, dam and dike construction projects. Since 2003, they are recovering following reflooding and destruction of dams. The water spatial distribution is still very controlled by inherited structures.



Figure 32. LWE derived from optical imagery based on different approaches over Al Hamar wetlands.

The LWE dynamic derived from satellite image, thanks to OTSU or SVM approaches are similar. Therefore, the values derived from SVM algorithm are higher than the OTSU ones. This is related to the fact that OTSU is more restrictive, taking into account free open water surface, rather than SVM which includes also a part of flooded vegetation.

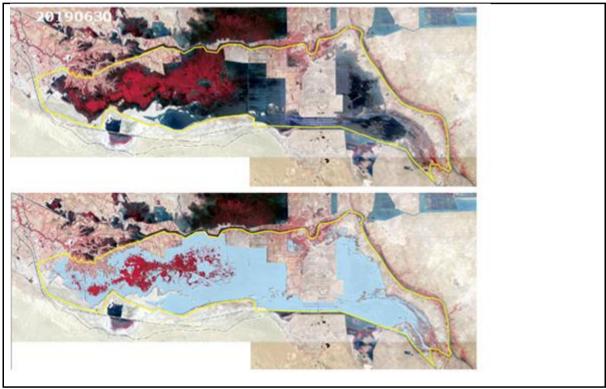


Figure 33. LWE extraction based on a SVM approach, initial image and resulting LWE: part of the flooded vegetation is integrated within the LWE.

3.3.1.8. Chad case study

The Chad case of study is particular in regard to the others tested lakes cases. It was an opportunity to:

- Test 5 different algorithms for LWE extraction based on optical imagery
- Compare results acquired from Sentinel2 and landsat8 data acquired the same day, but of course having different spatial resolution, ie 10 and 30m respectively.
- Compare LWE derived both from Optical and SAR imagery
- Compare the EO derived map with the with field tracks recorded from a boat.

Chad lake is a historically large, shallow, endorheic lake in Africa, which has varied in size over the centuries. According to the Global Resource Information Database of the United Nations Environment Programme, it shrank by as much as 95% from about 1963 to 1998.

This field survey was carried out in the Archipelago of the Chad lake, the Bol Reria area a disconnected part of interdune water bodies.

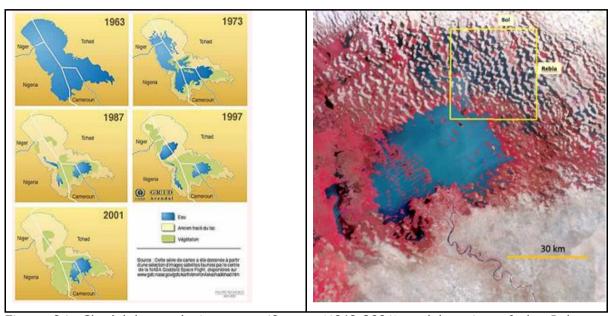


Figure 34. Chad lake evolution over 40years (1963-2001) and location of the Bol area within the Archipelago region of Chad Lake

In the Archipelago area, the islands and peninsulas are summits of remaining sand dunes' network flooded by lake Chad water. Dune sonnet area: this is the part of the island where there is a small group of huts and a few millet fields when the island is inhabited. The vegetation corresponds to trees and shrubs. On the shores of the islands, fringe of macrophytes can be observed. One characteristic of this part of Chad lake is the presence of Reeds islands that are anchored in shallow waters. Part of them separate from the main islands and form circular floating islands of vegetation named, kirtas. These are formed by Papyrus and Phragmites. Their sizes vary from few meters to several hundred meters. At the time of reversal of dominants winds, during June and October generally in the rainy season when the winds shifts are frequent, these islands move back and forth, modifying the aspect of the reeds islands and closing the channels of the Archipelago (Leonard, 1974; Ittis and Lemoalle, 1983).

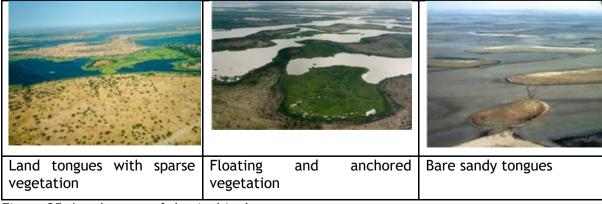


Figure 35. Landscapes of the Archipelago area.

Five different methods were tested, OTSU, Cany-OTSU, Kmean, SVM and RF. Obtained LWE are presented in the Figure 36.

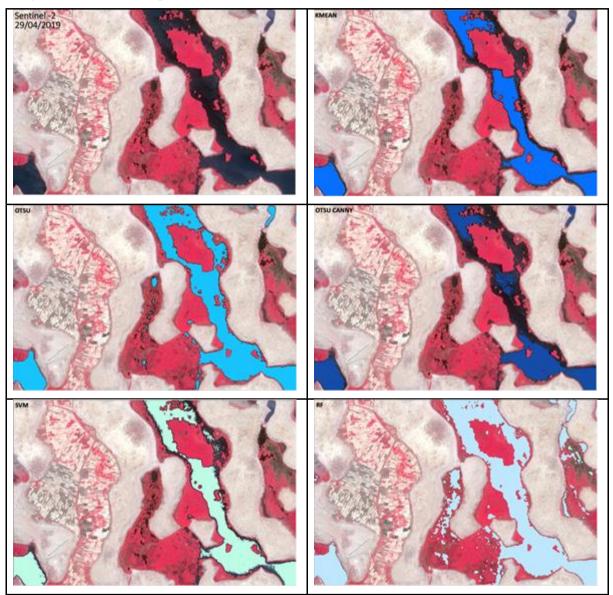


Figure 36. LWLs extracted exploiting different approaches from the Sentinel2 image acquired on the 29 of April 2019.

In this complex environment, the analysis and comparison of the derived LWE from Sentinel2 imagery are the following:

- Canny OTSU and Kmeans, a none negligible underestimate the LWE
- SWM a small under estimation
- OTSU, the most realistic over this landscape
- RF a small overestimation as integrating part of the wet/floating vegetation

All methods retrieve about 70 % of the potential observed water, all real open water surface. The differences are noticeable on very shallow water and on the immediate environment or inside floating islands.

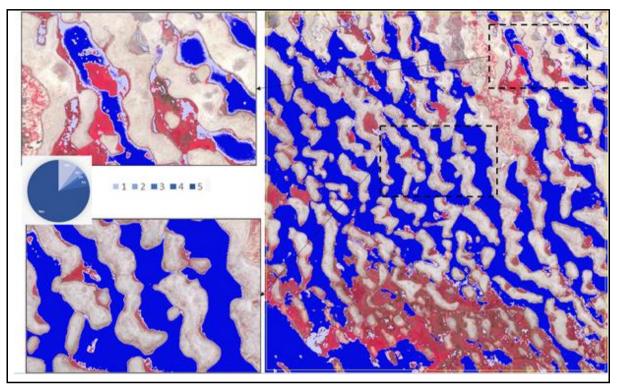


Figure 37. Occurrence map of observed water based on Sentinel2 imagery.

The same five approaches were also followed to retrieve LWE from the Landsat imagery. In this complex environment, the analysis and comparison of the derived LWE from Landsat imagery are the following:

- Kmeans, a none negligible underestimate the LWE
- Canny-OTSU presents a small under estimation
- SVM and OTSU, are very similar in terms of obtained results and present the most realistic over this landscape
- RF a small overestimation as integrating part of the wet/floating vegetation

Based on Landsat 8 imagery, i.e. with a pixel integrating more surface, all methods present in common more than $80\,\%$ of the potential observed water, all real open water surface. The slight differences are noticeable on very shallow water and on the immediate environment or inside floating islands.

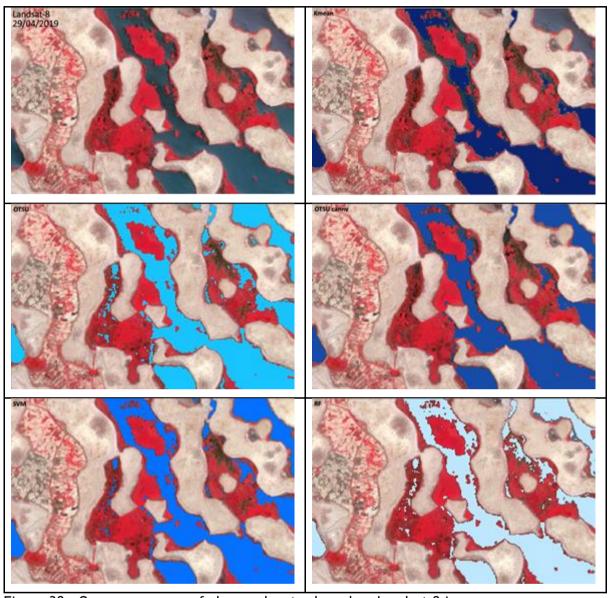


Figure 38. Occurrence map of observed water based on Landsat 8 imagery.

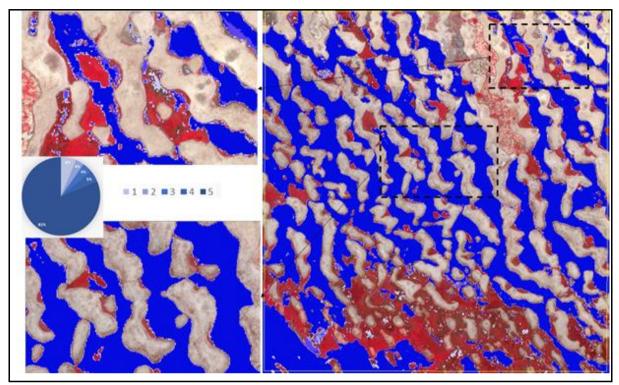


Figure 39. Map of occurrence of observed water based on Landsat8 imagery.

The comparison of LWE extraction from Sentinel2 and Landsat 8 is interesting, the extraction based on Landsat 8 is more coherent/continuous, with less difference between the results derived from the various methods.

Table 13. Comparison of the Sentinel 2 and Landsat 8 bands.	Table 13.	Comparison of	the Se	ntinel2 and	d Landsat	8 bands.
---	-----------	---------------	--------	-------------	-----------	----------

	Sentinel 2		Landsat 8
NIR	0,779-0,885 10m	В8	0.85-0.88 um
NIR	0,853-0,875 20m	B82	0.85-0.88
SWIR1	1,568 - 1,659		1.57-1.65
SWIR 2	2,114- 2,289		2.11-2.29

If in the SWIR1 and SWIR, the bands are located similarly and have same width, on the NIR there are noticeable difference between Sentinel2, band B8 10m, and landsat8, band 5 whereas the narrow band B8A of Sentinel2 is similar to the NIR band, band 4, of Landsat 8, with respectively a 20 and 30 meters of spatial resolution. Whereas, on the LWE derived from Sentinel2, the limits water/non water are more accurate, but some overestimation is observed with vegetation areas classified. On the LWE obtained from a RF approach over a Landsat 8 image, it can be observed an underestimation of water on the shore of the lake due to the lower resolution.

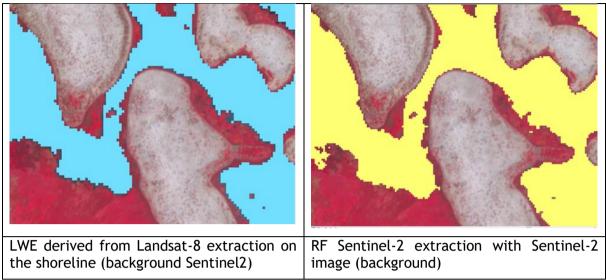


Figure 40. Comparison of Sentinel 2 and Landsat 8 LWE: influence of the spatial resolution.

In addition, a multitemporal analysis was done over a period of one month, exploiting five Sentinel 2 images. These data were acquired at the following dates.

- 09/04/2019
- 14/04/2019
- 24/04/2019
- 29/04/2019
- 04/05/2019

For each image, a water surface based on an SVM approach was derived. From these a mean surface was calculated and then each surface compared to this mean surface.

Table 14.

Date	LWE	Difference / Mean
09/04/2019	378.7875 km²	0,7%
14/04/2019	371.5667 km ²	1,4%
24/04/2019	372.1424 km²	-1,03%
29/04/2019	379.2515 km ²	0,85%
04/05/2019	381.6009 km²	1,48%

Very stable water bodies, with less than 2% of difference between the observations within a month. In addition, the LWE apparent dynamic is more related to the movement of floating vegetation (under wind direction/speed dependence) rather than LWL changes. Some also apparent changes are related to some artefact linked with small clouds on one date.

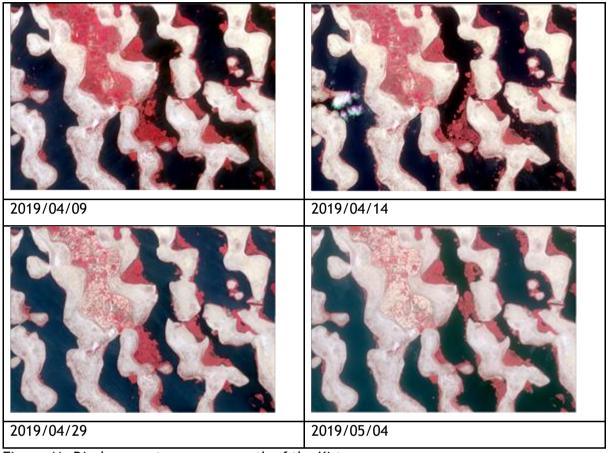
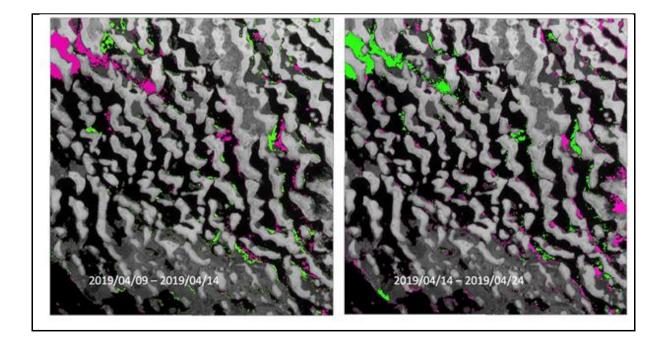


Figure 41. Displacement over one month of the Kirtas.



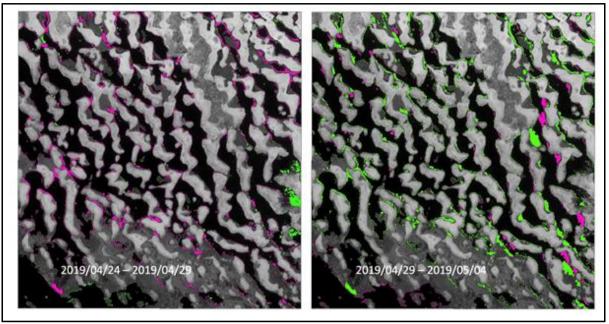


Figure 42. Displacement of vegetated islands from date to date. Green colour: No water in first image, water in last image. Lily colour: Water in first image, no water in last image.

3.3.1.8.1. Validation with boat track.

A field trip has been organized by LEGOS on parallel of the CCI lake work. Boat tracks on Landsat, using a GPS, receiver allowed to carefully map the water and the vegetation. It is allowed locally to validate the water limits, another part the boat is in the middle of bays. An interesting thing is also that the tracks cross within vegetated island, confirming well the "water" aspect of these island. All changes in the nature of the terrain crossed by the boat were noted allowing to precisely map the lake countour along the boat's route.

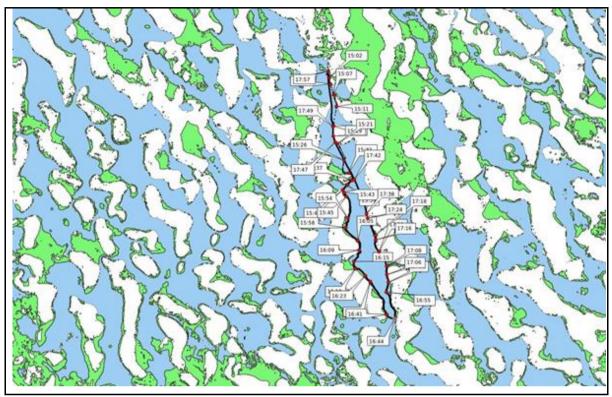


Figure 43. Tracks of boat survey, 10 of April 2019 within the Chad lake Archipelago (Courtesy of Legos).

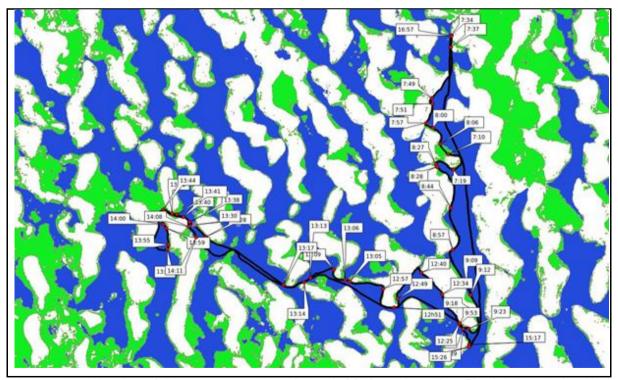


Figure 44. Tracks of boat survey, 11 of April 2019 within the Chad lake Archipelago (Courtesy of Legos).

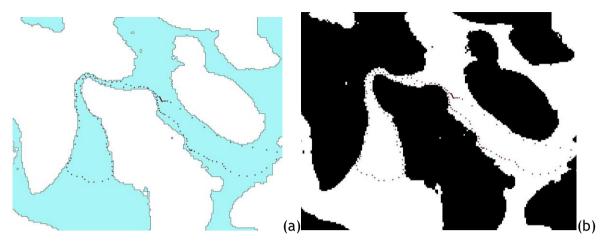


Figure 45 (a) Processing using normalized MNDWI index with threshold of -0.1 (b) using OTSU method. Both processing of the landsat 8 images allow to correctly map the transition between land and water in this complex archipelago system.

3.3.1.9. Argentino lake

Lago Argentino, also name El Calafate, is a lake in the Patagonian province of Santa Cruz, Argentina. It is the biggest freshwater lake in Argentina, with a surface area of 1,415 km2. It has an average depth of 150 m, and a maximum depth of 500 m. The lake lies within the Los Glaciares National Park in a landscape with numerous glaciers and is fed by the glacial meltwater of several rivers, the water from Lake Viedma brought by the La Leona River, and many mountain streams

The argentine lake case is interesting, due to the potential impact of environment on the retrieval of LWE. Two major elements have to be taken into account, the topographic position of the lake surrounding by mountains, and the presence of glaciers feeding the lakes and of snow covering the borders. By the way it is challenging case for testing the different methods of LWE based on optical imagery.

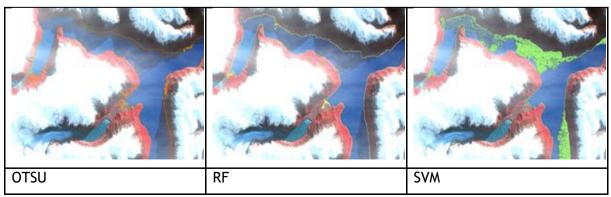


Figure 46. Artefacts observed over Argentino Lake when applying OTSU, RF Tile and SWM on Sentinel2 imagery acquired the 02-06-2019.

In this second set of tests for each processed image, SVM and RF was trained based on Pekel water mask occurrence.

It is appearing that:

- OTSU and SVM are sensitive to shadows on water surfaces.
- RF is less sensible to shadows effect and by the way presents a more realistic shoreline.

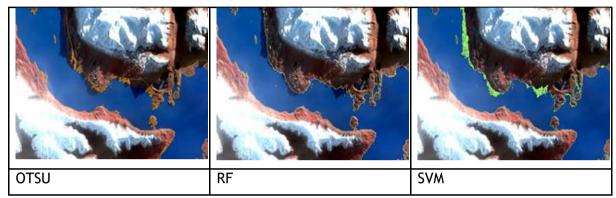


Figure 47. Artefacts observed over Argentino Lake when applying OTSU, RF Tile and SWM on Sentinel2 imagery acquired the 02-06-2019.

When analyzing the temporal evolution of the LWE, a relative coherence of the LWE evolution over the time obtained by the different process. Therefore, the OTSU tends to overestimate the water extent, whereas the SVM, underestimates it. This explains the differences between the methods at 02/06/2017. Large shadows are present at the limit between the lake and the land. RF Tile is more consistent than SVM which don't classify shadows on water and OTSU which classify shadows on land.

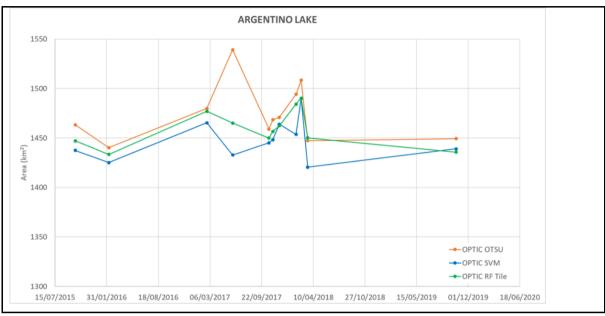


Figure 48. LWE derived from optical imagery based on different approaches over Argentino Lake

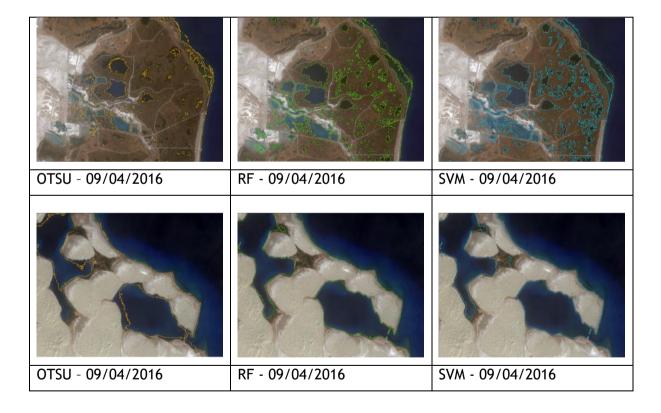
3.3.1.10. Bosten lake

The Bosten lake, also known as Bagrash lake, is a large freshwater lake on the North Eastern rim of the Tarim Basin, about 20 km East of Yanqi, Xinjiang, in the Bayingholin Mongol Autonomous Prefecture, an extremely arid region in the North West of China. Covering an area of about 1000 km2, it is the largest lake in Xinjiang. The mean water depth is 8,2m with a maximum depth of 17m. The lake is frozen during the winter (up to March). In the Western part: there is a huge wetlands complex that is totally dry in beginning spring. In addition, the environment of the lake varies a lot around the year why lot of vegetation in summer (as well on the nearly agricultural fields that could induce some potential confusion with the wetlands vegetation. In fall, the vegetation is relatively dry.

In term of data accessibility and quality, it has to be noticed that a part of the analysis is spoiled by the fact that a relative important amount, i.e. 7 dates, of Landsat 7 images covering an interesting hydrological period where not exploitable due to some instrument artefact.



Figure 49. LWE derived from optical imagery based on different approaches over Bosten lake.



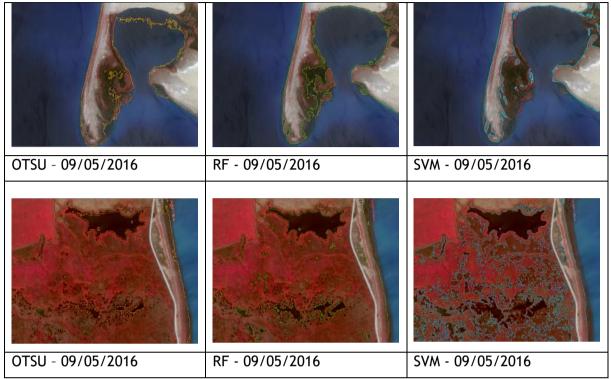


Figure 50. Comparison of LWE limits obtained from OTSU, Random Forest (RF) et SWM over Bosten wetlands.

Observations based on the graphic and vector extractions analysis are the following:

- OTSU: underestimates some shallow water
- RF: The most realistic over this landscape.
- SVM: overestimates water surface, especially in dry wetland
- The RF approach is the best technical solution to derive accurate LWEs over the Bosten Lake.

As illustrated above with water masks derived from the images Sentinel-2 acquired the 09/04/2016 and the 09/05/2016, OTSU is overestimating and SVM is underestimating on this data.

3.3.1.11. Khanka

Lake Khanka is located on the border of the People's Republic of China (China) and the Russian Federation (Russia). It is the largest lake in Northeast Asia, as well as a transboundary waterbody between China and Russia. The lake is called Lake Xingkai in China and Khanka Lake in Russia. The water plane of the lake varies between 4,000-4,400km2. The average lake depth is 4.5m and maximum lake depth is 10.6m.

There are 23 inflowing rivers to the lake, 8 draining from China and 15 draining from Russia. The Song'acha River is the only outflow river from the lake and is subsequently connected with the Wusuli/Ussuri River and the Heilong/Amur River system. The drainage basin of Lake Xingkai/Khanka is a habitat for important animal and plant species of both countries, particularly the wetlands surrounding the lake. The Russian Federation designated the lake as a Ramsar Convention wetland site, on the basis of its importance for migratory bird species.



Figure 51. LWE derived from optical imagery based on different approaches over Khanka Lake.

The comparison of the LWE obtained based on OTSU, SV and RF indicates:

- OTSU, Overestimation observed in dry wetlands and in crops that could extend far beyond the AOI boundaries.
- SVM, Underestimation in small water bodies.
- RF is Closest from the ground truth.

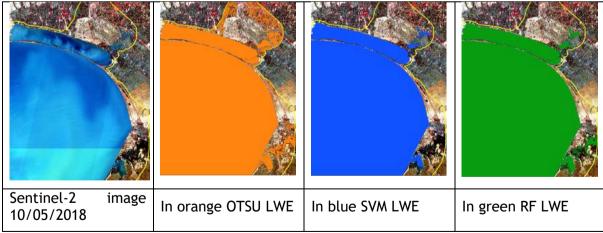


Figure 52. LWE obtained with OTSU, SV and RF.

Presented at local scale in the above figure, these trends are observed on the whole S2 image. Conclusion over Khanka case of study is that RF approach allows to derive accurate LWEs.

3.3.1.12. Ilmen lake

The Ilmen lake is a large lake in the Novgorod Oblast of Russia. The average surface area is 982 Km2, therefore it may vary between 733 km2 and 2090 square km2 depending on water level. The lake is fed by 52 inflowing rivers, the four main ones being the Msta, the Pola, the Lovat, and the Shelon. It is drained through a single outlet, the Volkhov, into Lake Ladoga.

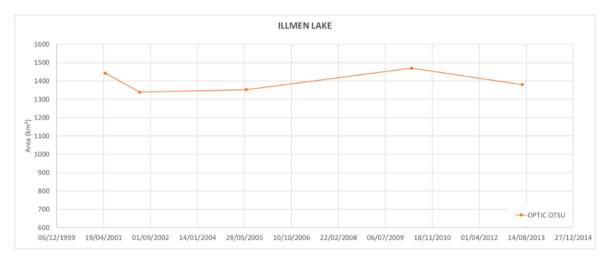


Figure 53. LWE derived from optical imagery based on different approaches over Illmen Lake.

Part of the processing was done exploiting OTSU approach from GEE. Therefore, not all the images selected from USGS sites, are accessible from GEE, by hence, the optical time series is not so large.

3.3.1.13. Sary kamysh Lake

The Sarykamysh or Sary-Kamysh, also known as Sarygamysh is situated in Central Asia approximately midway between the Caspian Sea and the Aral Sea. The lake sits in an oval depression of tectonic origin, which was later affected by aeolian erosion during successive glaciations from 2 million years ago, to 10 000BC. The Northern quarter of the lake belongs to the country of Uzbekistan, while the rest belongs to Turkmenistan. The Sarykamish lake was formed in 1971 as a result of flooding of a set of small lakes located within the depression. Now Sarykamish lake is a large drainage water body which has been used as a discharge collector of salty irrigation water from the fields. Since 1992 the lake has been progressively increased in size a, reaching maximal levels in the beginning of 2000, an increasing still on going with some recession phases. The mean water depth is 8 m with a maximum depth of 4m; its surface is about 4000km2.

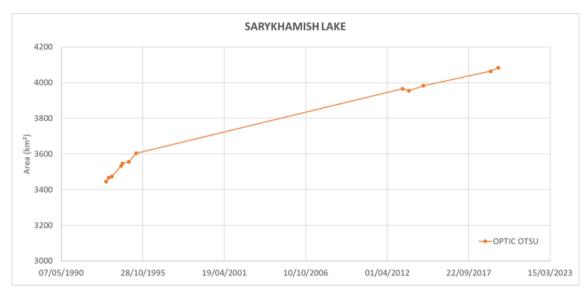


Figure 54.: LWEs derived from optical imagery based on different approaches over Sary Khamish Lake.

The idea was to have a selection of images covering the different phases of lake inflow. The selected data started from the beginning of the 90' up the recent years. Unfortunately, a relative long period, in the beginning of the 2000' was not covered due to the instrumental problem on Landsat 7.

3.3.2. Results analysis for SAR sensors-based approaches

As commented in Section 3.1.2, series of Sentinel-1 images have been exploited to derive their Lake Water Mask and their corresponding Lake Water Extent. The LWE measurements will be used along with the Lake Water Level in order to generate the corresponding hypsometric curves.

As described in the ATBD document, NORCE and TRE-Altamira have employed different approaches (even though sharing some common steps) and have used GRD and SLC images respectively. A main limitation with S1 data is that this constrains the time-series to the period after October 2014. S1 data are geocoded using the SRTM DEM and a most precise one when available (as for Lake Altevatnet).

The Lake Water Masks generated by NORCE and TRE-Altamira over the same images' dates (or presenting a low temporal difference) are used in order to compare both methodologies' relative performances. In order to maximize the contribution of the SAR LWE to the hypsometric curve calculation, NORCE and TRE-Altamira employed different S1 datasets (including different orbits) but also considering some common dates for comparison purposes. In the same way, images presenting a time difference of 12 days have also been considered. This can be questionable sometimes when the LWE changes rapidly, but in order to obtain sufficient amount of comparisons we have used this criterion. This analysis has been carried out over the following lakes: Chad (section 3.3.2.1), Illmen (section 3.3.2.2), Bosten (section 3.3.2.3), Argentino (section 3.3.2.4) and Khanka (section 3.3.2.5).

The comparison statistics are derived from the contingency matrix. For its calculation an analysis polygon covering the lake is selected and one of the two results is taken as the reference. Then the following parameters are calculated in %:

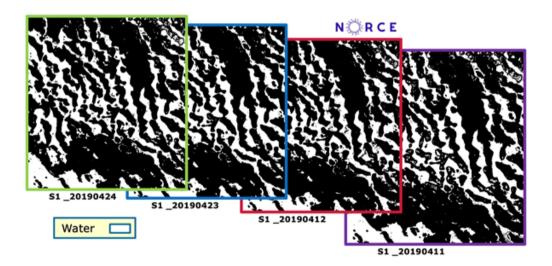
- True positive (TP): both results detecting water.
- False negative (FN): reference result detecting water and the other land.
- False Positive (FP): reference result detecting land and the other water.
- True negative (TN): both results detecting land.
- Accuracy rate (AR): TP + TN.

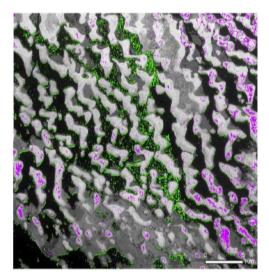
It is worth saying that the naming of the statistical parameters (TP, FN, FP and TN) is also somewhat arbitrary since it is not always clear which of the datasets that is closest to the ground truth. For the sake of comparison, we always set the first dataset as truth. In order to illustrate the comparison, the contingency matrix as well as the temporal LWE series are shown.

In section 3.3.2.6 a summary with the tabular contingency matrices results for all the lakes that are inter-comparable between the two SAR algorithms are provided.

It is also important to mention that an Envisat ASAR WSM database is available. For some of the lakes we have found interesting data from the period 2002-2012 which are useful for some of the lakes that have had significant changes in the LWE in the period after 2000. ASAR WSM data has in general coarser spatial resolution, and only one polarization so the quality is in general poorer than for \$1 data.

3.3.2.1. Chad





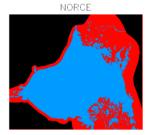
	LWE (km2)					
Date	NORCE	TRE- ALTAMIRA				
11/04/2019	354.127	324.364				
12/04/2019	339.114	299.972				
23/04/2019	354.471	308.309				
24/04/2019	346.339	335.068				

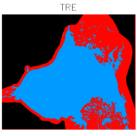
3.3.2.2. Illmen

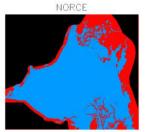
NORCE's and TRE-ALTA's lake water mask and their corresponding contingency matrices for a couple of coincident dates (20191024 and 20200116) which present a significant extent variation among them, are depicted in Figure 55. The overall agreement for those images is 97.73% and 96.19% respectively. Most of the differences between the two SAR approaches is related to the integration of inundated wetlands.

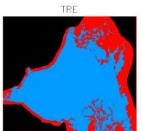
For Lake Illmen, the contingency matrix values for all NORCE's and TRE-ALTA's results corresponding to images presenting a temporal difference lower or equal to 12 days can be found in Table 15.

A representation of the temporal evolution of all NORCE's and TRE-ALTA's LWE is depicted in Figure 56. The same representation but just focusing on the 12 days difference images is depicted in Figure 57.









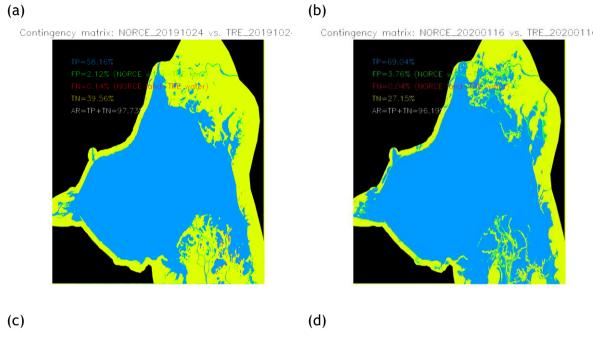


Figure 55. (a) NORCE and TRE-ALTA 20191024 lake water masks (blue) over a selected analysis polygon (red) and (c) its corresponding contingency matrix. (b) NORCE and TRE-ALTA 20200116 lake water masks (blue) over a selected analysis polygon (red) and (d) its corresponding contingency matrix.

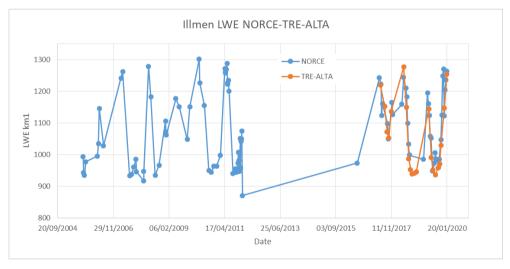


Figure 56. NORCE and TRE-ALTA LWE time series for Lake Illmen.



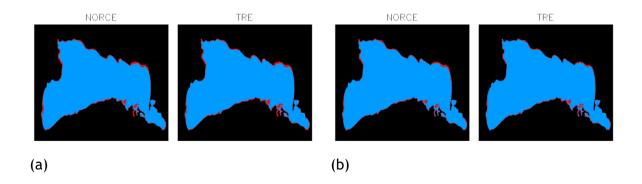
Figure 57. NORCE and TRE-ALTA 12 days maximum difference LWE time series for Lake Illmen.

3.3.2.3. Bosten

NORCE's and TRE-ALTA's lake water mask and their corresponding contingency matrices for a couple of coincident dates (20170729, 20190414 are depicted in Figure 58. The overall agreements for those images are 98.89% and 98.75% respectively.

For lake Bosten, the contingency matrix values for all NORCE's and TRE-ALTA's results corresponding to images presenting a temporal difference lower or equal to 12 days are represented in Table 17.

A representation of the temporal evolution of all NORCE's and TRE-ALTA's LWE is depicted in Figure 59. The same representation but just focusing on the 12 days difference images is depicted in Figure 60.



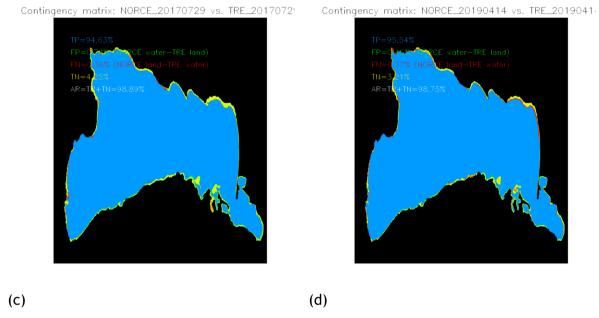


Figure 58. (a) NORCE and TRE-ALTA 20170729 lake water masks (blue) over a selected analysis polygon (red) and (c) its corresponding contingency matrix. (b) NORCE and TRE-ALTA 20190414 lake water masks (blue) over a selected analysis polygon (red) and (d) its corresponding contingency matrix.

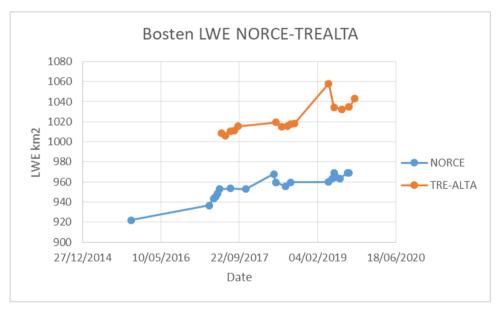


Figure 59. NORCE and TRE-ALTA LWE time series for lake Bosten.

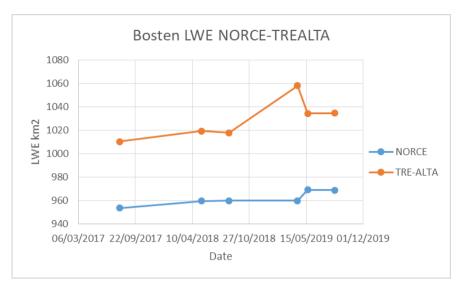


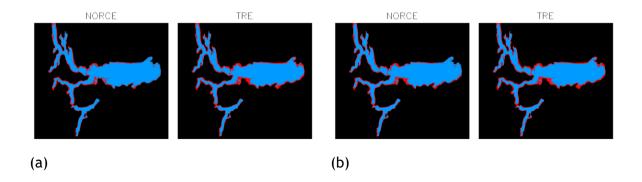
Figure 60. NORCE and TRE-ALTA 12 days maximum difference LWE time series for lake Bosten.

3.3.2.4. Argentino

NORCE's and TRE-ALTA's lake water mask and their corresponding contingency matrices for a 7 days difference couple (20171221-20171214) and a 5 days difference one (20191105-20191110) are depicted in Figure 61. The overall agreement for those images are 94.16% and 95.38% respectively.

For lake Argentino, the contingency matrix values for all NORCE's and TRE-ALTA's results corresponding to images presenting a temporal differences lower or equal to 12 days are represented in Table 19.

A representation of the temporal evolution of all NORCE's and TRE-ALTA's LWE is depicted in Figure 62. The same representation but just focusing on the 12 days difference images is depicted in Figure 63.



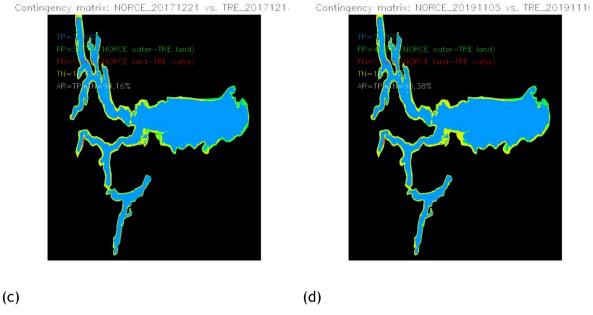


Figure 61. (a) NORCE and TRE-ALTA (20171221-20171214) lake water masks (blue) over a selected analysis polygon (red) and (c) its corresponding contingency matrix. (b) NORCE and TRE-ALTA (20191105-20191110) lake water masks (blue) over a selected analysis polygon (red) and (d) its corresponding contingency matrix.

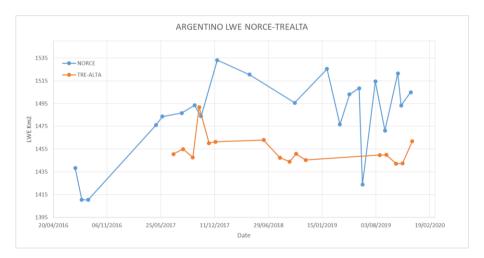


Figure 62. NORCE and TRE-ALTA LWE time series for lake Argentino.

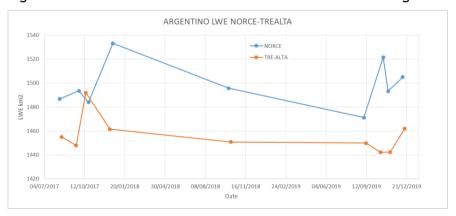


Figure 63. NORCE and TRE-ALTA 12 days maximum difference LWE time series for lake Argentino.

3.3.2.5. Khanka

NORCE's and TRE-ALTA's lake water mask and their corresponding contingency matrices for a couple of coincident dates (20180515, 20190510 are depicted in Figure 64. The overall agreements for those images are 98.83% and 97.62% respectively.

For lake Khanka, the contingency matrix values for all NORCE's and TRE-ALTA's results corresponding to images presenting a temporal difference lower or equal to 12 days are represented in Table 21.

A representation of the temporal evolution of all NORCE's and TRE-ALTA's LWE is depicted in Figure 65. The same representation but just focusing on the 12 days difference images is depicted in Figure 66.

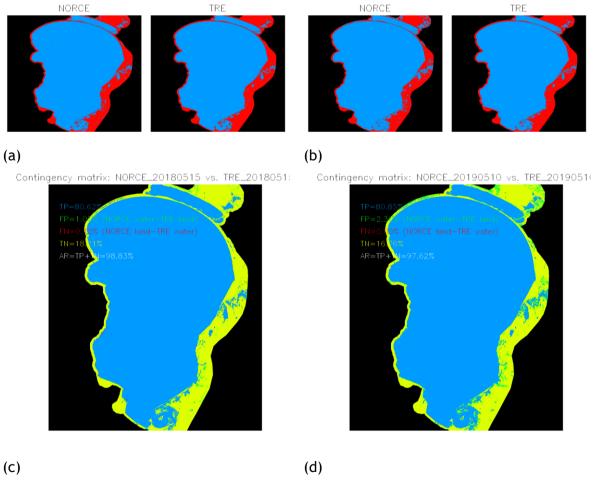


Figure 64. (a) NORCE and TRE-ALTA 20180515 lake water masks (blue) over a selected analysis polygon (red) and (c) its corresponding contingency matrix. (b) NORCE and TRE-ALTA 20190510 lake water masks (blue) over a selected analysis polygon (red) and (d) its corresponding contingency matrix.

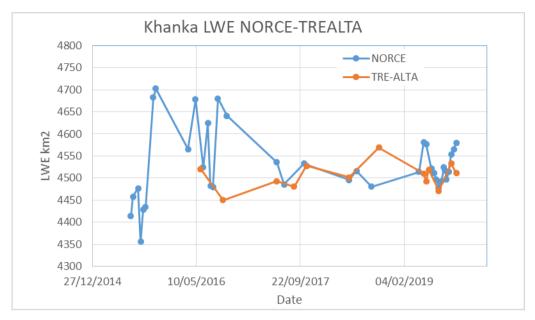


Figure 65. NORCE and TRE-ALTA LWE time series for lake Khanka.

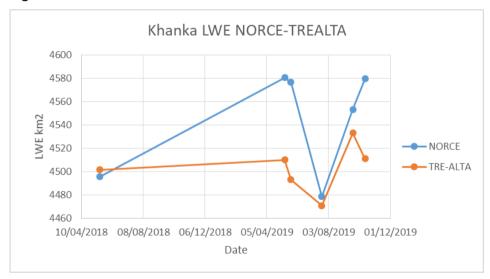


Figure 66. NORCE and TRE-ALTA 12 days maximum difference LWE time series for lake Khanka.

3.3.2.6. Overall comparisons between the two SAR algorithms

In this section we provide tabular contingency matrices for all the lakes that are intercomparable between the two SAR algorithms. Due to different selections of dates for the two SAR classifiers (NORCE and TRE-ALTA) different dates will be inter-compared. The main criterion for inter-comparison is that the two images (NORCE's and TRE-ALTA's) are close in time. For that, a maximum time difference to 12 days is set.

The following tables have been generated:

• Lake Illmen: contingency matrix in Table 15 and corresponding Area difference mean values in Table 16.

- Lake Bosten: contingency matrix in Table 17 and corresponding Area difference mean values in Table 18.
- Lake Argentino: contingency matrix in Table 19 and corresponding Area difference mean values in Table 20.
- Lake Khanka: contingency matrix in Table 21 and corresponding Area difference mean values in Table 22.
- Mean Values of all previous: Table 23

Tables' acronyms stand for: TP (True Positive in %), FP (False Positive in %), FN (False Negative in %), TN (True Negative in %), AR (Accuracy Rate in %), Td (time difference in days), Ad (Area LWE difference in %) and Ard (Area LWE difference in square kilometers).

After the comparison results, it can be stated the performance for the two SAR results is in general very similar, although differences are observed due to different dates/lakes and slight differences in methodology.

Table 15. NORCE and TRE-ALTA contingency matrix values summary for Lake Illmen

	Date 1	Date 2	TP	FP	FN	TN	AR	Td	Ad		Ard
Illmen	20170612	20170605	67.38	3.03	0.11	29.47	96.85	7	4.15		2.27
Illmen	20170916	20170909	59.83	3.37	0.25	36.52	96.36	7	4.92		25.97
Illmen	20170928	20171003	58.61	1.82	0.63	38.92	97.54	5	1.95		3.25
Illmen	20171115	20171108	63.48	3.58	0.04	32.88	96.36	7	5.28		28.01
Illmen	20180502	20180507	69.55	2.10	0.31	28.03	97.58	5	2.49		33.16
Illmen	20180607	20180612	64.11	5.58	0.00	30.29	94.40	5	7.99		59.6
Illmen	20180619	20180612	64.03	4.08	0.08	31.79	95.82	7	5.87		32.31
Illmen	20180713	20180718	55.73	3.77	0.16	40.33	96.06	5	6.06		45.68
Illmen	20180725	20180718	55.69	1.82	0.19	42.27	97.97	7	2.83		11.32
Illmen	20190427	20190502	64.06	2.82	0.13	32.97	97.03	7	4.03		16.86
Illmen	20190509	20190502	63.16	1.55	1.03	34.25	97.41	7	0.80		20.91
Illmen	20190602	20190607	56.13	4.49	0.01	39.35	95.49	5	7.39		61.91
Illmen	20190626	20190701	53.59	1.01	0.39	44.99	98.58	5	1.14		3.38
Illmen	20190801	20190806	53.16	4.78	0.00	42.04	95.21	5	8.25		69.85
Illmen	20190813	20190806	53.16	3.57	0.00	43.24	96.41	7	6.29		48.84
Illmen	20190918	20190911	54.02	1.22	0.36	44.37	98.40	7	1.55		0.82
Illmen	20190930	20191005	54.58	2.15	0.41	42.83	97.42	7	3.06		14.52
Illmen	20191024	20191024	58.16	2.12	0.14	39.56	97.73	0		3.27	17.92
Illmen	20191211	20191211	61.77	2.90	1.09	34.21	95.99	0	2.80		24.6
Illmen	20200116	20200116	69.04	3.76	0.04	27.15	96.19	0	5.10		9.67

Table 16. NORCE and TRE-ALTA mean and stdv LWE difference for Lake Illmen

Mean(Ar)	Stdv(Ar)	Mean(Ard)	Stdv(Ard)	
----------	----------	-----------	-----------	--

Illmen 96,74	1,09	26,54	20,97	
--------------	------	-------	-------	--

Table 17. NORCE and TRE-ALTA contingency matrix values summary for Lake Bosten

	Date 1	Date 2	TP	FP	FN	TN	AR	Td	Ad	Ard
Bosten	20170729	20170729	94.63	0.74	0.36	4.25	98.89	0	0.40	56.70
Bosten	20180513	20180513	95.12	0.85	0.32	3.69	98.81	0	0.55	59.68
Bosten	20180817	20180817	95.25	0.73	0.39	3.61	98.87	0	0.35	58.06
Bosten	20190414	20190414	95.54	0.46	0.77	3.21	98.75	0	0.32	98.22
Bosten	20190520	20190520	95.83	1.08	0.26	2.81	98.64	0	0.84	65.30
Bosten	20190824	20190824	95.95	0.94	0.20	2.90	98.85	0	0.76	65.78

Table 18. NORCE and TRE-ALTA mean and stdv LWE difference for Lake Illmen

	Mean(Ar)	Stdv(Ar)	Mean(Ard)	Stdv(Ard)
Bosten	98.80	0.09	67.29	15.61

Table 19. NORCE and TRE-ALTA contingency matrix values summary for Lake Argentino

	Date 1	Date 2	TP	FP	FN	TN	AR	Td	Ad	Ard
Argentino	20170811	20170816	76.49	4.24	1.11	18.14	94.63	5	3.86	31.7
Argentino	20170928	20170921	76.62	4.47	0.74	18.15	94.78	7	4.60	45.78
Argentino	20171022	20171015	76.99	3.59	2.31	17.09	94.09	7	1.58	7.94
Argentino	20171221	20171214	77.68	5.57	0.26	16.47	94.16	7	6.37	71.85
Argentino	20181005	20181010	76.56	4.65	0.83	17.94	94.51	5	4.70	44.94
Argentino	20190906	20190911	76.29	3.59	1.16	18.93	95.23	5	3.03	21.32
Argentino	20191024	20191017	76.76	5.86	0.32	17.04	93.81	7	6.71	79.42
Argentino	20191105	20191110	76.80	4.28	0.32	18.57	95.38	5	4.88	51.08
Argentino	20191211	20191216	77.78	3.94	0.46	17.80	95.58	5	4.25	43.09

Table 20. NORCE and TRE-ALTA mean and stdv LWE difference for Lake Argentino

	Mean(Ar)	Stdv(Ar)	Mean(Ard)	Stdv(Ard)
Argentino	94.68	0.61	44.12	22.50

Table 21. NORCE and TRE-ALTA contingency matrix values summary for Lake Khanka

	Date 1	Date 2	TP	FP	FN	TN	AR	Td	Ad	Ard
Khanka	20170601	20170601	80.58	1.83	0.02	17.56	98.14	0.00	2.19	6.01
Khanka	20180515	20180515	80.62	1.05	0.10	18.21	98.83	0.00	1.15	70.53
Khanka	20190510	20190510	80.85	2.36	0.00	16.76	97.62	0.00	2.83	83.5
Khanka	20190522	20190522	80.59	2.56	0.00	16.84	97.43	0.00	3.07	7.72
Khanka	20190721	20190721	80.21	1.15	0.01	18.61	98.83	0.00	1.40	20.23
Khanka	20190919	20190919	81.24	1.48	0.04	17.22	98.46	0.00	1.73	68.48
Khanka	20191013	20191013	80.88	2.32	0.00	16.78	97.66	0.00	2.78	6.01

Table 22. NORCE and TRE-ALTA mean and stdv LWE difference for Lake Khanka

	Mean(Ar)	Stdv(Ar)	Mean(Ard)	Stdv(Ard)
Khanka	98.13	0.58	42.74	35.15

Table 23. Mean Value for all previous contingency matrix AR, Td and Ad values

	AR	Ard	Td	Ad
Average	96.87	37.95	4.8	3.75

3.3.3. Comparison of optical and SAR sensors-based approaches

Similarly to what has been done in the previous section with the two SAR based methodologies, a comparison between the SAR and the optical LWE approaches is now done. On the SAR side Sentinel-1 and Envisat-SAR have been employed while on the optical side Sentinel2 and Landsat series have been used.

Temporal LWE series are generated and compared and for some of the lakes the contingency matrix is calculated whenever having images with a temporal difference equal or lower than 12 days.

3.3.3.1. Altevatnet

Comparison between SAR and optical LWEs have been carried out over more than 10 pairs of acquisitions, thanks to the high latitudes, the revisit is very high.

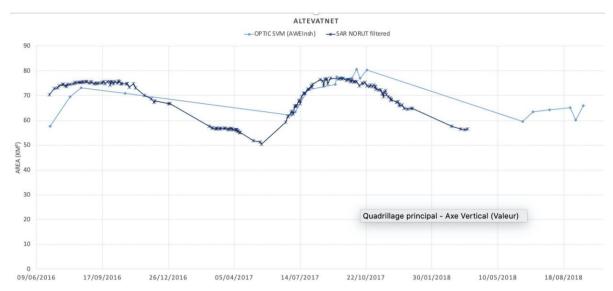


Figure 67. Comparison SAR and optical LWEs derived over Altevatnet

The value of the derived LWEs from optical and SAR are very coherent between them, with a correlation of 0.97

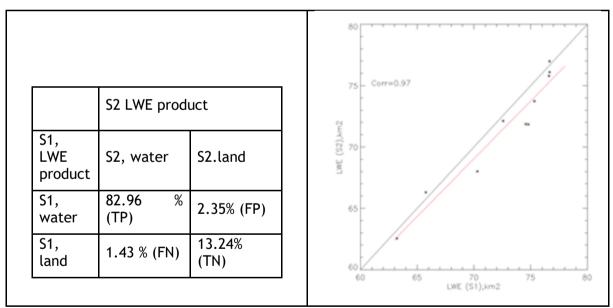


Figure 68. Analysis over 12 pairs of Sentinel1 and Sentinel2 of the commission and omission rates.

Of course, over this Nordic landscape, the observed commission and omission are the expected ones and are related to:

- Presence of cloud on optical imagery.
- Sensitivity to windy conditions and ice presence in SAR.

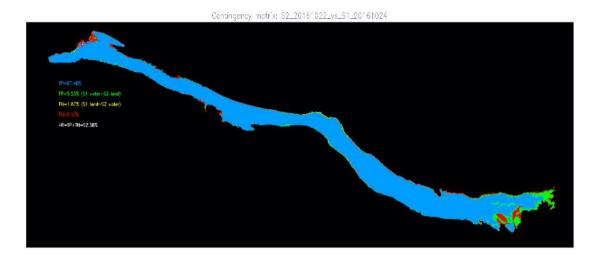


Figure 69. Analysis of the omission and commission between SAR and optical approaches exploiting Sentintel1 and Sentinel2.

3.3.3.2. Colhue

The period with dual observation is relatively limited, the trends are similar but with apparently an overestimation of water surface by SAR. This could be related to the nearby environment, i.e. desert and also sandy shore.

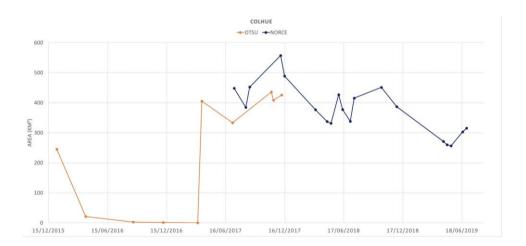


Figure 70. Comparison SAR and optical LWEs derived over Colhue Lake.

3.3.3.3. Al Hamar wetland

There are important differences between SAR and optical observations over the complex area of Al Hamar wetlands. Whereas optical imagery allows us to monitor the increase of

water surfaces, SAR LWEs presents a drop at the beginning of the series, and then stay at a lower level than optical. This is due to the facts:

- Optical LWE integrates a part of flooded vegetation whereas SAR retrieves open water surface.
- SAR is more sensitive to the environment, i.e. sandy flat area, than optical sensors.



Figure 71. Comparison SAR and optical LWEs derived over Al-Hammar.

3.3.3.4. Sassykol and Alakol

This case of study is particularly interesting, as it is two neighbouring lakes being connected through a wetland complex. In addition, Sassykhol is a relatively shallow lake, with surrounding wetland, whereas Alakol is a deep lake with a more classical shape.

Over Alakol lake, the difference of LWE between SAR derived LWE and Optical ones, is about 3% that is quite satisfactory.

Therefore, over Sasykholol, the difference is about 14%, this is related to the fact large part of the increase of water surface is related to the flooding of wetland, which is not observed by SAR.

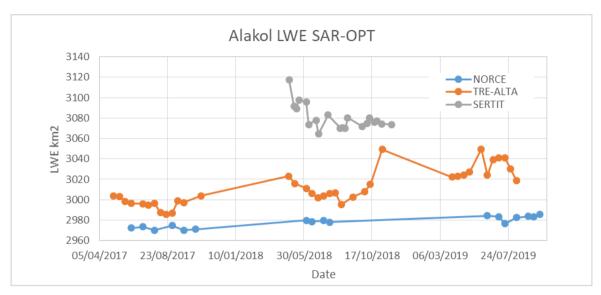
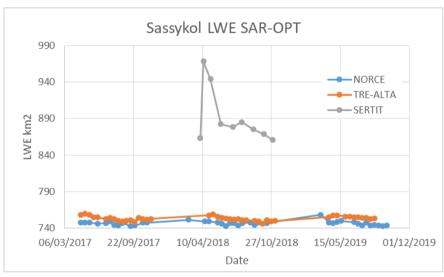


Figure 72. Comparison SAR and optical LWEs derived over Sassykol and Alakol lakes.



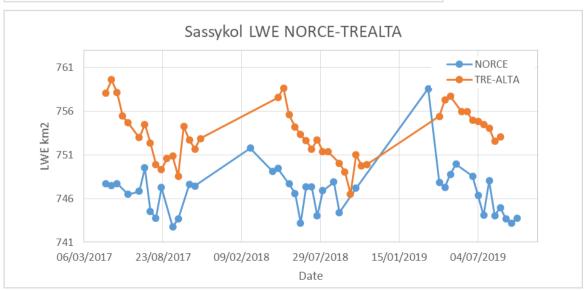


Figure 73. Comparison SAR and optical LWEs derived over Sassykol and Alakol lakes.

3.3.3.5. Chad

There were near simultaneous acquisition of Sentniel1 SAR image and Sentinel2 optical images on the 24^{th} of April 2019. This gives the opportunity to compare the capabilities and limitations of the two system.

Within the common AOI defined there were 14.68 km² detected as water in Sentinel 1 and none water in Sentinel 2. These areas correspond partially to very small and located clouds that have intercepted the optical signal. In this case SAR, and not the optical sensor, were closest to the truth. Therefore, most of the "SAR water alone" correspond to commission with bare soil on the NE part of the study area.

At the opposite, 40.75 km² detected as water on Sentinel 2 and no water on Sentinel 1. These areas correspond mostly to shoreline of the lake, plus some floating islands.

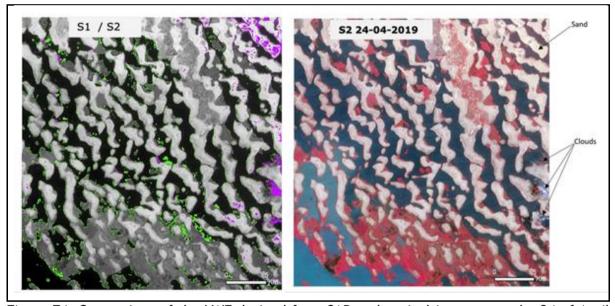


Figure 74. Comparison of the LWE derived from SAR and optical imagery on the 24 of April 2019: in Purple, detected as water on S1 and none water based on S2; in Green detected as water on S2 and none water on S1.

3.3.3.6. Chilwa lake

The difference between optical LWE and SAR classification is about 20%. This is related to the fact that large parts of the increase in water surface is related to the flooding of wetland, which is not observed by SAR. In this specific complex context, the SAR extraction seems underestimated over wetlands.

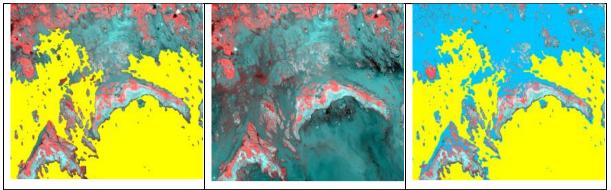


Figure 75. Comparison of SAR and optical LWEs; in yellow LWE derived from SAR with in background the Sentinel2, False colour Sentintel2 acquired the 24/04/2018; in blue the additional water surfaces derived from a Sentinel-2 image with the SVM algorithm.

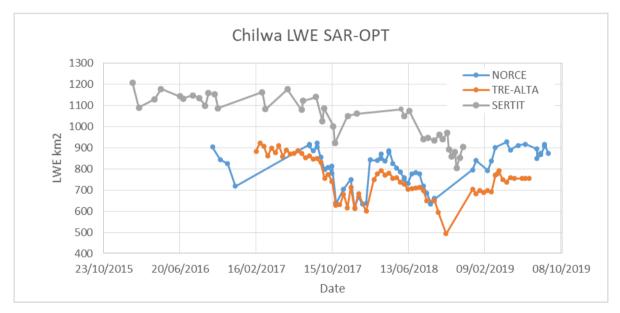


Figure 76. Comparison SAR and optical LWEs derived over Chilwa lake.

3.3.3.7. Argentino

NORCE-SERTIT and TRE-ALTA-SERTIT lake water mask and their corresponding contingency matrices for a couple of couple of images presenting a time difference smaller than 12 days (20171022-20171015 and 2071015-20171020 respectively) are depicted in Figure 77. The overall agreements for those images are 96.66% and 95.82% respectively.

For Lake Argentino, the contingency matrix values for all NORCE and SERTIT LWE and TRE-ALTA and SERTIT corresponding to images with a temporal difference lower or equal to 12 days are represented in Table 25 and Table 30 respectively.

A representation of the temporal evolution of all NORCE, TRE-ALTA and SERTIT LWE is depicted in Figure 78. The same representation but just focusing on the 12 days difference images is depicted in Figure 79.

Argentino is a lake surrounded by steeply mountains. The shape of the lake is defined by narrow valleys. Due to the altitude, the lake is often roughened by wind, and this is a challenge for the SAR retrieval of the LWE. When comparing LWEs derived from Sentinel1 and Sentinel2 acquired close in time, the differences are:

- Commission errors for SAR on lakeshore, particularly the Eastern parts of the lake, where the sandy, muddy shores are relative flat and soft surfaces that has a radar signature similar to water.
- Some SAR omission-errors that could be related to windy water surface, but also in the north branch of the lake, to some commission errors from the glacier terminus or lake ice.

Depending on the applied processing, the difference in terms of detected water surface can be significant (more than 5%) or relatively small (less than 1% - 3.3%).

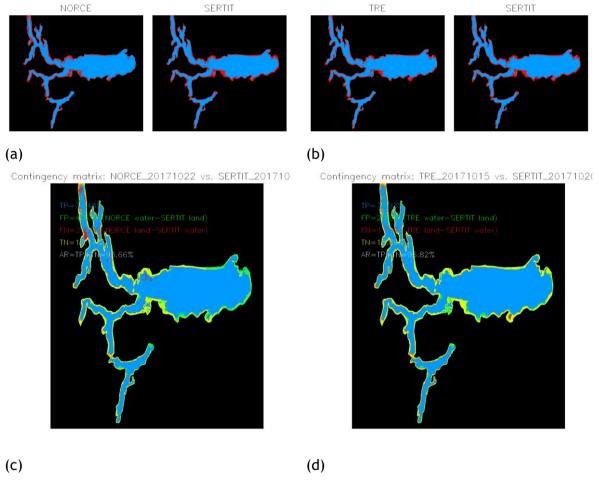


Figure 77. (a) NORCE-SERTIT (20171022-20171015) lake water masks (blue) over a selected analysis polygon (red) and (c) its corresponding contingency matrix. (b) TRE-ALTA-SERTIT (2071015-20171020) lake water masks (blue) over a selected analysis polygon (red) and (d) its corresponding contingency matrix.

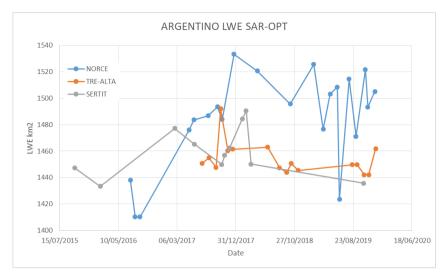


Figure 78. NORCE SAR, TRE-ALTA SAR and SERTIT LWE time series for lake Argentino.

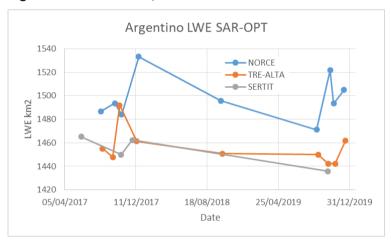


Figure 79. NORCE SAR, TRE-ALTA SAR and SERTIT Optical LWE time series for lake Argentino.

3.3.3.8. Illmen

In the Illmen lake case no temporally close images were processed among the three groups. In Figure 80, the three LWE series are depicted. From it, it is hard to evaluate SAR and Optical relative performances.

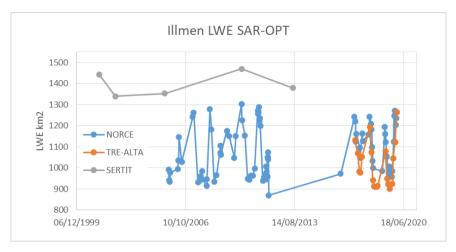


Figure 80. NORCE SAR, TRE-ALTA SAR and SERTIT Optical LWE time series for lake Illmen.

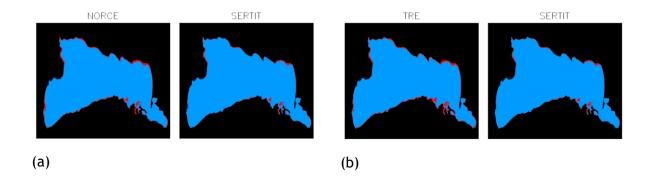
3.3.3.9. Bosten

NORCE-SERTIT and TRE-ALTA-SERTIT lake water mask and their corresponding contingency matrices for a couple of couple of images with a time difference smaller than 12 days (20170729-20170728 for both cases) are depicted in Figure 81. The overall agreement for those images are 98.02% and 97.65%, respectively.

The systematic bias between optical and SAR LWE-estimates is 4% to 8% depending on the SAR approach. This is related to the fact that large part of the increase in water extent occurs in wetlands east of the main lake, and SAR has limited capabilities in resolving the water in such wetlands. The bias between the NORCE and the TRE ALTAMIRA approach for Bosten is related to the use of different masks.

For Lake Bosten, the contingency matrix values for all NORCE and SERTIT LWE and TRE-ALTA and SERTIT corresponding to images with a temporal difference lower or equal to 12 days are represented in Table 24 and Table 29 respectively.

The temporal evolution of all NORCE, TRE-ALTA and SERTIT LWE-estimates are depicted in Figure 82. The same representation but just focusing on the 12 days difference images is depicted in Figure 83.



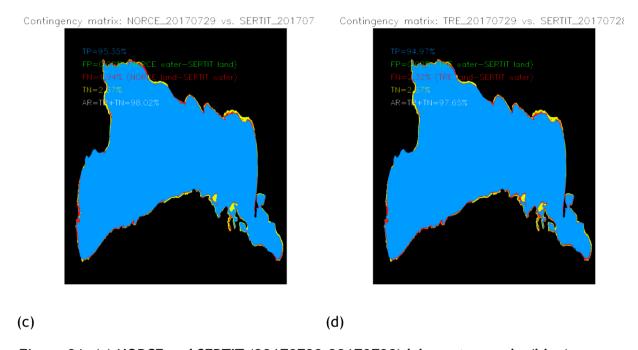


Figure 81. (a) NORCE and SERTIT (20170729-20170728) lake water masks (blue) over a selected analysis polygon (red) and (c) its corresponding contingency matrix. (b) TRE-ALTA and SERTIT (20170729-20170728) lake water masks (blue) over a selected analysis polygon (red) and (d) its corresponding contingency matrix.

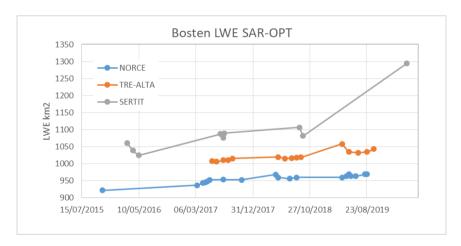


Figure 82. NORCE SAR, TRE-ALTA SAR and SERTIT LWE time series for lake Bosten.

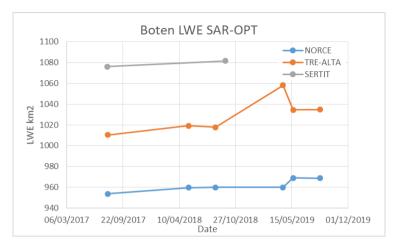


Figure 83. NORCE SAR, TRE-ALTA SAR and SERTIT Optical LWE time series for lake Bosten.

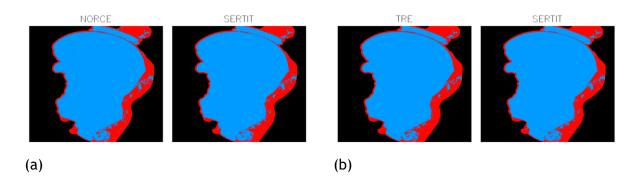
3.3.3.10. Lake Khanka

NORCE-SERTIT and TRE-ALTA-SERTIT lake water mask and their corresponding contingency matrices for a couple of couple of images with a time difference smaller than 12 days (20180515-20180510, 20180515-20180510 respectively) are depicted in Figure 84. The overall agreement for those images are 98.18% and 98.34%, respectively.

For Lake Khanka, the contingency matrix values for all NORCE and SERTIT LWE and TRE-ALTA and SERTIT corresponding to images with a temporal difference lower or equal to 12 days are represented in Table 26 and Table 31Table 29 respectively.

The temporal evolution of all NORCE, TRE-ALTA and SERTIT LWE-estimates are depicted in Figure 85. The same representation but just focusing on the 12 days difference images is depicted in Figure 86.

The difference between SAR and optical LWE is about 50 km², which represents about 1.1% of the water lake extent. It is interesting to compare pair by pair the evolution of commission and omission between the LWEs derived from Sentinel1 and from Sentinel2. Water detection errors in optical imagery is often related to poor cloud discrimination. In these examples, the omission and commission errors are very low. SAR omission, in red in the figure—corresponds to vegetated areas along the lake shore, areas than can be inundated. In the pair of the end of July 2019, effects of the presence of a clouds is noticeable, inducing both omission as water is not recognized and commission over the land surface. This example highlights the sensitivity and limitations of exploiting optical imagery for the recognition of water surface due to the presence of clouds.



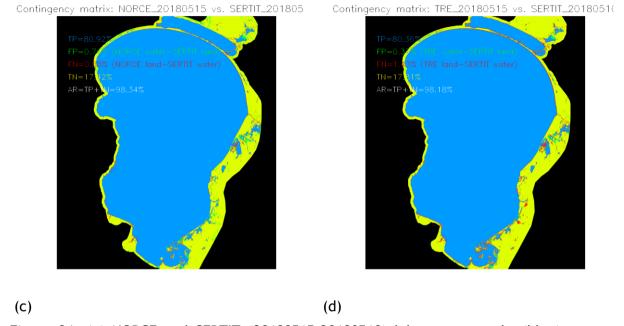


Figure 84. (a) NORCE and SERTIT (20180515-20180510) lake water masks (blue) over a selected analysis polygon (red) and (c) its corresponding contingency matrix. (b)TRE-ALTA-SERTIT (20180515-20180510) lake water masks (blue) over a selected analysis polygon (red) and (d) its corresponding contingency matrix.

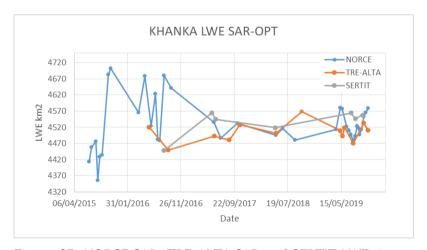


Figure 85. NORCE SAR, TRE-ALTA SAR and SERTIT LWE time series for lake Khanka.

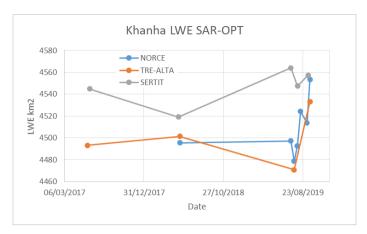


Figure 86. NORCE SAR, TRE-ALTA SAR and SERTIT Optical LWE time series for lake Khanka.

3.3.3.11. Sarykamish Lake

As already indicated, SaryKamish is located in a desert zone with very limited vegetation. The landscape consists mostly in sandy and muddy bare soil.

In the SAR LWE there is an important commission error between land and water of 5.3 %. This is related to the nature/texture/structure of the soil surrounding the lake. The area viewed as water in the SAR image, corresponds to a bare crust having similar backscatter as nearby water body.

The omissions are relatively scarce and corresponds to rough water surface and inundated vegetation within a delta-zone.

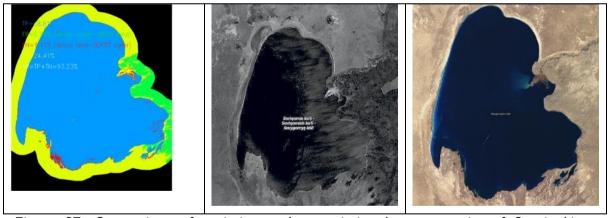


Figure 87: Comparison of omission and commission between pairs of Sentinel1, and Sentinel 2 acquired respectively the 2019-03-07, and 2019-03-11 over the Sary Kamish lake.

3.3.3.12. Summary of all inter-comparisons between optical and SAR classification

In this section we provide tabular contingency matrices for all the lakes that are intercomparable between the two SAR and the Optical algorithms. Due to different selections of dates for the three approaches different dates will be inter-compared. The main criterion for inter-comparison is that the pair of images to compare (either coming from any of NORCE's,TRE-ALTA's and SERTIT's) are close in time. For that, a maximum time difference to 12 days is set.

The following tables have been generated:

- Lake Bosten: NORCE-SERTIT Table 24 and TRE-ALTA -SERTIT Table 29.
- Lake Argentino: NORCE-SERTIT Table 25 and TRE-ALTA -SERTIT Table 30.
- Lake Khanka: NORCE-SERTIT Table 26 and TRE-ALTA -SERTIT Table 31.
- Mean Values of all previous: NORCE-SERTIT Table 28 and TRE-ALTA Table 32.

Tables' acronyms stand for: TP (True Positive in %), FP (False Positive in %), FN (False Negative in %), TN (True Negative in %), AR (Accuracy Rate in %), Td (time difference in days), Ad (Area LWE difference in %) and Ard (Area LWE difference in square kilometres).

Table 24. NORCE and SERTIT contingency matrix values summary for Lake Bosten

	Date 1	Date 2	TP	FP	FN	TN	AR	Td	Ad	Ard
Boster	20170729	20170728	95.35	0.02	1.94	2.67	98.02	1	2.01	122.45

Table 25. NORCE and SERTIT contingency matrix values summary for Lake Argentino

	Date 1	Date 2	TP	FP	FN	TN	AR	Td	Ad	Ard
Argentino	20170531	20170602	78.35	2.21	1.20	18.22	96.58	3	1.26	18.39
Argentino	20171022	20171020	76.48	4.10	2.22	17.17	93.66	2	2.34	34.02
Argentino	20191024	20191015	77.12	5.51	0.85	16.50	93.63	9	5.64	85.87

Table 26. NORCE and SERTIT contingency matrix values summary for Lake Khanka

	Date 1	Date 2	TP	FP	FN	TN	AR	Td	Ad	Ard
Khanka	20180515	20180510	80.92	0.76	0.88	17.42	98.34	5	0.14	23.68

Table 27. NORCE and SERTIT contingency matrix values summary for Lake Sarykamish

	Date 1	Date 2	TP	FP	FN	TN	AR	Td	Ad	Ard
Sarykamish	20190307	20190311	68.81	5.64	1.11	24.41	93.23	4	6.07	217,78

Table 28. Mean Value for all previous NORCE and SERTIT contingency matrix AR, Td and Ad values

	AR	Td	Ad	Ard
Average	94.69	4,1	4.01	83,69

Table 29. TRE-ALTA and SERTIT contingency matrix values summary for Lake Bosten

	Date 1	Date 2	TP	FP	FN	TN	AR	Td	Ad	Ard
Bosten	20170729	20170728	94.97	0.02	2.32	2.67	97.65	1	2.42	65.74
Bosten	20180910	20180901	95.64	0.03	1.89	2.42	98.07	9	1.94	87.72

Table 30. TRE-ALTA and SERTIT contingency matrix values summary for Lake Argentino

	Date 1	Date 2	TP	FP	FN	TN	AR	Td	Ad	Ard
Argentino	20171015	20171020	76.93	2.38	1.78	18.89	95.82	5	0.76	41.96
Argentino	20171120	20171129	77.41	0.71	1.99	19.87	97.29	9	1.63	2.24
Argentino	20191017	20191015	75.86	1.22	2.10	20.79	96.66	8	1.14	6.45

Table 31. TRE-ALTA and SERTIT contingency matrix values summary for Lake Khanka

	Date 1	Date 2	TP	FP	FN	TN	AR	Td	Ad	Ard
Khanka	20170601	20170609	80.40	0.21	1.91	17.47	97.87	8.00	2.10	9,10
Khanka	20180515	20180510	80.36	0.37	1.43	17.81	98.18	5.00	1.31	23.68
Khanka	20190919	20190912	80.91	0.38	1.50	17.19	98.11	7.00	1.38	4.01

Table 32. Mean Value for all previous TRE-ALTA and SERTIT contingency matrix AR, Td and Ad values

	AR	Td	Ad	Ard
Average	94.82	5.5	5.41	30,11

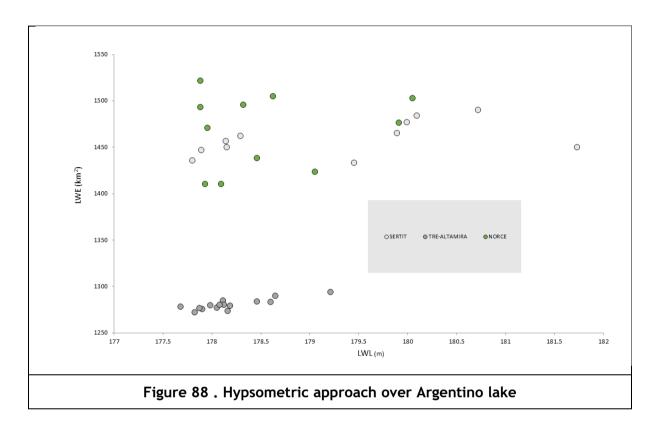
3.3.4. Results analysis for hypsometric approach

The objectives are to compare the results of lake area from different methods and type of images in term of building resulting hypsometry curve. This approach was tested over the second set of lakes, Khanka (China_Russia), Illmen (Russia), Sary Kamysh lake (Turkménistan - Ouzbékistan), Bosten lake (China), and Argentino, (Argentina).

3.3.4.1. Hypsometric approach for the Argentino Lake

Over Argentino Lake, three series were compared: one from optical imagery and two from SAR imagery. Following remarks can be dressed

we see that SERTIT and NORCE solutions have the same mean value of lake water extent, since it is much lower for TRE-Altamira. The optical solution (SERTIT) presents non monotonic trend but this is likely due to two evident outliers as seen on the figure. The NORCE solution looks very disturbed with high dispersion of the hypsometry data. The TRE-Altamira looks quite coherent, with low RMS and monotonic trend but over a lower range of water level observations (no data for high water level above 180 m as for the 2 other solutions).



The hypsometry coefficient can also be calculated using each of the solution with also the estimation of uncertainty, which here is characterized by the RMS of the difference between the theorical hypsometry and the data used to calculate it.

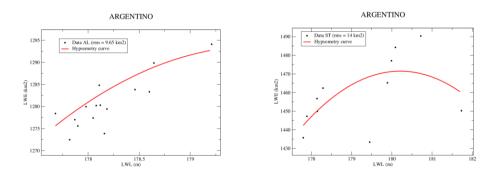


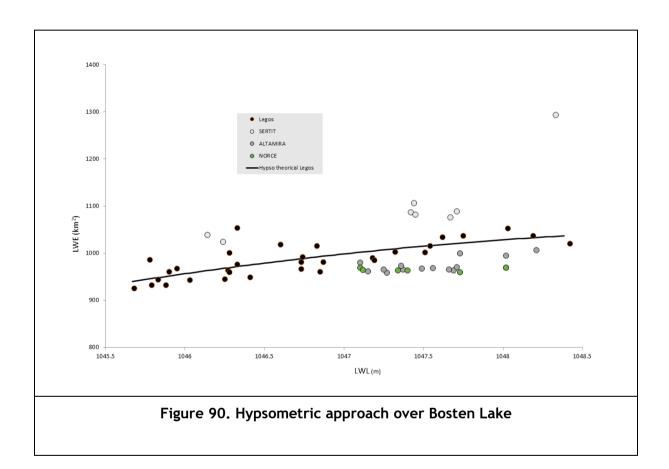
Figure 89. Hypsometry coefficient for Argentino lake

We see from this figure that the sensitivity of hypsometry to outliers (like for the optical solution: SERTIT) can be very high.

3.3.4.2. Hypsometric approach for the Bosten Lake

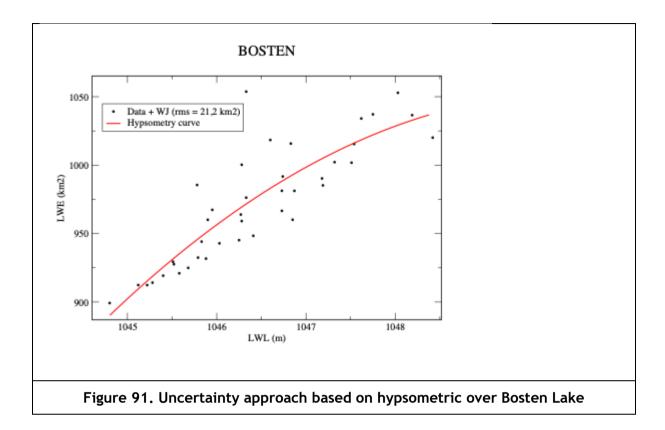
Over Bosten Lake, four series were compared, two from SAR imagery and the two from optical data. Following remarks can be dressed:

- The 4 solutions look quite coherent
- It remains some outliers



It highlights the need to extract water extent of the lake over the largest range of water level variations. This minimizes the impact of dispersion of the solution on the calculation of the hypsometry coefficients, although it may increase the RMS.

The uncertainty using the theorical hypsometry curve has been calculated only with LEGOS solution. The RMS of LEGOS Solution is of 5%



3.3.4.3. Hypsometric approach for the Illmen lake

Over Illmen Lake, three series were compared, two from SAR imagery and one from optical data. Following remarks can be dressed:

•Optical imageries look (Sertit) overestimates the LWE when compared to the 2 radar solutions, but not enough images have been processed

We observed from the two radar imagery datasets a small bias between both solutions: NORCE slightly overestimates the extent compared to TRE-Altamira. Both series looks however consistent, with monotonic trend and a good coverage of the water level changes over time

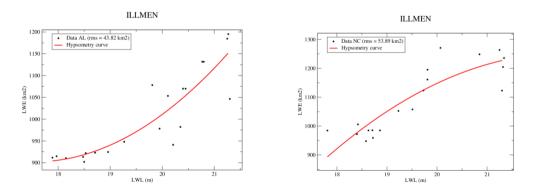


Figure 92. Lake Illmen. Theorical hypsometry for the 2 radar solutions

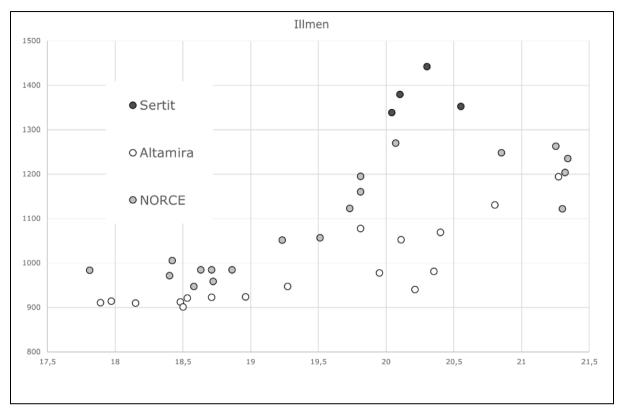


Figure 93. Hypsometric approach over Illmen Lake

3.3.4.4. Hypsometric approach for the Khanka Lake

Over Khanka Lake, four series were compared, two from SAR imagery and the two from optical data. Following remarks can be dressed:

•Very complicate case: 2 groups of solutions (optical Sertit/SAR NORCE & optical LEGOS/SAR Altamira) with strong bias

The LEGOS solution has a strong RMS, due to high dispersion, but has a monotonic trend which is more realistic. The NORCE solution cumulate monotonic trend and relatively low RMS, but over a much shorter range of observations.

•The 2 radar solutions don't have variation of LWE w.r.t variations of LWL (no trends).

They both doesn't have enough images processed. This is a typical case where hypsometry analysis does not really help building realistic relationship between LWE and LWL.

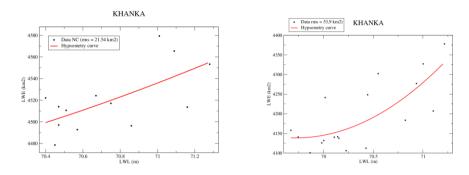


Figure 94. Theorical hypsometry from NORCE and LEGOS solutions

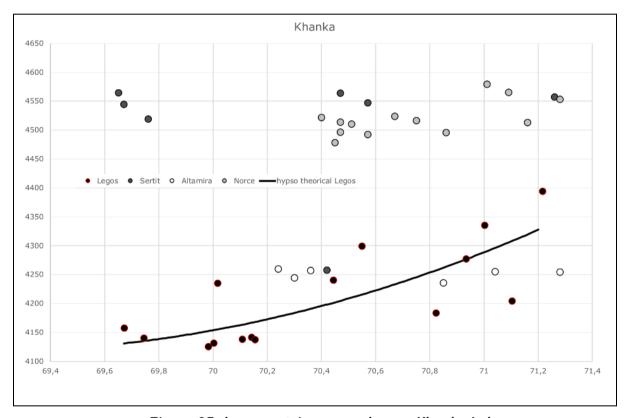


Figure 95. hypsometric approach over Khanka Lake

The uncertainty using the theorical hypsometry curve has been calculated only with LEGOS solution. The RMS of LEGOS Solution is of 1,3%

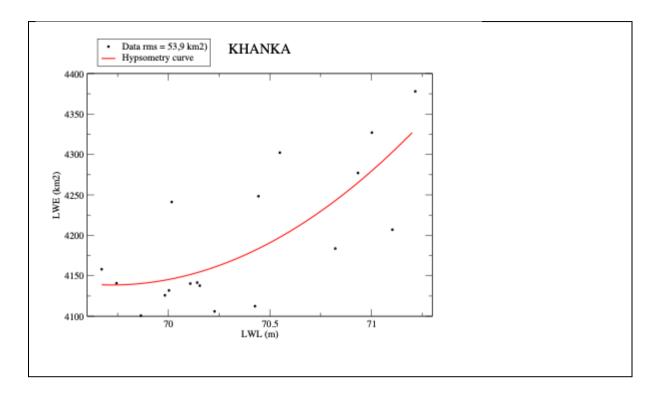


Figure 96. Uncertainty approach based on hypsometric over Khanka Lake

3.3.4.5. Hypsometric approach for the Sary -Kamisk Lake

Over Sary-Kamish Lake, four series were compared, two rom SAR imagery and wo from optical data. Following remarks can be dressed:

- The tow optical solutions are very coherent together (remains 2 outliers in the Sertit solution)
- Due to high LWL variation in time, LEGOS solution is uncompleted for small LWL
- It remains some outliers in the SERTIT solution
- the two solutions based on radar imagery are limited to a too low range of variations, since the water level change of this lake has strongly increased over the last 20 years. Radar imageries has been obtained only over the last 3-4 years when the lake was already at high level, as we may see from the X axis for these two solutions (level around 8 meters). This explains that the shape of the hypsometry depends too much on the dispersion of the LWL/LWE data and in such case hypsometry cannot be calculated using only these data.
- The uncertainty of the hypsometry using the two optical imagery's solution is lower than 1% of the total extent of the lake

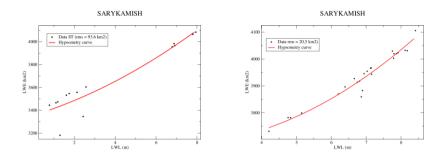


Figure 97 Theorical hypsometry from SERTIT and LEGOS solutions

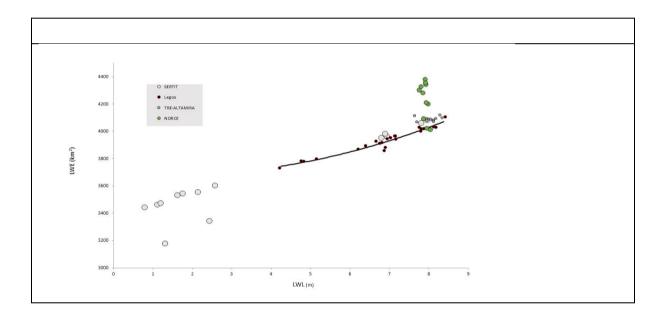


Figure 98. Hypsometric approach over SaryKamish Lake

The conclusion here is that for the Sarykamish lake, we may simply remove the two outliers (seen on the Sertit solution) and recalculate the hypsometry coefficient using the Legos, the TRE-Altamira and the Sertit solutions. It will allow to extend the domain of validity of the hypsometry from very low to very high lake water level.

The uncertainty using the theorical hypsometry curve has been calculated only with LEGOS solution. The RMS of LEGOS Solution is of 1,3%

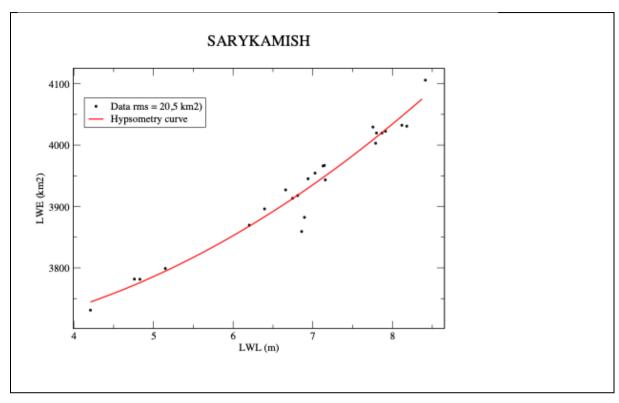


Figure 99. Uncertainty approach based on hypsometric over SaryKamish Lake

3.4. Conclusions and recommendations

We analysed the behaviour of two supervised and one unsupervised approaches on five lakes with various characteristics that provided a relatively good overview of what can be observed when estimating LWE from high-resolution optical imagery.

Each method has its weaknesses and strengths. Of course, we had to deal with the expected classical traps which are related to external condition such as clouds cover and associated shadows, local environment, the presence of vegetation within the water bodies, floating or submerged ones. Many of the characteristics of water conditions are related to the colour of the water in shallow water.

One main problem is also the definition of LWE. Do we have to consider open free water or open free water plus ice cover? In addition, are we speaking about open water, in case of water bodies surrounding or imbricated with wetlands? The delimitation between these entities is not so obvious. Do we have to consider a floating island of vegetation as water or not? Depending of the followed approach and or the exploited indices nor training samples selection process, omission or commission would /could happen.

The conclusion and recommendations at this stage, after two sets of benchmarking of methods, are the following:

• For optical data exploitation, the best method is a random forest one with a testing procedure on the training set based on Pekel GWS database.

• For SAR processing, KMeans approaches are satisfactory, but in some challenging environments, such as deserts or lakes with very fragmented or vegetated shoreline, it can result in large errors. Estimating the LWE during ice cover is also a challenging task, although it can be possible during the cold winter period (Vickers et al., 2019).

For an operational application, i.e. monitor LWE based on LWL, this require to generate de precise hypsometric curves, a work to be done lake after lake. More works are needed in order to select the appropriate images, to access these images as there are some restriction in term of data access exploiting either GEE, nor ESA SciHub.

4. Lake Surface Water Temperature (LSWT)

The Lake Surface Water Temperature LSWT is defined as the temperature of the water at the surface of the water body (surface skin temperature). The CCI LSWT dataset consists of a long-term climate data record (CDR). The validation and comparison of the LSWT is based on matchups between in situ and satellite measurements. In situ measurements are collated once per year from limnologists worldwide who are willing to share their in-situ data of lake surface temperature measurements on a personal, ad hoc basis supplementing the partial collections of some agencies. The annual collation is undertaken towards the end of the calendar year and involves a considerable effort every year to enlarge and quality-control the in-situ database with measurements that are suitable for satellite validation. Most of the data is collated through personal communications.

4.1. Data description

4.1.1. Satellite data

LSWT time series have been computed from sensors on multiple satellites and lake-specific consistency adjustments between sensors have been applied using the MetOpA AVHRR instrument as a reference (see RD- 1). In the current version, LSWT v4.3, MetOpA AVHRR has the best combination of length of record and data density for this purpose. For the LSWTs derived from MODIS, only quality levels 4 and 5 have been used for the dataset and a constant bias correction (0.11 for quality level 5 and 0.19 for quality level 4) has been applied across all lakes after a comparison with the LSWTs from MetOpA AVHRR. The same form of algorithms has been used to retrieve the LSWT from all sensors in order to obtain consistent time series for each of the CCI target lakes. The target list can be found at RD-3.

The time periods used for each satellite/instrument are provided in Table 33. Not all lakes include LSWT from all sensors in the series because of differing density and geometry of observation. The temperatures in the dataset are only available for cloud and ice-free observations during (in this version) day time, so gaps in time and space are common for all the lakes due to cloud cover and swath geometry of the instruments.

Table 33 - Time periods for	the satellite/instrument	used to generate the LSWTs

Satellite	Instrument	Time Period
ERS-2	ATSR2	Jun1995 - Jun2003
Envisat	AASTR	May2002 - Apr2012
MetOpA	AVHRR	Mar2007 - Aug2019
MetOpB	AVHRR	Dec2012 - Aug2019
Terra	MODIS	Feb2000 - Dec2020
Sentinel3A	SLSTR	Jun2016 - Dec2020
Sentinel3B	SLSTR	Aug2020 - Dec2020

A detailed description of the product generation is provided in the Algorithm Theoretical Basis Document (ATBD RD- 1) with further information on the product given in the Product User Guide (PUG RD- 3).

4.1.2. In situ data

The in situ dataset currently used for validation has been constructed from the in situ temperature data collected through the ARCLake project, the GloboLakes project, the EU Surface Temperature for All Corners of Earth (EUSTACE) project and the Copernicus Climate Change Service (C3S) product. At present, this dataset consists of 272 observation locations covering 132 lakes. However, the number of lakes are 79 with a total of 208 sites. Details of the in situ observation locations with their sources are given in Table 34 which reports all locations for the target lakes where there are matches. The geographical distribution of the sites is reported in Figure 100 which shows that most of the globe is covered with a big proportion of the sites located in North America and Europe.

Table 34 - List of the sources of the in situ data

Source	Lake names (number of observation sites)
NDBC - National Data Buoy Centre (USA)	Superior (3) Huron (2) Michigan (2) Erie (1) Ontario (1)
FOC - Fisheries and Oceans Canada (Canada)	Superior (1) Huron (4) Great Slave (2) Erie (2) Winnipeg (3) Ontario (4) Woods (1) Saint Claire (1) Nipissing (1) Simcoe (1)
Michigan Technological University (USA)	Superior (2) Michigan (1)
University of Minnesota (USA)	Superior (2)
Northern University of Michigan (USA)	Superior (3)
Superior Watershed Partnership (USA)	Superior (1)
U.S. Army Corps of Engineers (USA)	Superior (1)
Technical University of Kenya (Kenya)	Victoria (1)
GLERL - Great Lakes Environmental Research Lab (USA)	Huron (3) Michigan (2)
University of Wisconsin-Milwaukee (USA)	Michigan (2)
Northwestern Michigan College (USA)	Michigan (1)
University of Michigan CIGLR (USA)	Michigan (2)
Limno Tech (USA)	Michigan (3) Erie (4)
Illinois-Indiana Sea Grant and Purdue Civil Engineering (USA)	Michigan (2)
Leibniz Institute for Freshwater Ecology and Inland Fisheries (Germany)	Tanganyika (1)
Pierre Denis Plisnier	Tanganyika (4)

Source	Lake names (number of observation sites)
Irkutsk State University (Russia)	Baikal (1)
Regional Science Consortium (USA)	Erie (1)
UGLOS - Upper Great Lakes Observing System (USA)	Erie (2) Douglas (1)
LEGOS - Laboratoire d'Etudes en Géophysique et Océanographie Spatiales (France)	Issykkul (1)
SLU - Swedish University of Agricultural Science (Sweden)	Vanern (6) Vattern (2) Malaren (9) Hjalmaren (1) Siljan (1) Bolmen (2) Ekoln (1) Roxen (1)
Uppsala University (Sweden)	Vanern (1) Erken(1)
Sao Paulo State University (Brazil)	Tucurui (1) Itaipu (1) Tres Marias (1) Serra da Mesa (1) Itumbiara (1)
Junsheng Li (China)	Taihu (1)
KU Leven (Belgium)	Kivu (1)
SYKE - Finnish Environment Institute (Finland)	Inarinjarvi (1) Paijanne (3) Pielinen (4) Oulujarvi (1) Keitele (1) Nasijarvi(1) Lokan (1) Onkivesi (1) Puulavesi (1) Hoytiainen(1) Koitere(1) Vanajavesi (1) Pyhajarvi(1) Lappajarvi (1) Mallasvesi(2) Vuohijarvi(1) Lentua (1) Myekojarvi (1) Pyhajarvi (1)
Vermont EPSCOR - Established Program to Stimulate Competitive Research (USA)	Champlain (1)
SUNY Plattsburgh Center for Earth and Environmental Science (USA)	Champlain (1)
Nipissing University (Canada)	Nipissing (2)
National Park Service (USA)	Mead (3) Mohave (2)
NIWA (New Zealand)	Taupo (3) Rotorua (1)
GLEON - Global Lake Ecological Observatory Network	Tanganyika (3) Balaton (1)
BLI - Balaton Limnological Institute (Hungary)	Balaton (6)
KDKVI - Central Transdanubian (Regional) Inspectorate for Environmental Protection, Nature Conservation and Water Management (Hungary)	Balaton (3)
Hungarian Met Service (Hungary)	Balaton (1)
UMR CARRTEL - Centre Alpin de Recerche sur le Réseaux Trophique des Ecosystèmes Limniques (France)	Geneva (1)

Source	Lake names (number of observation sites)			
UC-Davis Tahoe Environmental Research Center (USA)	Tahoe (1)			
Utrecht University (Nederlands)	Garda (1)			
Italian National Research Council (Italy)	Garda (8) Trasimeno (2) Maggiore (2) Como Mezzola (5) Bolsena (1) Iseo (2) Bracciano (1)			
NOAA National Ocean Service Water Level Observation Network (USA)	St John River (3)			
Estonian University of Life Sciences (Estonia)	Vorstjarv (4)			
Environmental Protection Agency - Ireland	Corrib (2) Derg (1)			
Martin Dokulil - Austria	Neusiedl (1)			
Israel Oceanographic and Limnological Research (Israel)	Sea of Galilee (2)			
National Institute for Environmental Studies (Japan)	Kasumigaura (5)			
Universidad del Valle de Guatemala - Guatemala	Atilian (1)			
Universitá degli Studi di Perugia (Italy)	Trasimeno (1)			
University of Waikato and the Bay of Plenty Regional Council - New Zealand	Rotorua (1)			
Centre for Ecology and Hydrology - Edinburgh - UK	Lomond (1) Leven (1)			
Freie Universitat Berlin/Fondazione Edmund Mach (Germany/Italy)	Iseo (1)			
University of Latvia and Latvian Environmental Geology and Meteorology Centre - Latvia	Razna (1)			
University of Wisconsin-Madison (USA)	Mendota (1)			
NTL LTER - North Temperate Lakes Long- Term Ecological Research (USA)	Mendota (1) Trout (1)			
The Ohio State University (USA)	Douglas (1)			

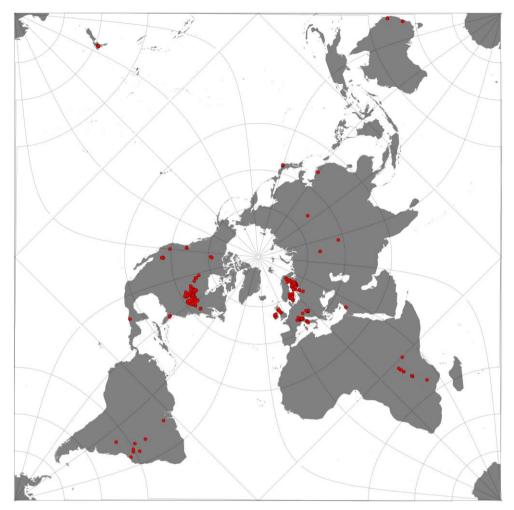


Figure 100 - Geographical distribution of sites used for LSWT validation.

Table 35. Lakes_cci lakes with in situ data. lists the 79 lakes together with their maximum distance from land [Carrea et al. 2015], which is an indication of each lakes' size that is meaningful for LSWT remote sensing. The distance to land for lake Iseo in Italy is shown in Figure 101. The best resolution of the instruments used for the retrieval of the LSWT is 1 km. If the lake has a maximum distance to land of 1.7 km such as lake Iseo, the LSWT retrieval is very likely to be available only for that part of the lake and only for a limited proportion of overpasses (clear sky and observations relatively central within the swath). In particular, a combination of factors has to occur: 1) the satellite image locations line up so that some pixels are nominally fully water pixels, which requires the satellite view zenith angle (which affects the on-the-ground resolution) to be such that the half-pixel size is smaller than the distance to coast; 2) these pixels are cloud free; 3) image geolocation errors (which can be of order 1 pixel uncertainty) are small enough that the nominally water-filled pixels are truly water-filled meaning that the water detection tests are passed.

Table 35. Lakes_cci lakes with in situ data.

Lake id	Name	Country	N sites	Max distance to land (km)
2	Superior	Canada/USA	13	73.5
3	Victoria	Tanzania	1	84.1
5	Huron	Canada/USA	9	73.3
6	Michigan	USA	15	63.8
7	Tanganyika	Tanzania	8	34.1
8	Baikal	Russia	1	33.7
11	Great Slave	Canada	2	44.6
12	Erie	Canada	10	45.6
13	Winnipeg	Canada	3	40.1
15	Ontario	Canada	5	36.1
25	Issykkul	Kyrgyzstan	1	26.9
29	Vanern	Sweden	7	20.3
44	Woods	Canada	1	11.8
52	Tucurui	Brazil	1	6.4
65	Itaipu	Paraguay	1	3.8
66	Taihu	China	1	16
67	Kivu	Zaire	1	13
95	Vattern	Sweden	2	9.9
96	Serra da Mesa	Brazil	1	4
144	Inarinjarvi	Finland	1	4.1
146	Saint Claire	Canada	1	13
157	Paijanne	Finland	3	3.8
163	Malaren	Sweden	9	2.7
165	Champlain	USA	2	5.8
188	Tres Marias	Brazil	1	2.3
195	Pielinen	Finland	4	4.1
198	Nipissing	Canada	3	9
202	Oulujarvi	Finland	1	6
236	Simcoe	Canada	1	8.4
238	Itumbiara	Brazil	1	3
278	Mead	USA	3	3.8
295	Taupo	New Zealand	3	9.6
310	Balaton	Hungary	11	6

Lake id	Name	Country	N sites	Max distance to land (km)
327	Geneva	Switzerland	1	6.2
346	Keitele	Finland	1	2.2
369	Nasijarvi	Finland	1	2.9
376	Lokan	Finland	1	4.4
380	Tahoe	USA	1	8.2
387	Hjalmaren	Sweden	1	4.7
422	Onkivesi	Finland	2	2.4
505	Garda	Italy	9	5.2
507	St John River	USA	1	2.4
530	Puulavesi	Finland	1	2.1
654	Siljan	Finland	1	5.4
657	Hoytiainen	Finland	1	3.2
679	Vorstjarv	Estonia	4	6.2
948	Maggiore	Italy	2	2.4
1028	Bolmen	Sweden	2	2.7
1057	Corrib	Ireland	2	2.6
1115	Neusiedl	Austria	1	3.6
1196	Sea of Galilee	Israel	2	5.6
1201	Vanajavesi	Finland	1	3.4
1204	Kasumigaura	Japan	5	3.7
1240	Pyhajarvi	Finland	1	3.9
1246	Lappajarvi	Finland	1	3.5
1479	Atilian	Guatemala	1	4
1519	Derg	Ireland	1	1.9
1529	Trasimeno	Italy	3	4.3
1596	Bolsena	Italy	1	5.2
1679	Ekoln	Sweden	1	1.6
1893	Roxen	Sweden	1	2.9
2054	Vuohijarvi	Finland	1	1.9
2516	Lomond	UK	1	1.5
3307	Bracciano	Italy	1	3.9
3379	Razna	Latvia	1	3.2
4503	Mendota	USA	2	2.5

Lake id	Name	Country	N sites	Max distance to land (km)
6785	Erken	Sweden	1	1.5
12262	Leven	UK	1	1.5
12471	Trout	USA	1	1.4
13377	Douglas	USA	2	1.5
300001112	Koitere	Finland	2	2.2
300001141	Mallasvesi	Finland	2	2.1
300001274	Como Mezzola	Italy	5	2
300009360	Mohave	USA	2	2.8
300011716	Myekojarvi	Finland	1	1.5
300012023	Lentua	Finand	1	2.1
300012614	Pyhajarvi	Finalnd	1	1.7
300014185	Iseo	Italy	3	1.7
300016649	Rotorua	New Zealand	2	3.1

A good portion of the lakes that have been used for the validation are small, for which, given the previous discussion, the LSWT retrieval is most challenging.

Moreover, some of the locations of in situ measurements are situated close to the coast even for large lakes, which means that the nearest water-filled pixels may not overlap the in situ measurement, thus increasing the uncertainty in the comparison from spatial representativity.

As the in situ data are from a variety of sources, with different formats, considerable effort has been put in to consolidate each new source of data to a standard format for use in validation. A quality control procedure for checking the in situ data is also necessary, since they are not always plausible. This is partly automated and partly by manual inspection. The quality control procedure was initiated within the ARCLake project and updated within GloboLakes and C3S.

The in situ data have a range of characteristics:

- the measurements have been taken at different depths up to 1 m;
- the temporal sampling of the measurements ranges from 15 minutes to few times a year;
- the temporal availability of the in situ measurements varies from few months up to covering all the satellite period;
- for some locations the measurements are averages while for others they have been taken instantaneously at the reported time;
- none of the in situ measurements which have been collected are provided with an uncertainty estimate.

While part of the data are available online, the majority has been collected through personal communications and in a proportion of cases we are not licensed to redistribute the data because of the provider's data policy.

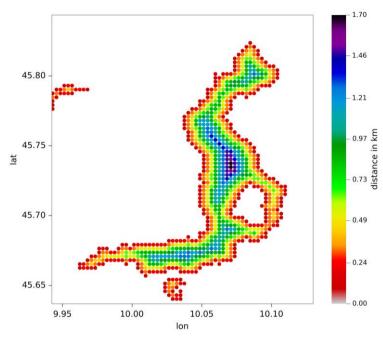


Figure 101. Distance to the nearest land for each pixel on water for lake Iseo in Italy at about 300m resolution.

4.2. Comparison methods

The validation of the Lake Surface Water Temperatures consists of the comparison with independent in situ data. The satellite-to-in situ-matches are created at the original satellite coordinates, at L2. The Lakes_cci products are not on satellite coordinates but are gridded in a regular grid at $1/120^{\circ}$ resolution and "supercollated" (combined across the available sensors) making a gridded product technically referred to as "L3S". The LSWT of the L3S grid cell of the combined and regridded temperature are therefore directly validated to assess the products as seen by users. The validation of the LSWT is performed using conventional and robust statistics, the latter being less sensitive to outliers and more descriptive of the majority of data.

4.2.1. Generation of the L2 matchups

A per-sensor matchup is created at L2 and it contains satellite and in situ data as nearly as possible co-incident and space and time. The match defines the reference temperature and time from the in situ location and the associated LSWTs, quality level and uncertainty from the L2 LSWT product. The matchup is created for satellite observations based on the following criteria:

- Spatially within 3 km from the location of the in situ measurement and
- **Temporally** within 3 hours for the in situ measurements where the measurement time was available. For some of the lakes only daily mean temperature was recorded or the time-of-day of the measurements was not reported, and in these cases the time criterion was to match the day of observation.

4.2.2. Validation of the L3S CCI LSWT v4.3

The differences between the L3S LSWT and in situ data are analysed using both standard and robust statistics. Robust statistics is less influenced by outliers in the distribution of differences, which can be also caused by poor in situ measurements. Quality control measures have been applied to in situ measurements, but many different instruments have been used to take measurements and the operating methods of the instruments and the reporting vary strongly between sites. Time series of the absolute temperatures together with their difference are generated differentiating the quality levels. "Violin" plots where the distribution of the difference is shown are produced for each quality level. The robust statistics is also investigated per quality level for each year and for each lake.

4.2.3. Validation of the LSWT uncertainty

The validation of the L3S Lakes_CCI LSWT v4.3 is carried out comparing the satellite minus in situ temperature difference with the combination of the satellite uncertainty (present in the products) and an estimate of the in situ uncertainty (which is relatively poorly known). In an ideal case, the standard deviation of the differences between the satellite LSWT and a reference LSWT would equal the combined measurement uncertainty plus the uncertainty attributable to representativity effects.

4.2.4. Number of CCI lakes with LSWT

An assessment of the lakes with no retrieved LSWT is reported. Most of the lakes are too small in comparison with the satellite resolution and other can be included in the next version.

4.3. Description of work

The matchup is carried out per sensor over the 208 locations on 79 lakes. The total number of matches is 81,436 for any quality level and 66,407 excluding satellite LSWT of quality level equal to 1. The number of matches varies per year and since the AVHRR and the MODIS sensors have a larger swath than the ATSR sensors (ATSRs swath is 500 km, AVHRRs swath is ~2900km and MODIS swath is 2330 km), after 2000 the number of matches clearly increases as it is shown in Figure 102. We can notice two other clear increase in 2007 the AVHRR on MetOpA is introduces and in 2013 when the AVHRR on MetOpA is used together with the AVHRR on MetOpB. After 2019 the number of matches decreases since the AVHRRs are only used until August 2019. The number of matches depends also on the availability of the in situ since a different number of locations is available every year as shown in Figure 103. The number of locations where in situ measurements have been taken has almost tripled since 1995; however, a portion of the measurement temporal frequency is daily.

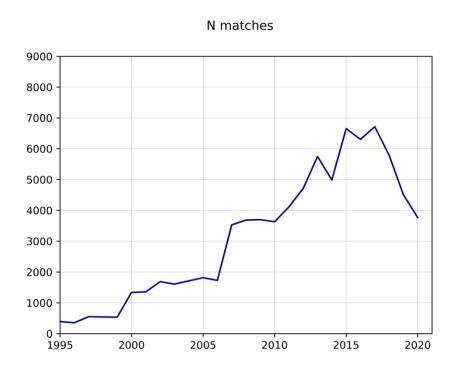


Figure 102. Number of matches for the CCI lakes at L3 per year.

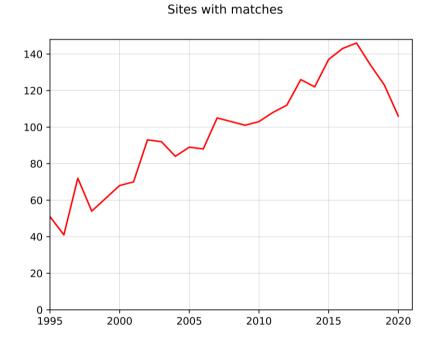


Figure 103. Number of sites with matches for the CCI lakes at L3 per year.

4.4. Result analysis

4.4.1. Validation of LSWTs

Robust statistics and the traditional statistics per quality level are reported in Table 36. Robust and traditional statistics of the LSWT satellite minus in situ difference. for the matches across all the locations where in situ measurements were available as reported in Table 35.

Table 36. Robust and traditional statistics of the LSWT satellite minus in situ difference.

QL	N	Median	RSD	Mean	SD
5	30080	-0.150	0.504	-0.127	0.936
4	17802	-0.220	0.756	-0.249	1.203
3	11767	-0.300	1.023	-0.413	1.472
2	6758	-0.590	1.460	-0.792	1.811
1	15029	-2.280	3.988	-3.474	4.874

In Table 36 the number of matches per quality levels are listed together with the median and the robust standard deviation of the satellite minus in situ temperature difference and the traditional metrics, the mean and the standard deviation. The difference between the median and the mean is less than 0.025 K for quality level 5 and it increases as the quality levels get lower (suggesting, as expected, a higher incidence of cold-biased observations for low quality levels). The agreement between satellite and in situ measurements varies according to the quality levels in a way that is expected.

The best agreement is for quality levels 4 and 5, which are the levels that reflect a higher degree of confidence in the validity of the satellite estimate. Our recommendation to users is to use the quality level 4 and 5 for lake-climate applications in general, although lower quality level data may be relevant to users where they have specifically verified their fitness for a given lake for their application. Quality level 3 data comparison with the in situ data shows an agreement that may be acceptable to some users; however, they have to be used with care. Quality level 1 data should never be used and they are classified as "bad data".

A contribution to the difference on average is the expected skin effect. Infrared radiometers are sensitive to radiation emitted between the air-surface interface and 20mm below the interface while the in situ measurements considered here are taken at a distance up to 1m from the air-surface interface. During the night, the surface of the water is generally cooler than the subsurface by ~0.2 K [Saunders, 1967], [Embury et al, 2012]. However, during the day, if the wind speed is low enough, thermal stratification due to solar heating contributes a positive offset to the difference in temperature between the radiometric lake surface and the in situ measurement depth (up to 1 m). The positive thermal stratification would be expected to be in the range <<1 K for most observations and but occasionally of order a few kelvins. The degree of near-surface stratification to be expected in different lakes depends on fetch, weather conditions (radiative balance and wind speed), the depth of in-situ measurement, and any local vertical mixing perturbations introduced by the presence of the in-situ measurement system. The aggregate effect of these factors is not currently well quantified. Overall, it

is plausible that for day time LSWT observations the mean stratification effect is of order one or a few tenths, as has been determined over the oceans. In summary, a geophysical contribution to the satellite minus in situ temperature difference is the expected skin effect of -0.2 K, but other positive geophysical offsets are similar in magnitude and difficult to quantify precisely. In this context, a mean agreement of the physics-based retrievals and validation within +/-0.2 K is a convincing result. In terms of scatter, as well as the retrieval uncertainty and variability in the vertical stratification effects, the scatter includes in situ uncertainty and horizontal variability. Again, quantitative understanding of the scatter from these effects is not yet mature, and for this reason full uncertainty budget validation remains a research aspiration (see also E3UB).

The distributions of the satellite minus in situ temperature differences per quality level 2,3,4,5 are reported in Figure 104 as "violin" plots. The distributions become more stretched and less symmetric with longer tail towards negative differences as the quality levels decreases.

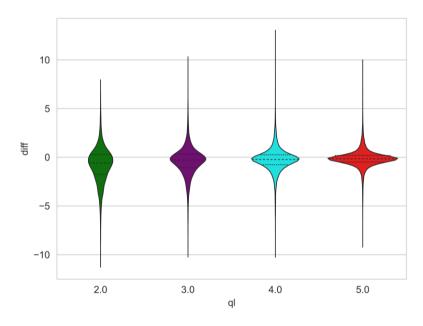


Figure 104. Distributions of the satellite minus in situ temperature difference per quality level as "violin" plots where the widths indicate the density of data for a given difference.

The median and the robust standard deviation per quality level per year for all the lakes is shown in Figure 105 and Figure 106 together with the number of matches. For high quality levels the median and the robust standard deviation of the satellite minus in situ differences are consistently small throughout the years when different instruments have been adopted and a different number of matches is available. They deteriorate as the quality goes lower. The number of matches for quality level 5 is consistently the highest.

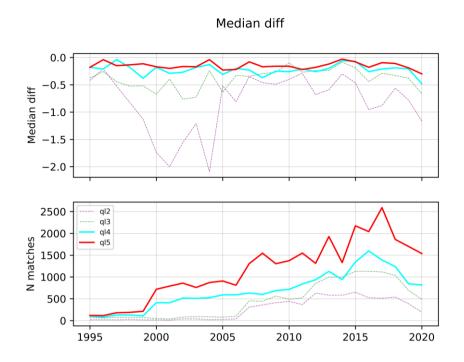


Figure 105. Satellite minus in situ temperature difference median per year (upper plot) and number of matches (lower plot) per quality level.

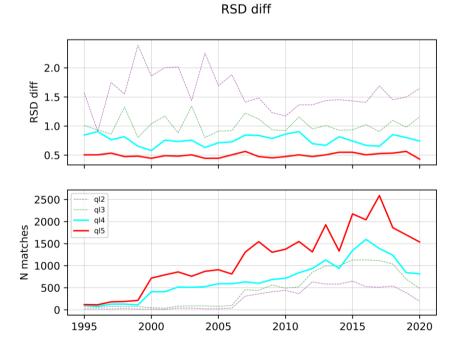


Figure 106. Satellite minus in situ temperature difference robust standard deviation per year (upper plot) and number of matches (lower plot) per quality level.

The median and robust standard deviation have been inspected also for each lake. Figure 107 and Figure 108 show the plots together with the correspondent number of matches. Higher numbers of matches are for lakes where data were available for longer periods but

also where hourly/subhourly measurements were available and for sites far from the coast. The median and robust standard deviation are consistently better for quality level 5 throughout the lakes. For both quality level 5 and 4 the instances of greater variation are related to lower numbers of matches. However, for some lakes/quality-level combinations the in situ measurements are very few: for example, for lake Taihu (lake ID 66), only one match with LSWT of quality level 4 is available. For lake Paijanne (lake ID 157), the median difference and the robust standard deviation are unusually large. The in situ data for this lake come from two different originators for three sites as shown in Figure 109. Figure 110 shows the satellite LSWTs, the in situ values and the climatology (for reference) for the three sites on lake Paijanne in 2015. The majority of the measurements have been taken at the site 1 which is very close to the coast where the satellite minus in situ difference shows cooler satellite LSWT in the first part of the year and warmer in the second part. This behaviour is consistently throughout the years and is consistent with an effect of shallow-water energy balance on the in situ measurement that differences from the satellite location of the matches that are at a spatial distance up to 3 km; for comparing to measurements close to the coast in shallow water, this is a significant distance. For the other two sites the in situ and satellite measurements have a good agreement but the in situ data have a lower temporal resolution.

Median diff per lake

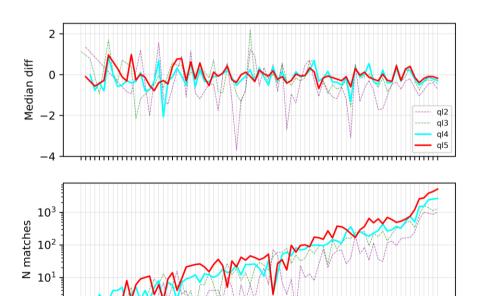


Figure 107. Satellite minus in situ temperature difference median per lake (upper plot) and number of matches (lower plot) per quality level.

10⁰

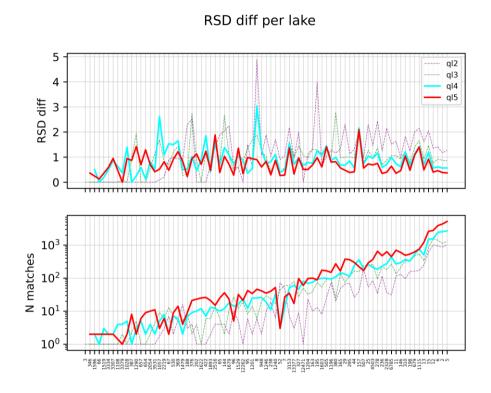


Figure 108. Satellite minus in situ temperature difference robust standard deviation per lake (upper plot) and number of matches (lower plot) per quality level.

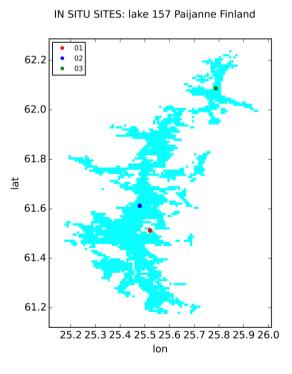


Figure 109. Locations of in situ measurements for lake Paijanne in Finland. Each dot represents a 1/120°x1/120° cell.

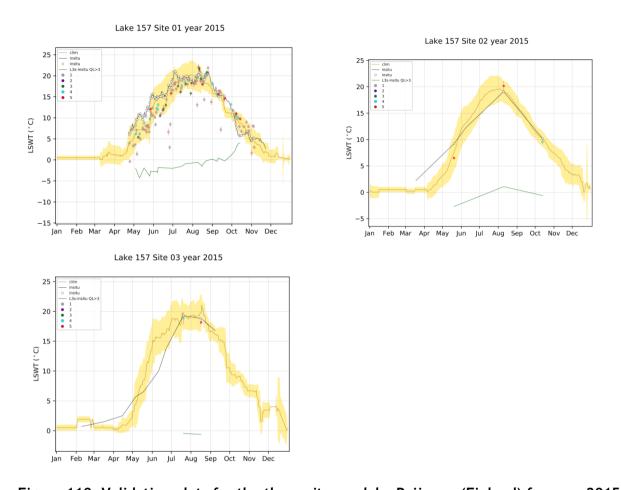


Figure 110. Validation plots for the three sites on lake Paijanne (Finland) for year 2015 where the yellow plot is the climatology, the black line and the white dots are the in situ measurements, the coloured dots are the satellite LSWTs per quality level and the green line is the satellite minus in situ difference.

As another example, consider lake Superior where many sites are available. The robust statistics of the difference for all the matches of quality level 3,4,5 have been plotted per sites in Figure 111 showing consistency for sites near each other, and a higher variability of the differences for sites close to the coast than those far into the lake, as expected. Figure 112 and Figure 113 show the LSWT and the in situ measurements in 2014 respectively for three sites on the lake Superior. In 2014 a sharp increase in temperature can be observed in the beginning of August consistently for the three sites and consistently for the satellite and in situ measurements. The timing of temperature increase is consistent at the three very different locations, one being closer to the coast and the other two more offshore. Figure 113 (right hand side) shows the position of the in situ measurement sites on lake Superior.

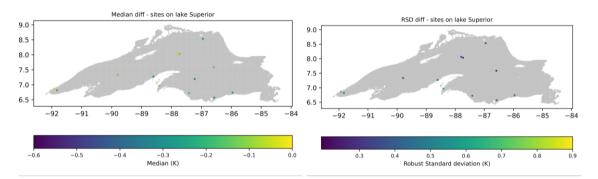


Figure 111. Satellite minus in situ temperature difference median and robust standard deviation for all the sites on lake Superior for quality level 3,4,5.

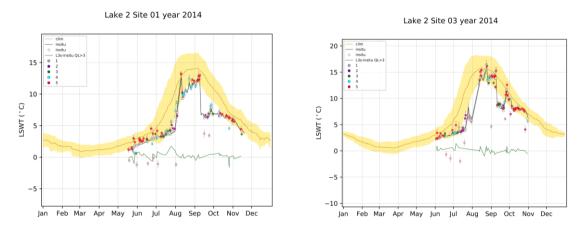


Figure 112. Satellite observations (dots), in situ matches (white dots), in situ measurements (black line), satellite minus in situ T difference for quality levels 4,5 (green line) and climatology (golden line with climatological variability as the yellow band) for lake Superior in 2014 at site 01 (on the left) and site 03 (on the right).

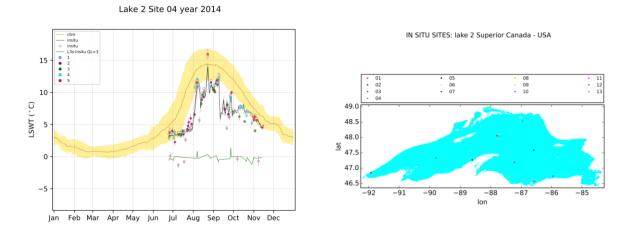


Figure 113. Satellite observations (dots), in situ matches (white dots), in situ measurements (black line), satellite minus in situ T difference for quality levels 4,5 (green line) and climatology (golden line with climatological variability as the yellow band) for lake Superior in 2014 at site 04 (on the left) and location of the in situ measurement sites on lake Superior on a 1/120° grid (on the right).

The time series of the satellite and the in situ temperature together with their difference have been inspected and they are reported here for two "difficult" validation cases as examples. The first is a small lake (lake Erken in Sweden). The second is lake Kasumigaura in Japan, a lake with low-temporal-frequency data.

The location where the in situ measurements have been collected on Lake Erken in Sweden is shown in Figure 114 (red dot). Figure 115 shows the satellite observations and the in situ measurements in 1995 when only ATSR2 was utilised and in 2014 when observations from AVHRR-A, AVHRR-B and MODIS were used. For both the years the satellite observations follow remarkably well the in situ measurements, which were very high frequency measurements. The peak in the difference (green line) in 1995 is very likely due to a slight temporal mismatch. Despite the peculiar behaviour of the temperatures through the year in both cases with LSWT being well beyond one stardard deviation from the climatological reference, the satellite and in situ are mimicking each other remarkably well. Note that in this case the measurements site is close to the shore but matching within 3 km does not have a strong influence in this case because the lake is small and more consistent in LSWT across its area.

IN SITU SITES: lake 6785 Erken Sweden

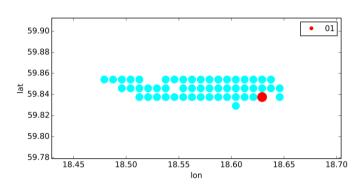


Figure 114. Location of in situ measurements for lake Erken in Sweden. Each dot represents a 1/120°x1/120° cell.

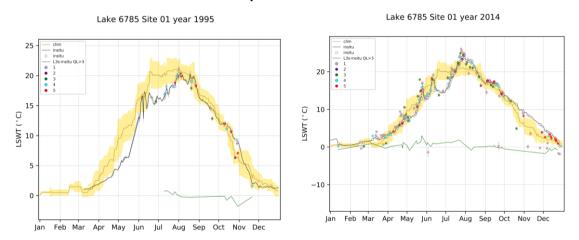


Figure 115. Satellite observations (dots), in situ matches (white dots), in situ measurements (black line), satellite minus in situ T difference for quality levels 4,5 (green line) and climatology (golden line with climatological variability as the yellow band) for lake Erken in Sweden in 1995 (on the left) and 2014 (on the right).

Figure 116, Figure 117, Figure 118 present lake Kasumigaura in Japan at the four sites in year 2016 and Figure 119 presents the site 05 in 2016 and 2020. The sites 01 and 03 are very close to the coast, so the matches are lower in number than for the other sites. However, a consistency between in situ and satellite LSWT can be observed for all the sites.

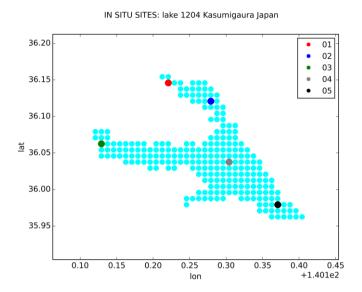


Figure 116. Location of in situ measurements for lake Kasumigaura in Japan. Each dot represents a 1/120°x1/120° cell.

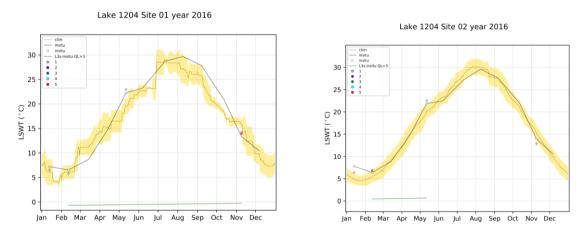


Figure 117. Satellite observations (dots), in situ matches (white dots), in situ measurements (black line), satellite minus in situ T difference for quality levels 4,5 (green line) and climatology (golden line with climatological variability as the yellow band) for lake Kasumigaura in Japan in 2016 for site 01 (on the left), site 02 (on the right).

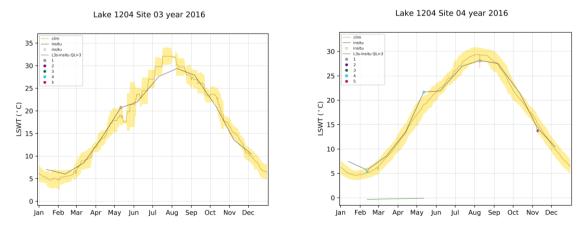
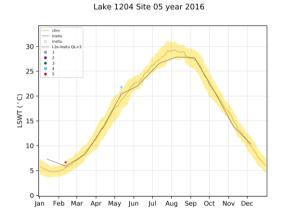


Figure 118. Satellite observations (dots), in situ matches (white dots), in situ measurements (black line), satellite minus in situ T difference for quality levels 4,5 (green line) and climatology (golden line with climatological variability as the yellow band) for lake Kasumigaura in Japan in 2016 for site 03 (on the left), site 04 (on the right).



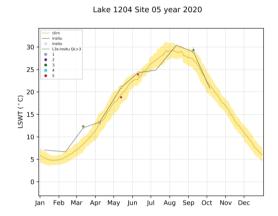


Figure 119. Satellite observations (dots), in situ matches (white dots), in situ measurements (black line), satellite minus in situ T difference for quality levels 4,5 (green line) and climatology (golden line with climatological variability as the yellow band) for lake Kasumigaura in Japan in 2016 and 2020 for site 05.

4.4.2. Validation of the uncertainty LSWT v4.3

The LSWT uncertainty estimate has been validated comparing the difference satellite minus in situ temperatures and the correspondent LSWT and in situ uncertainties. The following quantity is calculated for each match:

$$\Delta = \frac{T_{LSWT} - T_{INSITU}}{\sqrt{\sigma_{LSWT}^2 + \sigma_{INSITU}^2 + \sigma_{repr}^2}}$$

where T indicates temperature, for LSWT and in situ as indicated in the subscripts. σ means the standard deviation from measurement uncertainty (for LSWT and in situ) and from real differences because of point-to-pixel representativity effects.

The in situ measurements uncertainty is not known for the data we have. We explore two assumptions: $\sigma_{\text{INSITU}}=0.2\,$ K, a value based on deployment of similar measurement technologies to the ocean, and $\sigma_{\text{INSITU}}=0.5\,$ K which would be at the upper end of our expectations for in situ uncertainty. The representativity effect is presently unquantified and we set it to 0 K for the present; neglecting representativity has the tendency to make the LSWT uncertainty look underestimated.

Lakes_cci products, σ^2_{LSWT} is context sensitive and varies from match to match, which is why the validation approach involves the calculation of the above metric: the distribution of Δ should be a Gaussian distribution with mean equal to 0 and standard deviation equal to 1 when all standard deviations are well estimated and the retrieval is unbiased relative to the in situ and any mean geophysical effect. Figure 120 shows the histograms of the

uncertainties per quality level where also the fitted Gaussian and the target Gaussian distributions are shown. In Figure 120 an in situ uncertainty of 0.2 K has been assumed.

N(0,1) N(-0.31,2.17) QL=5 N(0,1) N(-0.62,3.02) QL=4 0.3 ⊆ □ 0.2 0.0 0.4 0.3 c 0.2 \subseteq 0.1 0.0 -10 _ _5 Ó 10 Ó -10

LSWT uncertainty validation, insitu unc = 0.2

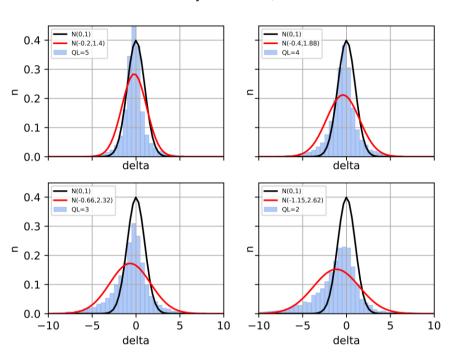
Figure 120. LSWT uncertainty validation (in situ uncertainty = 0.2K) per quality level (indicated in legend): histograms of D

delta

delta

For quality level 5, the Gaussian fit has width 2.17, which means that observed differences are more different than expected from the quoted uncertainties. This may be partly because the product uncertainties are underestimated but could also arise to the degree that lake in situ data (being more diverse) have larger uncertainty than the assumed value (based on experience of ocean observations), and because representativity is neglected. Interpretation of this outcome is therefore currently ambiguous, and research is needed to better understand the in situ uncertainty and representativity effects.

We used an in situ uncertainty of 0.5 K to explore the level of in situ uncertainty that would better fit the Gaussian. Figure 121 shows the uncertainty validation for this value. The width of the Gaussian fit for quality level 5 is much closer to one, and thus 0.5 K may be closer to the combined effect of in situ and representativity uncertainty. However, more investigations of in situ uncertainty and representativity need to be carried out to understand this better.



LSWT uncertainty validation, insitu unc = 0.5

Figure 121. LSWT uncertainty validation (in situ uncertainty = 0.5K) per quality level (indicated in legend): histograms of D

4.4.3. Lakes with no LSWT

For 14 target lakes out of the 2025 lakes of the CCI Lake v2 list, LSWT has been obtained with less than 100 days of observations of quality level>1, largely due to the fact that they are too small. The lakes are listed in Table 37 together with the estimated maximum distance to land. The majority of the lakes is not feasible because of their small size, except for lakes 200000071 and 20000072 where in the mask used during the processing they were labelled as sea rather than lakes.

CCI Lake id	Name	Country	Max distance to land (km)	N days of observation	Comments
1	Caspian	Azerbaijan/Russ ia/Turkmenista n/Iran	164.8	0	Labelled as sea
18089	Macnean	United Kingdom	0.6	0	Too small
164651	Portmore	United Kingdom	0.5	3	Too small
208840	Mantua	Italy	0.5	0	Too small
215215	Morse	USA	0.5	0	Too small

Table 37. Lakes with no LSWT

CCI Lake id	Name	Country	Max distance to land (km)	N days of observation	Comments
215311	Geist	USA	0.5	0	Too small
215339	Eagle Creek	USA	0.6	0	Too small
10000004	Mourve	United Kingdom	0.2	0	Too small
20000071	Patos Lagoon	Brazil	25.2	0	Labelled as sea
200000072	Maracaibo	Venezuela	46.5	0	Labelled as sea
300134644	Rihpojavri	Norway	0.7	0	Too small
16662	Jijila	Romania	1.3	3	Too small
300016021	Ambussel	Tanzania	1.8	19	Too small
799	Hawizeh marshes	Iraq/Iran	1	22	Too small/dry

4.5. Conclusions and recommendations

The validation of the LSWT shows very good mean agreement (comfortably within +/-0.2 K) between satellite LSWT and independent in situ temperature measurements.

The uncertainty validation principles are clear, but the results are not mature enough to make strong validation statements regarding the uncertainty information since in situ uncertainty and representativity uncertainty are poorly known for lakes: experience from the ocean suggests in situ/representativity uncertainty of order 0.2 K, but for lakes this may be an underestimate. An initial analysis suggests 0.5 K is closer to the case.

The LSWT and uncertainty validation show that the quality level accompanying the LSWT are a very important variable for the proper use of the data. Quality levels provide the confidence on the LSWT retrieval. We recommend to use quality levels 4 and 5 for lake-climate applications.

Manual inspection of all products for more than 2000 water bodies is impossible and in most cases requires local knowledge. The validation of the products is, and always will be, based on a small sample of well-studied areas. Users of these products are therefore advised to inspect the results for their area of interest before generating derivative products and any feedback to Lakes_cci would be most useful.

4.6. Acknowledgement

We would like to thank lestyn R. Woolway for having established the contacts to set up the in situ database and all the institutions listed in Table 2 that have provided us with in-situ data and in particular: Enner Alcantara (Saõ Paulo State University, Saõ Paulo, Brazil), Gil Bohrer (The Ohio State University, Columbus, USA), Jean-Francois Cretaux (LEGOS, Toulouse, France), Margaret Dix (Universidad del Valle de Guatemala, Guatemala), Hilary Dugan (Center for Limnology, University of Wisconsin-Madison, USA), Martin Dokulil (Mondsee, Austria), Gideon Gal (Yigal Allon Kinneret Limnological Laboratory, Israel

Oceanographic and Limnological Research, Migdal, Israel), Claudia Giardino (Istituto per il Rilayemento Elettromagnetico dell'Ambiente, National Research Council of Italy, Italy). April James (Nipissing University, Canada), Ilga Kokorite (University of Latvia and Latvian Environmental Geology and Meteorology Centre, Latvia), Johanna Korhonen (SYKE, Helsinki, Finland), Ben Kraemer (Leibniz Institute for Freshwater Ecology and Inland Fisheries, Berlin, Germany), Alo Laas (Estonian University of Life Sciences, Tartu, Estonia), Eric Leibensperger (Center for Earth and Environmental Science, SUNY Plattsburgh, USA), Junsheng Li (Institute of Remote Sensing and Digital Earth, Chinese Academy of Science, China), Alessandro Ludovisi (Dipartimento di Biologia Cellulare e Ambientale, Universita degli Studi di Perugia, Italy), Chris MacBride (University of Waikato and the Bay of Plenty Regional Council. New Zealand). Shin-ichiro Matsuzaki (National Institute for Environmental Studies, Japan), Linda May (Centre for Ecology and Hydrology, Edinburgh, UK) , Ghislaine Monet (UMR CARRTEL, Thonon le Bains, France), Tiina Nogesand Peeter Noges (Estonian University of Life Sciences, Tartu, Estonia), Sajid Pareeth (Freie Universitat Berlin/Fondazione Edmund Mach, Germany/Italy), Sebastiano Piccolroaz (Institute for Marine and Atmospheric Research, Department of Physics, Utrecht University), Don Pierson (Uppsala University, Sweden), Pierre-Denis Plisnier, Merja Pulkkanen (SYKE, Helsinki, Finland), Antti Raike (SYKE, Helsinki, Finland), Alon Rimmer (Yigal Allon Kinneret Limnological Laboratory, Israel Oceanographic and Limnological Research, Migdal, Israel), Michela Rogora (CNR Institute for Water Research (CNR IRSA), Italy), Geoffrey Schladow (UC-Davis Tahoe Environmental Research Center, USA), Eugene Silow (Irkutsk State University, Russia), Lewis Sitoki (Department of Earth Environmental Science and Technology, Technical University of Kenya, Nairobi), Evangelos Spyrakos (Biological and Environmental Science, University of Stirling, Scotland, UK), Wim Thiery (Department of Earth and Environmental Sciences, KU Leuven, Belgium), Piet Verburg (NIWA, New Zealand), Gesa Weyhenmeyer (Department of Ecology and Genetics, Uppsala University, Sweden), Caroline Wynne (Environmental Protection Agency, Ireland).

5. Lake Ice Cover (LIC)

Lake ice cover (LIC) refers to the extent (or area) of a lake covered by ice. In Lakes_cci, LIC is a daily merged level 3 (L3) product generated from MODIS Terra/Aqua Calibrated Radiance 5-Min L1B Swath (MOD02/MYD02), Collection 6.1, data (see ATBD for details). In the LIC product, each pixel is assigned one of four possible class labels: water (value=1), ice (value=2), cloud (value=3), and bad (value=4; case where a retrieval was not possible due to poor data quality). Only pixels labelled as water, ice or cloud are considered in product validation. It is important to note that validation is performed on Terra and Aqua derived LIC individually, before merging into the daily Terra/Aqua L3 product and before aggregating into the ca. 1 km harmonized grid.

Product validation can be performed following three approaches: (1) comparison against ground-based nearshore observations; (2) comparison with LIC products generated by other algorithms/groups such as NASA's MODIS Snow Cover products from Terra and Aqua (MOD10/MYD10); and (3) validation against groups of pixels (Areas Of Interest or AOI) extracted for a selection of lakes from visual interpretation of original MODIS Terra and Aqua imagery used as input into the LIC retrieval algorithm (i.e. MOD02/MYD02). Here, validation involved a mix of approaches (2) and (3) whereby the thematic accuracy (water, ice and cloud) of Lakes_cci LIC (Aqua and Terra severally) was evaluated, and results of Lakes_cci LIC and NASA's MODIS Snow Cover products were compared by visual inspection. Approach (1) was not considered herein since ground-based nearshore observations were unavailable or non-existent at the time of writing of this report; however, most importantly, such observations are of a more limited value than the two other approaches for lake-wide validation of the LIC product.

5.1 Data description

5.1.1. MODIS Terra/Aqua Calibrated Radiances Level 1B product (MOD02/MYD02)

MODIS Terra/Aqua Calibrated Radiances L1B 5-Min Swath, Collection 6.1 (MODIS Characterization Support Team, 2017) - MOD02/MYD02 false color composites (R: band 2, G: band 2, B: band 1) with a 250 m spatial resolution were used as reference images to manually collect AOIs with assigned labels (lake ice, open water, and cloud) to assess the accuracy of the Lakes_cci LIC. MODIS band 1 is centered at 645nm (red) and band 2 at 865 nm (near-IR). In addition to false color composites, true color composites were also used for visual inspection (R: band 1, G: band 4, B: band 3). MODIS band 3 is centered at 469nm (blue) and band 4 at 555nm (green) with 500m.

5.1.2. MODIS Snow Cover product (MOD10/MYD10)

MODIS Terra/Aqua Snow Cover 5-Min L2 Swath 500 m, Collection 6.1 (C6.1), daily products (MOD10/MYD10) were also validated and compared with the Lakes_cci LIC product. In MOD10/MYD10, lake ice cover is identified using a similar criteria as snow over land (i.e. Normalized Difference Snow Index (NDSI) ≥ 10 and data screens) and a land/water mask provided in MOD/MYD03 products (Riggs et al., 2019). The NDSI is derived from MODIS radiance data acquired by Terra/Aqua satellites (i.e. MODIS Level 1B product (MOD02/MYD02) top-of-the-atmosphere (TOA) calibrated radiance/reflectance data). MODIS Cloud Mask products (MOD/MYD35_L2) have been also applied to filter cloud cover

for MOD10/MYD10 products. Input data to the Snowmap algorithm used in the production of MOD10/MYD10 are shown in Table 38. Therefore, four labels (i.e., ice, water, cloud, other) are presented in MOD10/MYD10 snow maps to indicate ice cover over lake areas.

Table 38. MODIS data inputs for the collection 6.1 (C6.1) snow algorithm (adapted from Riggs et al., 2019)

Input Data	Product Description
MOD/MYD02HKM	Calibrated Radiances (500 m)
MOD/MYD021KM	Calibrated Radiances (1 km)
MOD/MYD03	Geolocation Fields (1 km)
MOD/MYD35_L2	Cloud Mask and Spectral Test Results (250 m and 1 km)

^{*} Depends on platform, MOD indicates Terra, MYD indicates Agua

The MODIS Snow Cover product experiences issues in differentiating between clouds and snow/ice, and both errors of commission and omission can be found in the product. Omission occurs when the MODIS cloud mask (MOD/MYD 35) identifies an area of snow/ice as certain cloud, therefore excluding it from the MODIS snow product (Riggs et al., 2019). Errors of commission can occur on the periphery of cloud cover or where clouds appear similar to snow/ice and are excluded from the cloud mask and instead identified as snow/ice (Riggs et al., 2019). Cloud cover can be removed from the product by extracting the classification value from the previous and proceeding pixel to determine the possible classification of a cloud covered pixel. High solar zenith angle can also prevent accurate classification of snow/ice cover, which is important when studying freeze-up at northern high latitudes. Pixels are screened using a solar zenith angle mask of > 70° and pixels with a solar zenith angle $\geq 85^{\circ}$ are classified as night.

5.2 Comparison methods

Validation of the Lakes_cci LIC product and NASA MODIS Snow Cover product (also used for comparison) has been performed through computation of confusion matrices built on independent statistical validation. The reference data for validation were collected for water, ice and cloud as AOIs from the visual interpretation of the false colour composite surface reflectance images (MOD09/MYD09), as described in section 5.1.1, over three ice seasons (freeze-up and break-up periods) interspersed across the 20-year MODIS record by skilled ice analysts.

In addition to statistical validation, comparison of the Lake_cci LIC product and NASA MODIS Snow Cover product was performed through visual inspection. The reference data include false colour composite and true colour composite images.

5.3 Description of work

Work on product validation and intercomparison is comprised of three parts. The first part includes algorithm development and validation for the Lakes_cci LIC product derived from MODIS Terra. Validation for the Lakes_cci LIC product generated from MODIS Aqua is

presented in the second part. The last part shows intercomparison of the Lakes_cci LIC and NASA Snow cover products by visual inspection.

In the first part, a total of 17 lakes have been selected across the Northern Hemisphere to serve for the purpose of both development and validation of the Lakes cci LIC product derived from MODIS Terra (Figure 122). Samples were collected for three ice seasons (2002-2003, 2009-2010, 2016-2017) as to include MODIS data from Terra (2000-present) and Agua (2002-present), and provide a good temporal spread over the full record to ensure algorithm stability. For each lake, one image from the freeze-up period and one image from the break-up period were selected for both Terra and Agua images if available. False color composites (R: Band 2, G: Band 2, B: Band 1) with a 250 m spatial resolution were used as reference images to manually extract AOIs with labels (ice, water, and cloud). In total 54 images (20 from FU, 34 from BU) were selected and 276,003 pixels were sampled from the selected AOIs. A 100-fold cross-validation was used to conduct model comparison and examine model transferability. In order to evaluate the performance of the algorithm for the freeze-up and break-up periods separately, additional samples from MODIS Terra were collected for a full ice season (2018 - 2019) from Great Slave Lake and Lake Ladoga. A total of 10,075,081 pixels taken from 229 swaths over the two lakes were used to present classification performance by confusion matrices.

The second part shows the validation of the Lake_cci LIC product generated by MODIS Aqua. Similar to the work for the Lake_cci LIC product from MODIS Terra, 1,665,188 samples were collected based on false colour composite images produced from the MYDO2 Aqua product over Great Slave Lake (156 swaths), Lake Ladoga and Lake Onega (61 swaths) in 2020. A confusion matrix computed by using the samples is provided to present the classification performance.

In the last part, the Lakes_cci LIC product is compared with the NASA MODIS Snow Cover product through visual inspection. Several classification challenges for the two products, such as thin ice detection, thin cloud/fog retrieval, and confusion between ice and cloud are examined.

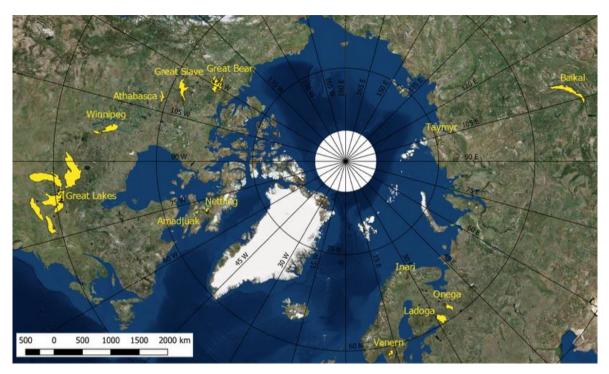


Figure 122. Geographical distribution of lakes used for Lakes_cci LIC algorithm development and validation

5.4 Result analysis

Figure 123 shows the accuracies computed from a 100-fold cross-validation (CV) using the samples of MODIS Terra from the 17 lakes. Random forest (RF) was found to outperform two other machine learning algorithms (multinomial logistic regression, MLR, and support vector machine, SVM) and comparable to gradient boosting trees (GBT) for lake ice cover, open water and cloud classification in a recent paper by the developers of the current LIC product (Wu et al. 2021). Furthermore, RF provided consistent results based on a comprehensive accuracy assessment (random k-fold as well as spatial and temporal CV as shown in Table 39).

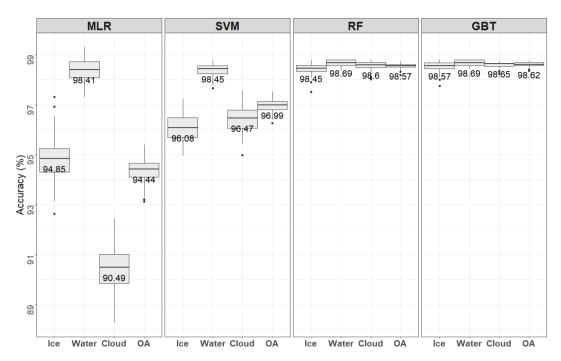


Figure 123. Comparison of accuracies (%) obtained using random 100-fold CV across classifiers for the ice, water and cloud classes individually, and overall (OA) (Wu et al., 2021)

Table 39. Accuracy assessment using temporal and spatial CV methods (adapted from Wu et al., 2021)

	MLR	SVM	RF	GBT
Temporal CV accuracy	93.21%	83.00%	95.49%	95.15%
Spatial CV accuracy	90.98%	79.36%	95.64%	95.26%

Table 40 provides the accuracies reached by the Lakes_cci LIC product with input from MODIS Terra for the break-up (BU) and freeze-up (FU) periods taken individually. There is no notable difference in the accuracy of the Lakes_cci LIC product between the BU (OA: 98.14%) and FU (OA: 96.83%) periods, and the classification accuracies are consistent across classes. Table 41 presents classification results for MODIS Aqua Lake_cci LIC, which are comparable to the performance of MODIS Terra Lake_cci LIC shown in Table 40.

Table 40. Confusion matrices with retrieval accuracies for Lakes_cci LIC derived from MODIS Terra (break-up and freeze-up periods individually) (adapted from Wu et al., 2021)

Freeze-Up				Break-Up					
		RF (classification)			RF (classification)				
Lakes Terra	_cci LIC	Ice	Water	Cloud	Accuracy	Ice	Water	Cloud	Accuracy
d	lce	740,105	3,305	15,520	97.52%	774,412	1,213	14,537	98.01%
User- defined	Water	613	509,797	4,863	98.94%	927	776,296	4,350	99.32%
) #	Cloud	154,578	15,790	4,700,459	96.50%	45,418	6,475	2,306,423	97.80%
Overall Accuracy: 96.83%			Overall A	ccuracy: 9	8.14%				

Table 41. Confusion matrices with retrieval accuracies for Lakes_cci LIC derived from MODIS Aqua

Lakes_cci LIC Aqua		RF (classification)				
		Ice	Water	Cloud	Accuracy	
, p	Ice	246,032	2,487	4,751	97.14%	
User- defined	Water	1,743	538,668	4,655	98.83%	
) ap	Cloud	19,738	5,324	841,790	97.11%	
Overall Accuracy: 97.68%						

In NASA's MOD10/MYD10 products, lake ice cover is estimated as snow cover with the scenario where NDSI is greater than 10 (Riggs et al., 2019). Besides NDSI, low visible reflectance screen thresholds of band $2 \le 0.10$ or band $4 \le 0.11$ are applied by the MODIS Snowmap algorithm in inland water bodies to detect lake ice cover (Riggs et al., 2019). Therefore, congelation ice with low visible reflectance is misclassified as open water in MOD10/MYD10 as shown in Figure 124, Figure 125, Figure 126, and Figure 127. Congelation ice usually occurs in the early ice formation or late melting period. Therefore, the performance of congelation ice detection affects the quality of lake ice phenology estimation in MOD10/MYD10. The Lakes_cci LIC product presents more accurate classification results of lake ice cover in the four cases below.

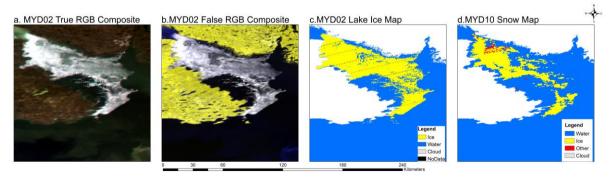


Figure 124. MODIS true RGB colour composite (a), MODIS false RGB colour composite (b), MODIS-derived Lakes_cci LIC (c), and MODIS Snow Cover product (d) for Great Slave Lake (Canada) on 19 June 2020 (UTC 19:25) by Aqua

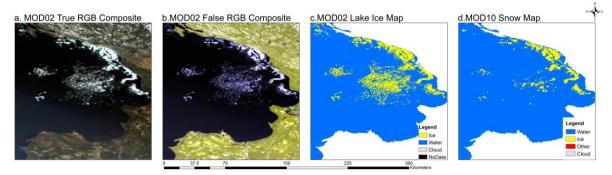


Figure 125. MODIS true RGB colour composite (a), MODIS false RGB colour composite (b), MODIS-derived Lakes_cci LIC (c), and MODIS Snow Cover product (d) for Lake Ladoga (Russia) on 26 April 2019 (UTC 09:45) by Terra

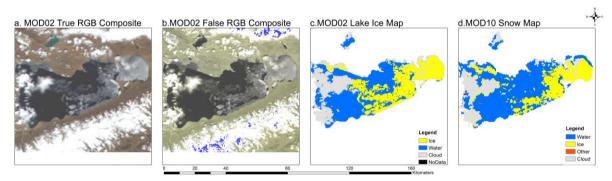


Figure 126. MODIS true RGB colour composite (a), MODIS false RGB colour composite (b), MODIS-derived Lakes_cci LIC (c), and MODIS Snow Cover product (d) for Namtso Lake (China) on 9 May 2020 (UTC 04:40) by Terra

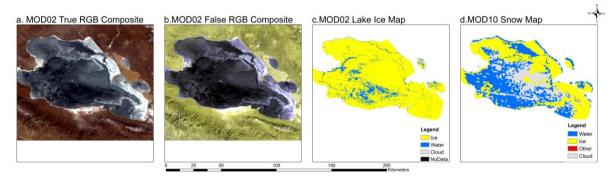


Figure 127. MODIS true RGB colour composite (a), MODIS false RGB colour composite (b), MODIS-derived Lakes_cci LIC (c), and MODIS Snow Cover product (d) for Qinghai Lake (China) on 4 January 2020 (UTC 04:25) by Terra

As mentioned in section 5.1.2, the MODIS Cloud Cover product (MOD/MYD35_L2) is used in the generation of MOD10/MYD10. Hence, the omission and commission errors of the MODIS Cloud Cover product lead to misclassification in MOD10/MYD10. Figures Figure 128 and Figure 129 show two cases of misclassification of thin ice in the MODIS snow cover product due to cloud overestimation by MOD/MYD35_L2. In contrast, the Lakes_cci LIC product can detect the lake ice cover in the two cases correctly. Moreover, as Figure 130 illustrates, the commission errors of cloud cover in MOD10 also occur under high solar zenith conditions, whereas the Lakes_cci LIC product performs better in the classification of lake ice cover.

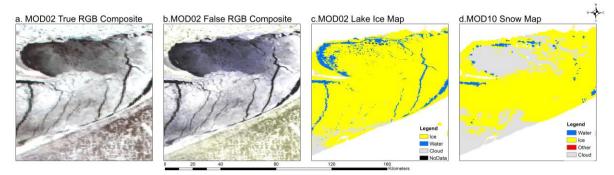


Figure 128. MODIS true RGB colour composite (a), MODIS false RGB colour composite (b), MODIS-derived Lakes_cci LIC (c), and MODIS Snow Cover product (d) for Lake Erie (Canada) on 2 February 2004 (UTC 15:50) by Terra

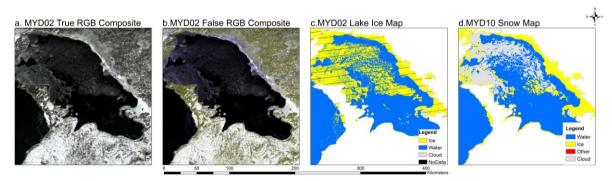


Figure 129. MODIS true RGB colour composite (a), MODIS false RGB colour composite (b), MODIS-derived Lakes_cci LIC (c), and MODIS Snow Cover product (d) for Lake Huron (Canada) on 10 February 2005 (UTC 18:30) by Aqua

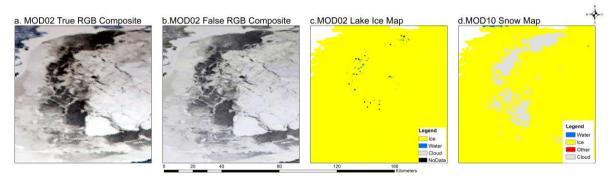


Figure 130. MODIS true RGB colour composite (a), MODIS false RGB colour composite (b), MODIS-derived Lakes_cci LIC (c), and MODIS Snow Cover product (d) for Great Slave Lake (Canada) on 30 December 2018 (UTC 19:20) by Terra under high solar zenith angle conditions (higher than 80 degrees)

The Lakes_cci LIC product also captures better the spatial distribution of cloud cover compared to MYD10/MOD10 (Figures Figure 131 and Figure 132). MOD10/MYD10 Snow Cover products underestimate cloud cover on two dates for the lakes shown as examples. With omission errors of cloud detection occurring, MYD10/MOD10 misclassify the thin cloud pixels to ice/snow cover because of high NDSI values as Figure 131 shows. Furthermore, MOD/MYD35_L2 can mislabel low-level clouds and/or fog as "clear sky", resulting in

classification errors of cloud to ice in MOD10/MYD10 as shown in Figure 132. However, Lakes_cci_LIC can tackle the two classification challenges properly.

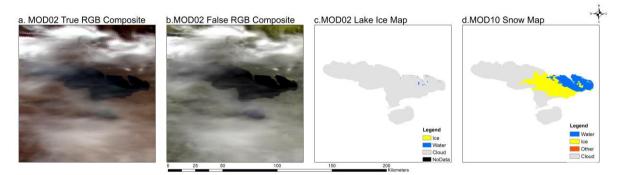


Figure 131. MODIS true RGB colour composite (a), MODIS false RGB colour composite (b), MODIS-derived Lakes_cci LIC (c), and MODIS Snow Cover product (d) for Khyargas Lake (Mongolia) on 13 May 2020 (UTC 04:10) by Terra

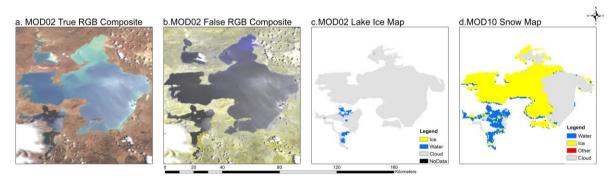


Figure 132. MODIS true RGB colour composite (a), MODIS false RGB colour composite (b), MODIS-derived Lakes_cci LIC (c), and MODIS Snow Cover product (d) for Siling Lake (China) on 1 July 2020 (UTC 04:50) by Terra

The basic assumption where a water body is deep and clear, thus absorbing solar radiation incident upon it, is employed by the Snowmap algorithm for surface classification over inland water bodies for MOD10/MYD10 (Riggs et al., 2019). Therefore, water with high turbidity or algal blooms can be incorrectly classified as ice cover. As Figure 133 shows, the Lakes_cci LIC product retrieves the water area that is misclassified as ice by MOD10.

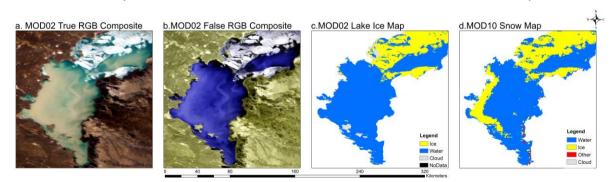


Figure 133. MODIS true RGB colour composite (a), MODIS false RGB colour composite (b), MODIS-derived Lakes_cci LIC (c), and MODIS Snow Cover product (d) for Lake Balkhash (Kazakhstan) on 25 March 2020 (UTC 06:45) by Terra

Another source of error for the MODIS Snow Cover (MOD10/MYD10) product is the geolocation of some lakes. MOD10/MYD10 employs the MODIS Geolocation L1A product in order to mask land and water (Riggs et al., 2019). However, as shown in Figure 134, the MODIS Snowmap algorithm classifies the majority of Lake Nettilling (Canada) as ocean resulting in no classification for the ice cover on the lake. However, the more accurate lake area mask used for the generation of the Lakes_cci LIC product provides correct classification of the ice conditions on that day.

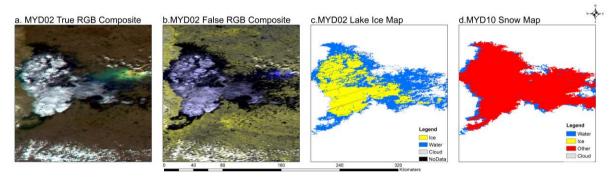


Figure 134. MODIS true RGB colour composite (a), MODIS false RGB colour composite (b), MODIS-derived Lakes_cci LIC (c), and MODIS Snow Cover product (d) for Lake Nettilling (Canada) on 25 July 2020 (UTC 15:45). The red colour corresponds to "ocean" in this MODIS Snow Cover product

5.5 Conclusions and recommendations

For Lakes_cci, LIC CDRP V2.0 is generated from a RF algorithm using MODIS Terra/Aqua Calibrated Radiance 5-Min L1B Swath (MOD02/MYD02), Collection 6.1, as primary data input. Lakes_cci LIC is a gridded product where cells are assigned one of four possible labels: water, ice, or bad (case where a retrieval was not possible due to poor data quality).

Validation of the Lakes_cci LIC product was performed via computation of confusion matrices built on independent statistical validation. Reference data for validation of class labels (water, ice, cloud) were collected from three dataset of samples collected through visual interpretation of the false color composite surface reflectance images (MOD02/MYD02) by skilled ice analysts. The reference data also served to validate lake ice/water/cloud retrievals contained in NASA's MODIS Terra/Aqua Snow Cover products (MOD10/MYD10), which were then compared to those from Lakes_cci LIC.

Results show that the retrieval algorithm (RF) implemented for Lakes_cci LIC production provides robust classification of lake ice cover. Retrieval accuracies are also found to be more consistent between classes and also ice periods for the Lakes_cci algorithm. Moreover, RF also produced comparable classification results from MODIS Terra and Aqua. Individual class accuracies are all above 90% (errors less than 10%) for LIC CDRP V2.0 which meet uncertainty requirements of 10% set by GCOS for LIC ECV (see E3UB document for details on determination of errors and uncertainty in LIC product). Additionally, according to visual inspection, compared to NASA's MODIS snow cover product (MYD10/MOD10), the Lakes_cci LIC product performs more accurately in several difficult classification cases illustrated in section 5.4.

Further assessment of the Lakes_cci LIC product and its comparison with ice products other than NASA's MOD10/MYD10 is planned in the future. This includes comparison with: 1) nearshore in situ ice/open water observations if such observations are available for any of the 2,024 lakes forming ice cover in LIC CDRP V2.0 release; 2) NOAA's Interactive Multisensor Snow and Ice Mapping System (IMS) daily 1 km global product (available since December 2014); 3) daily ice charts of the Great Lakes produce by the North American Ice Service; and 4) weekly ice fraction product from the Canadian Ice Service. Ice cover observations from these various sources will also be valuable for further assessing the quality of LIC CDRP V2.0 following its release.

6. Lake Water-Leaving Reflectance (LWLR)

Lake Water-Leaving Reflectance (LWLR), also referred to as water colour, is the measurement of the quantity of sunlight reaching the remote detector after interaction with the water column. The validation and comparison of the LWLR products, including LWLR, chlorophyll-a (Chl-a) and total suspended matter (either as TSM or expressed as Turbidity), is based on matchups between in situ and satellite measurements.

Lake Water-Leaving Reflectance (LWLR) is the result of atmospheric correction of top-of-atmosphere radiance over water pixels. This correction is the result of model optimization and subject to the possibility of ambiguous solutions. The main effects that introduce uncertainty are mixing of reflectance from water and nearby land in the atmosphere, bottom effects, in-water bio-optical model ambiguities and limited sensor band configurations to bound the mentioned numerical optimisation. A lack of in situ reference data and a bias favouring turbid, productive and large lakes in the data sets that do exist, further hampers uncertainty characterisation over a range of water types and lake geophysical and geospatial characteristics.

The problem of lacking in situ data is somewhat overcome when the end-to-end validation of biogeochemical products derived from LWLR is concerned. Here, a spatio-temporal sampling bias still exists with most of the available in situ data having been collected since the launch of MERIS. Systematic error in the LWLR retrieval can be compensated in algorithm calibration.

In this update, to fill the gap from April 2012 to April 2016 between MERIS and OLCI, independent validation and comparison of LWLR and water quality products was performed for the MODIS candidate products. Subsequently, for selected large lakes (the largest 250 lakes in CRDPv2.0), initial inter-sensor comparisons between MERIS/OLCI and MODIS for overlapping periods of three years were conducted. Based on the inter-sensor comparison results, lakes that passed consistency checks between sensors were selected for MODIS products to fill the gap.

6.1. Data description

6.1.1. In situ data

The validation dataset used in this study comprised 17 individual datasets from lakes and inland water bodies across the globe requested from the LIMNADES repository. This combined data set consisted of 1982 individual observations of remote sensing reflectance (Rrs, sr $^{-1}$), 28726 for Chl-a (mg/m 3) and 6955 of total suspended matter (TSM, mg/m 3). TSM measurements were more numerous than Turbidity observations. In the satellite products the two are interchangeable through a single conversion factor so results will be comparable.

6.1.2. Satellite data

Due to scarcity of recent in situ data for the OLCI observation period, the validation study described in this section is still based on MERIS. The 3rd reprocessing of MERIS full resolution L1B data was used for CDRP V1.0. These were processed to LWLR by applying radiometric calibration (SNAP toolbox, see ATBD for details) and Polymer (v4.6; Steinmetz

et al. (2011)). We note that the version of POLYMER used in the CDRP v1.0 is more recent (v4.12) but this does not cause differences in the handling of MERIS data. MODIS L1B data were obtained from the National Aeronautics and Space Administration (NASA) for the period 2009-2019, providing overlap with MERIS and OLCI to evaluate inter-sensor bias. The L1B data were masked using a combination of L1B and Idepix flags (see ATBD for details). The masks applied were Cosmetic, duplicated, glint risk, suspect, land/ocean, bright, coastline, and invalid from the L1B product and invalid, cloud, cloud ambiguous, cloud sure, cloud buffer, cloud shadow, snow ice, bright, white, coastline, land and glint risk from Idepix.

6.2. Comparison methods

Product validation of the LWLR and derived products (Chl-a and TSM) is based on comparison against in situ observations. Results presented here are for in situ validation carried out against satellite observations with the MERIS sensor, for which the most in situ data are available by far.

The satellite matchups were extracted from 3×3 pixel windows with a temporal window of 3 days. The mean value was calculated from the macro-pixel for the MERIS and MODIS paths which most closely matched the time of the observations. If the same satellite pass was found to be a valid match-up for multiple in situ observations from the same location, then the nearest in situ value in time was selected for the match-up.

Depending on which product was being validated, statistical measures of performance included the coefficient of determination (R²), Root-Mean-Square difference (RMS), Normalized RMS difference (NRMS), Mean Absolute Percentage difference (MAP) and bias.

6.3. Description of work

6.3.1. Lake Water-Leaving Reflectance

For water-leaving reflectance, comparisons between the in situ and satellite measurements were performed for each band. In this report, the atmospheric-corrected MERIS water-leaving reflectance were converted to remote sensing reflectance (Rrs = Rw/ π [sr⁻¹]) to facilitate the comparison between the in situ and satellite measurements.

6.3.2. Chlorophyll-a and TSM

6.3.2.1 MERIS algorithms and their assignment to OWTs

After extensive product validation in the GloboLakes project, the water constituent retrieval algorithms listed in Table 42 were identified as showing individual best performance against sets of Optical Water Types (Neil et al. 2019). For Chl-a, the algorithms included for validation are Gons05 (Gons et al. 2005), the NASA OC2 algorithm (O'Reilly et al. 1998), R708/R665 (Gilerson et al. 2010), and QAA (Mishra et al. 2013a; Mishra et al. 2013b). For TSM, the algorithms include those of Zhang et al. (2014), Vantrepotte et al. (2011), and Binding et al. (2010).

These algorithms are mapped to pixels depending on their similarity to a set of Optical Water Types (OWT), determined from in situ reflectance data in the GloboLakes project (Spyrakos et al. 2018). The assignment of algorithms to each OWT is shown in Table 42,

which also provides the calibrated algorithm coefficients. It is noted that while the methodology of algorithm tuning is as described in Neil et al. (2019), tuned coefficients may differ since the former are derived from in situ reflectance data against LIMNADES while *Calimnos* uses coefficients optimized for satellite-derived Polymer-corrected water-leaving reflectance (Rw).

Table 42 Chlorophyll-a/TSM algorithms per optical water type and tuned parameters for MERIS

Product	Algorithm	Optical Water Type number	Parameters	Tuned
Chl-a	OC2 oceancolor.gsfc.nasa.gov/cms/atb d/chlor_a	3, 9, 10, 13	a0, a1, a2, a3, a4	0.1731, - 3.9630, - 0.5620, 4.5008, - 3.0020
	708/665 empirical band ratio based on Gilerson et al. (2010)	2, 8, 11, 12	a, b, c	79.62, 0.7393, -54.99
	Semi-analytical NIR-Red band algorithm for MERIS based on Gons et al. (2005).	1, 4, 5, 6	aph*	0.025
	Adapted QAA algorithm according to Mishra et al. (2013)	7	S _{CDOM}	0.0135
TSM	Based on Zhang et al. (2014)	1, 7, 10	a, b	2524, 1.113
	Based on Vantrepotte et al. (2011)	2, 4, 6, 8, 12	a, b, c	206.4, 20460, 0.7921
	Based on Binding et al. (2010)	3, 5, 9, 11, 13	b* _{SPM}	0.664

6.3.2.2 MODIS algorithms assignment to OWTs

A Round-robin comparison was performed to assess the selected algorithms for MODIS. For Chl-a, 9 algorithms were compared, including three blue-green band ratio algorithms, two NIR-red band ratio algorithms, one peak height algorithm, and two semi-analytical algorithms (Table 43). For TSM, a total of 11 algorithms were included in the comparison, which were all empirical algorithms of red or/and NIR bands (Table 44).

Table 43 Summary of validated Chl-a models tested for MODIS

Model	Architectur al approach	Bands	Original training range (mg.m ⁻³)	Reference
OC3	Blue-green band ratio	min[443, 488], 547	0.012 - 77	https://oceancolor.gsfc.nasa.gov/ atbd/chlor_a/
OC2	Blue-green band ratio	488, 547	0.012 - 77	https://oceancolor.gsfc.nasa.gov/ atbd/chlor_a/
OC2_ HI	Blue-green band ratio	469, 555	0.012 - 77	https://oceancolor.gsfc.nasa.gov/ atbd/chlor_a/
R748_ 667	NIR-red band ratio	748, (667 or 678)	4 - 240	Dall'Olmo et al. 2005; Gitelson 1992; Gitelson et al. 2008; Gitelson et al. 2007; Gurlin et al. 2011
Shi	NIR-red band ratio	645, 859	6.6 - 113.7	Shi et al. 2015a
Appel		645, 859, 469	2.9 - 91	El-Alem et al. 2012
FLH	Peak height	665, 677, 746	1 - 10	Letelier and Abbott 1996
QAA_ v6	Semi- analytical	\	0 ~ 70	https://www.ioccg.org/groups/Sof tware_OCA/QAA_v6_2014209.pdf
GSM	Semi- analytical	\	0.02 - 10	Maritorena et al. 2002

Table 44 Summary of validated TSM models tested for MODIS

Model	Architectural approach	Bands	Original training range	Reference
Petus	single red band	645	TSM <30 mg L-1	Petus et al. 2010
ChenS	NIR-red ratio	log(859)/log(645)	5.8 ~ 577.2 mg L-1	Chen et al. 2015
Wang	single NIR band	869	Turb 1~300 NTU	Wang et al. 2012
Doxaran	NIR-red ratio	859/645	Turb 77.4 ~2193 NTU	Doxaran et al. 2009
			TSM 77 ~ 2182 mg L-1	
Dogliotti	red or NIR	645, 859	Turb 1.8 ~ 988 FNU	Dogliotti et al. 2015
Miller	single red band	645	0 ~ 55 mg L-1	Miller and McKee 2004; Sipelgas et al. 2006

Model	Architectural approach	Bands	Original training range	Reference
Shi	single red band	645	0 ~ 300 mg L-1	Shi et al. 2015b; Zhao et al. 2011
Ondrusek	single red band	645	0 ~ 16 mg L-1	Ondrusek et al. 2012
ChenZ	single red band	645	Turb 0.9 ~ 8 NTU	Chen et al. 2007
Hu	red and NIR	645-859	0 ~ 12 mg L-1	Hu et al. 2004
Zhang	three bands	488, 555, 645	1~ 300 mg L-1	Zhang et al. 2010

The above listed Chl-a/TSM algorithms were evaluated in three different forms:

- Original form (denoted ORG), which adopts the proposed original parameterization of the model coefficients from the literature.
- Calibrated form (denoted CAL), model coefficients were tuned for each model using the entire in-situ training dataset.
- Clustered form (denoted CLUS), coefficients were fitted using a subset of the dataset for each OWT to estimate Chl-a/TSM. For Chl-a, the matchups with top-40% of membership scores for each OWT were adopted in the algorithm tuning; while for TSM, matchups with top-70% of membership scores for each OWT were used in the algorithm tuning.

Afterward, the MODIS water constituent retrieval algorithms listed in Table 45 were identified as showing individual best performance against the assigned OWTs. For Chla, the algorithms included in the per-OWT blending and validation were the NASA OC2, OC3, OC2_HI algorithms (O'Reilly et al. 1998), and R748_667 algorithm (Dall'Olmo et al. 2005). For TSM, the algorithms included in the per-OWT blending and validation were Miller (Miller and McKee 2004), Ondrusek (Ondrusek et al. 2012), ChenZ (Chen et al. 2007), Petus (Petus et al. 2010), and Zhang (Zhang et al. 2010). For all of the selected MODIS algorithms, it is notated that they were specifically tuned for each OWT using matchups with relatively high membership scores (top-40% for Chla and top-70% for TSM), meaning that the tuned coefficients vary between OWTs even for the same algorithm.

Table 45 Chlorophyll-a/TSM algorithms per optical water type and tuned parameters for MODIS

Product	Algorithm	Optical Water Type number
Chl-a	OC2 oceancolor.gsfc.nasa.gov/cms/atbd/chlor_a	1, 5, 7, 9, 12,13
	OC3 oceancolor.gsfc.nasa.gov/cms/atbd/chlor_a	2, 3, 8
	748/667 empirical band ratio based on Dall'Olmo et al. (2005)	4, 6, 11
	OC2_HI oceancolor.gsfc.nasa. gov/cms/atbd/chlor_a	10
TSM	Based on Miller and McKee (2004)	1, 12
	Based on Ondrusek et al. (2012)	2, 6, 11, 13

Product	Algorithm	Optical Water Type number
	Based on Chen et al. (2007)	3, 5, 9
	Based on Petus et al. (2010)	4, 8, 10
	Based on Zhang et al. (2010)	7

6.3.2.3 Weighted-blending procedureThe satellite retrieved water constituent products provided in the CDRP are based on a weighted blending procedure, recombining the individual algorithm results with the weighting determined by their OWT membership scores. The selected algorithms are mapped to individual satellite measurement (per pixel) from the OWTs with the top-3 classification scores. The algorithm results corresponding to those three OWTs are averaged using the membership score as weighting factor, after normalizing the scores between 0 and 1 where 1 is the highest score and 0 is the score of the 4th ranking OWT. The derived Chl-a or TSM satellite products following this procedure were compared with the in situ matchups for validation for this analysis.

In this report, the validation of the selected algorithms is performed based on the in situ TSM observations. Turbidity is obtained using a conversion factor of $1.17 \text{ NTU/g m}^{-3}$ as formulated by (Nechad et al. 2016; Nechad et al. 2010).

6.3.3. Inter-sensor consistency evaluations

For the 250 largest lakes, initial consistency evaluations were performed between MERIS/OLCI and MODIS during 3 years of overlap in 2009-2011 (MERIS vs MODIS) and 2017-2019 (OLCI vs MODIS), respectively.

6.3.3.1 Data extraction

L3 per-day composited data were generated using SNAP v7.0, which were re-projected to the same grid at 1 km for MERIS, OLCI, and MODIS, respectively. For each sensor on each day, three pixels on the margin of lakes were firstly removed to reduce the influence of land adjacency, percentiles of 20, 50, and 80 were then calculated and saved for further consistency evaluations.

6.3.3.2 Statistical analysis

Two variables were selected for the evaluation including one water quality parameter (Chl-a) and one water-leaving reflectance band (polymer corrected Rw at 488/490 nm). It is noted that 488 nm is a MODIS band, and 490 nm is the corresponding MERIS/OLCI band. Based on previous inter-sensor analyses on LWLR, $R_w(488-490)$ was one of the bands showing the best consistency between MODIS and MERIS/OLCI. For brevity, $R_w(488)$ was used to refer this band for the three sensors thereafter.

For all of the large lakes, two statistic analyses were conducted on extracted daily median values of the two variables (Chla and Rw 488 nm).

Two-sided t-test (scipy.stats.ttest_rel in python) on related samples: This is a dependent t-test for the null hypothesis that 2 related or repeated samples have identical average (expected) values. If the calculated *p*-value is below the threshold chosen for statistical significance (0.001 was chosen for this study), then the null hypothesis is rejected in favor of the alternative hypothesis.

Linear regression analysis was also performed on the daily extracted median values between sensors for each lake, correlation coefficient (denoted R) was calculated as an indicator of the strength and direction of the linear relationship between the observations of two sensors.

Following the above analysis, lakes with P>0.001 (from t-test) or R>0.5 were selected for a further visual check, which finally resulted in 38 lakes that show sufficient consistency. It is also noticed that Turbidity is not included from MODIS as the algorithm performance is not adequate.

6.4. Validation results

6.4.1. Validation results for MERIS

6.4.1.1. LWLR from MERIS

Previous round-robin comparisons of atmospheric correction algorithms for MERIS carried out in GloboLakes showed that Polymer yielded the statistically most robust retrieval of reflectances, although a systematic negative bias was observed. This led to POLYMER being adopted as the state-of-the-art atmospheric correction processor for Lakes_cci to provide LWLR. Figure 146 shows R_{rs} matchup results for MERIS in 11 lakes at 11 wavebands from 412 nm to 779 nm, for which in situ data were available in LIMNADES. In general, significant linear relationships were found between the MERIS and in situ R_{rs} , with the highest R = 0.83 returned in the 560 nm band and outliers in the blue bands associated with the hypereutrophic Hartbeespoort Reservoir giving the worst performance. Systematic underestimation of MERIS R_{rs} is still observed. From inspecting the individual components of the atmospheric correction procedure, the underestimation is understood to be the result of overestimating the atmospheric path radiance with increasingly turbid waters, rather than a failure of the in-water bio-optical model or glint retrieval.

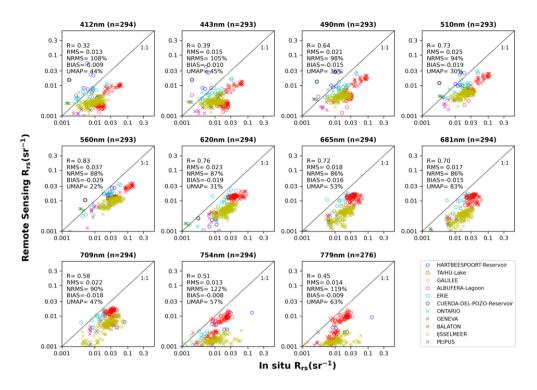


Figure 135 Comparison between in situ and MERIS R_{rs} in each band

To show the level of spectral consistency in the validation data set, Figure 136 presents averaged spectra of all R_{rs} matchups for each individual lake. There is general agreement between the in situ and satellite derived reflectance spectra, with the exception of the Cuerda-del-Pozo Reservoir.

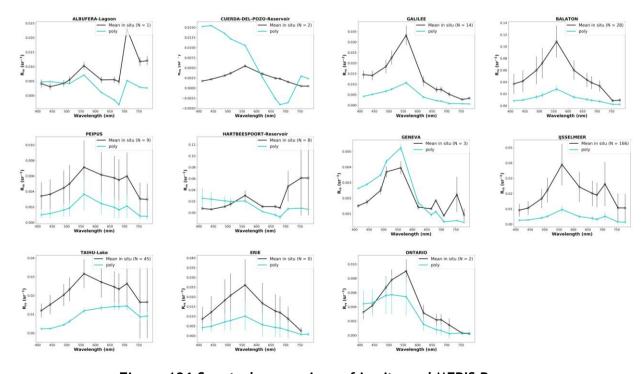


Figure 136 Spectral comparison of in situ and MERIS R_{rs}

6.4.1.2. Chlorophyll-a from MERIS

Scatter plots of Chl-a validation results for each pre-tuned algorithm are shown in Figure 137 using the whole data set regardless of optical water type classification. For the OC2 algorithm, although it returned the highest Normalized Root Mean Squared Error (NRMS) of 172%, the differences are evenly distributed around unity resulting in a slope of linear regression line close to 1. Saturation of the algorithm appears, as may be expected, at concentrations > 10 mg m⁻³. Algorithms R708_R665, Gons05, and QAA show a general overestimation at low Chl-a values, and an underestimation at high Chl-a values. The highest R² of 0.42 and lowest Normalized-Root-Mean-Squared-Error (NRMS) of 38% are returned by the QAA among the four algorithms.

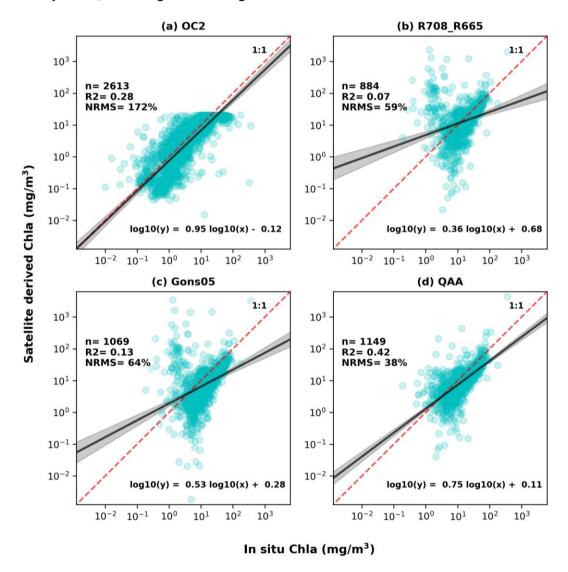


Figure 137 Comparison between in situ and MERIS-derived chlorophyll-a using the (a) OC2, (b) R708_R665, (c) Gons05, and (d) QAA algorithms.

The agreement between in situ and satellite retrieved Chl-a improves dramatically when the input algorithms are blended according to the per-pixel Optical Water Type membership (Figure 138). The regression line is then close to unity, with an R^2 of 0.69 and

NRMS of 78%, both measured in log space. It is noted that these results do not provide independent validation of the algorithms, since there is large overlap with the data sets used for algorithm calibration. The result does illustrate the positive effect of the weighted blending procedure.

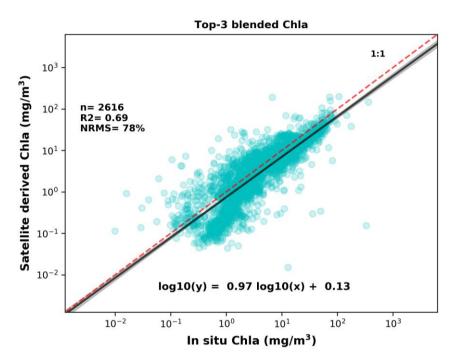


Figure 138 Comparison between in situ and top-3 blended chlorophyll-a derived from MERIS.

6.4.1.3. Total Suspended Matter from MERIS

The performance of individual TSM algorithms is shown in Figure 139, comparing in situ TSM measurements to those retrieved from MERIS matchup data. A general underestimation at the high TSM section was observed for all of the three algorithms. The slope of the regression line (expressed in linear scaling) for the Zhang algorithm approached unity with an intercept of -0.296 g m⁻³ suggesting systematic underestimation. The Vantrepotte algorithm showed the lowest NRMS (58%) among the three algorithms, although the lowest R² (0.29) was also returned by this algorithm indicating it lacks broad sensitivity. The Binding algorithm showed the highest R² of 0.65 and with the largest NRMS of 65%. The Binding algorithm shows no decrease in sensitivity with increasing concentrations, despite the increasing bias. These results suggest that a re-tuning of the Zhang and Binding algorithms from the original analysis carried out in the GloboLakes project would likely improve overall performance.

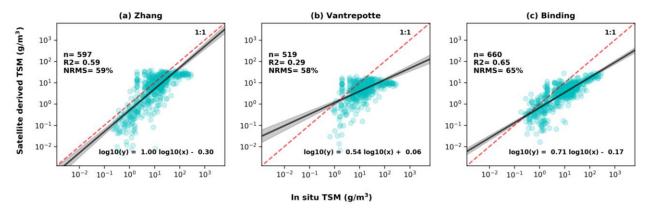


Figure 139 Comparison between in situ and retrieved TSM using the (a) Zhang, (b) Vantrepotte, and (c) Binding algorithms.

Agreement between the in situ and satellite retrieved TSM slightly improved with algorithm blending by OWT membership (Figure 140). Negative bias at the high TSM values is significant and confirms that additional algorithm candidates or re-tuning are needed. At present, the R^2 is 0.61 with NRMS = 54%.

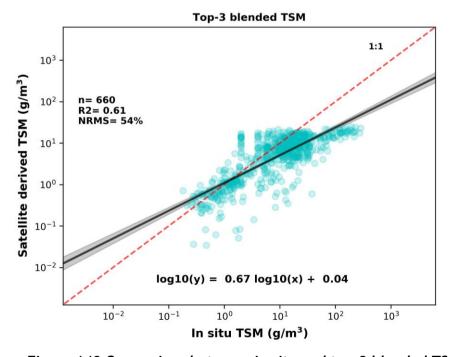


Figure 140 Comparison between in situ and top-3 blended TSM

6.4.2. Validation results for MODIS

6.4.2.1. LWLR from MODIS

Two atmospheric correction algorithms were compared for MODIS: POLYMER and l2gen. The performance of POLYMER-corrected MODIS R_{rs} matchups was assessed with in situ Rrs at 11 bands from 412 nm to 748 nm (Figure 141). Significant linear correlations were found for all bands between POLYMER-corrected and in situ R_{rs} , with the highest R = 0.83 returned at 547 nm. A systematic underestimation was observed with bias ranging from -0.004 sr-1 at 412 nm to -0.023 sr-1 at 555 nm.

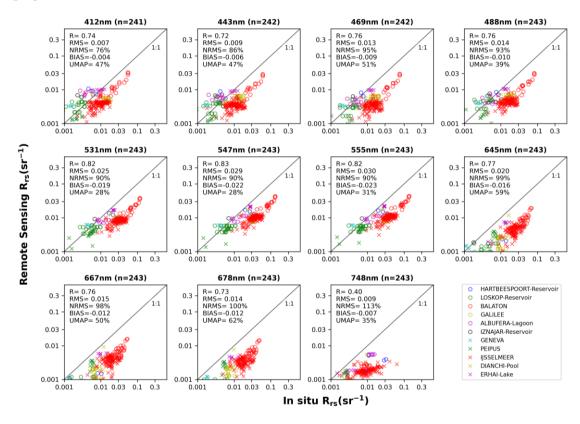


Figure 141 POLYMER v4.12 matchups of MODIS with in situ reflectance data from LIMNADES, using a ±3 day matchup window and 3x3 pixel extraction window.

The performance of l2gen was also evaluated against the in situ reflectance data for 11 bands from 412 to 748 nm. Similar to POLYMER, significant linear correlations were found for all bands (Figure 142). The highest R = 0.85 was returned at band 555nm and the lowest R of 0.33 was returned in band 748 nm. A systematic underestimation was observed with bias ranging from -0.005 sr⁻¹ in band 748 nm to -0.027 sr⁻¹ in band 555 nm.

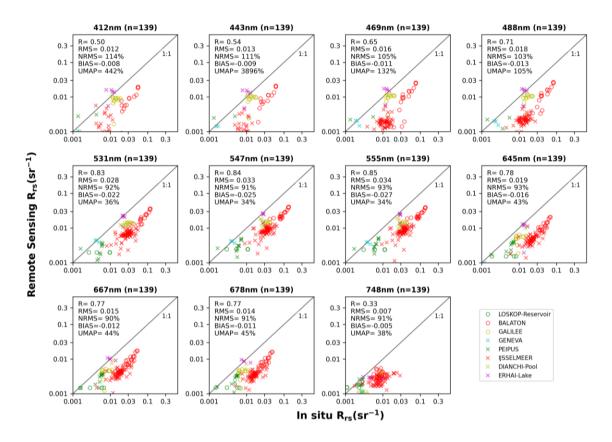


Figure 142 l2gen matchups of MODIS with in situ reflectance data in LIMNADES, using a ±3 day matchup window and 3x3 pixel extraction window.

Correlation analysis between POLYMER and l2gen reflectance products show significant linear relationships for bands between 412 nm and 748 nm, with highest R = 0.974 returned at 555 nm (Figure 143). Distributions are close to unity for bands between 469 nm and 678 nm, with high R values (>0.8). POLYMER results are lower at short wavebands (bands 412 nm and 443 nm), while for red-NIR bands (748 nm, 859 nm, and 869 nm) POLYMER is generally higher than l2gen-corrected reflectance.

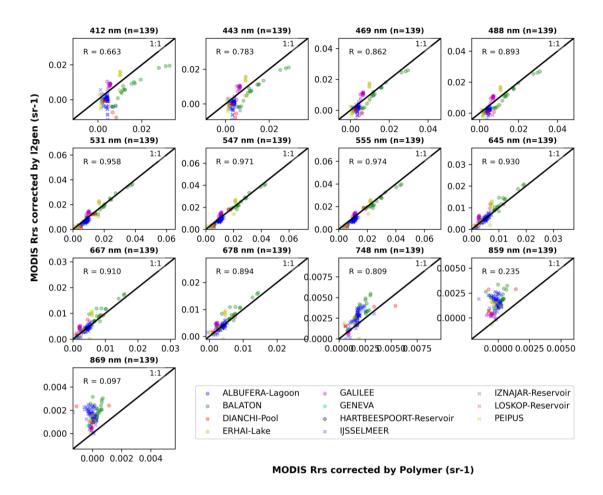


Figure 143 comparison between POLYMER and I2gen corrected reflectance

6.4.2.2. Chlorophyll-a from MODIS

Figure 144 shows Round-robin comparison results of Chl-a generated from each of the examined algorithms against the in situ measurements. Corresponding error metrics are shown in Figure 145. OCX, R748_667, and QAA algorithms demonstrate a combined capability to estimate Chl-a. Notably, algorithms OC2 and OC3 produce R values of 0.66 and NRMS of ~40% when compared to in situ measurements. Algorithms Shi, Appel, and FLH are shown to perform poorly over the observed concentration range. The apparent failures that occur with algorithms Shi and Appel may be related to their usage of $R_w(859)$, which is not used in other solutions, and leads to negative estimates of the calculated index. Among the three model forms, the per-cluster optimized algorithms ('Clus') provide the most accurate estimation of Chl-a.

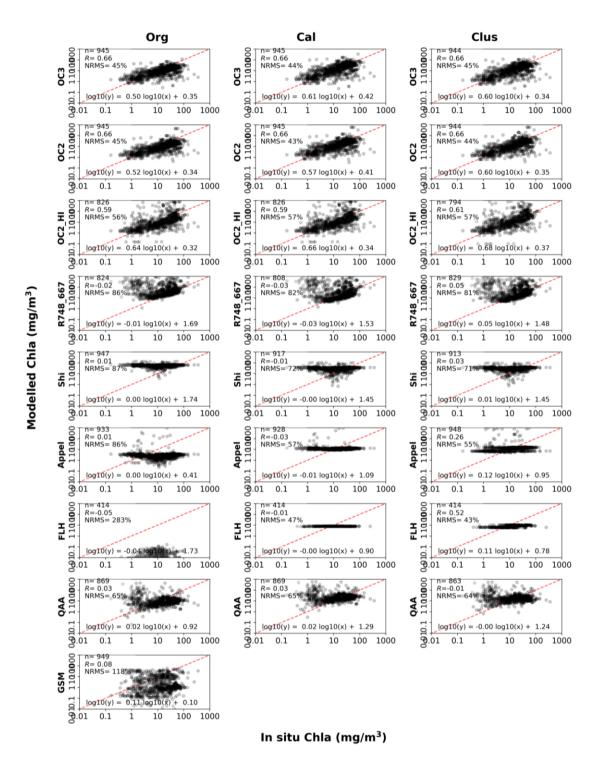


Figure 144 Comparison of in situ measured and algorithm retrieved Chl-a for MODIS with the original algorithm definition in the first column, calibration against the whole dataset in the second column and per-cluster optimized algorithms in the third column. Red line marks unity.

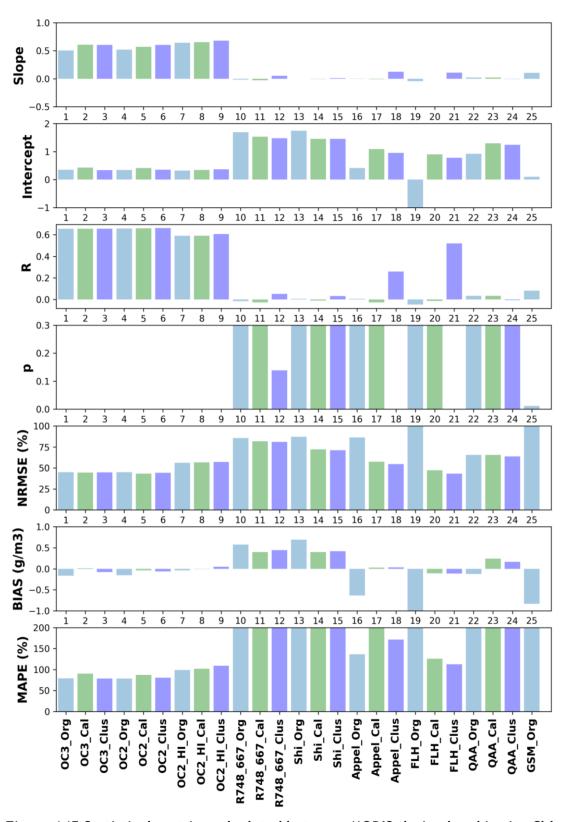


Figure 145 Statistical metrics calculated between MODIS-derived and in situ Chl-a

Based on the Round-robin comparison shown above, four Chl-a algorithms were selected and tuned per-OWT for the Chl-a retrieving, including NASA OC2(O'Reilly et al. 1998), OC3,

OC2_HI algorithms (O'Reilly et al. 1998), and R748_667 algorithm (Dall'Olmo et al. 2005). The final Chl-a product is then the result of the weighted-blending procedure described in section 6.3.2.3, with weights determined by the per-pixel OWT membership scores. The comparison between in situ and final Chl-a product is shown in Figure 146, with R value of 0.69 and NRMS of 43%.

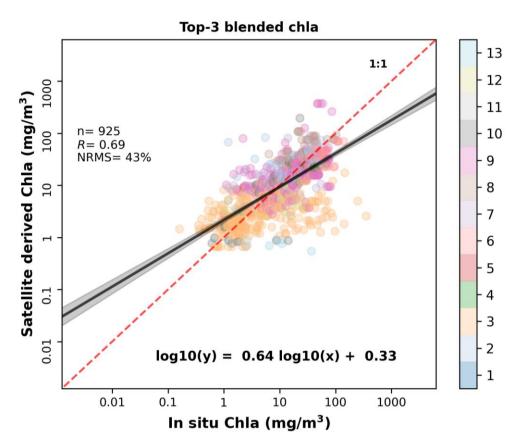


Figure 146 Comparison between in situ and top-3 blended Chl-a for MODIS. Colour coding refers to the most similar OWT.

6.4.2.3. Total Suspended Matter from MODIS

Figure 147 shows Round-robin comparison results of TSM generated from each of the examined algorithms against the in situ measurements, with corresponding error metrics shown in Figure 148. It is found that Miller, Ondrusek, ChenZ, Petus, and Zhang algorithms demonstrate the best TSM retrieval ability, after OWT-specific tuning. Algorithms ChenS, Wang, Doxaran, Dogliotti, Shi, and Hu performed relatively poorly. Several algorithms yielded a limited number of valid retrievals (e.g., ChenS, Wang, and Dogiliotti), which may be related to the usage of the 859 nm waveband leading to negative estimates.

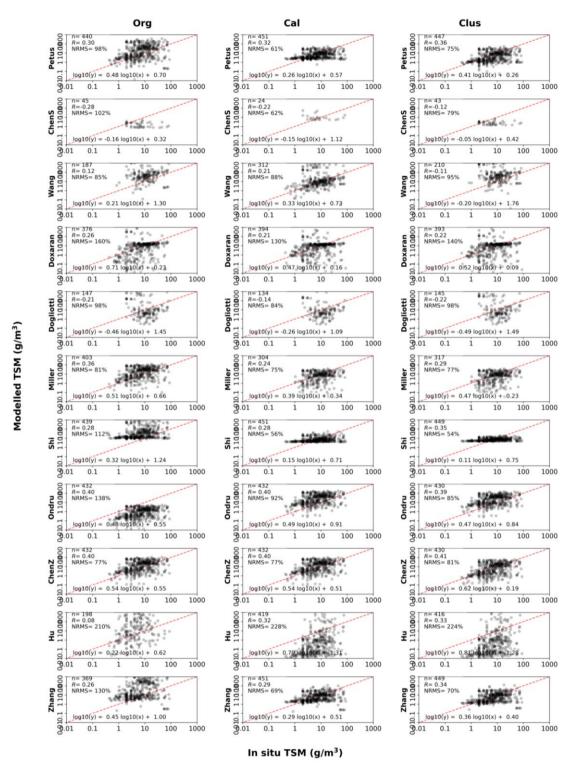


Figure 147 Comparison of in situ and MODIS-derived TSM using original, tuned and per-OWT tuned algorithms. The 1:1 line between in situ measured and algorithm retrieved TSM is represented by a red dashed line.

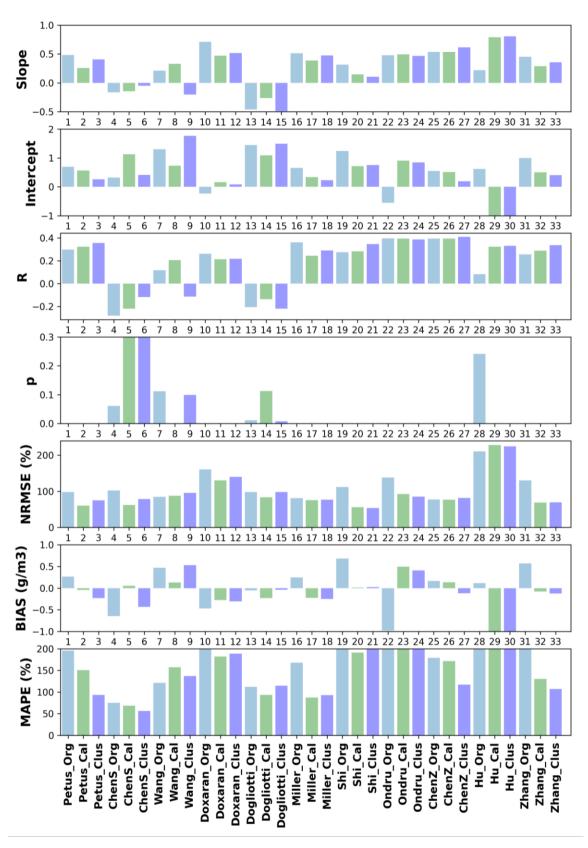


Figure 148 Statistical metrics calculated between in situ and MODIS-derived TSM

From the Round-robin experiment conducted above, several TSM algorithms were identified as having potential for TSM retrieval from MODIS, including Miller (Miller and

McKee 2004), Ondrusek (Ondrusek et al. 2012), ChenZ (Chen et al. 2007), Petus (Petus et al. 2010), and Zhang (Zhang et al. 2010). These algorithms were tuned per-OWT for best performance, following which the satellite retrieved TSM was calculated and blended according to per-pixel OWT membership scores, following the blending procedure described in section 6.3.2.3. The comparison between in situ and satellite retrieved TSM is shown in Figure 149, with a slope of 0.59 and an intercept of 0.34. Large dispersion is found in data points with highest similarity to OWT 3 (clear water). Overall, the optimized solution for TSM is not considered to perform well, and TSM (Turbidity) retrieval with MODIS is not recommended.

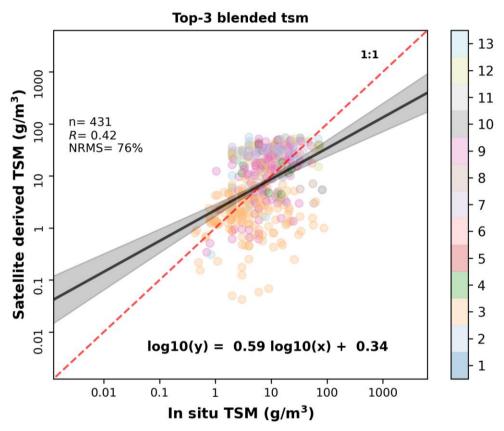


Figure 149 Comparison between in situ and top-3 blended TSM for MODIS. Color coding shows the dominant OWT.

6.4.3. Inter-sensor consistency evaluation

Following statistical analysis of Chl-a and LWLR (at 488nm), 38 lakes out of the largest 250 lakes in the Lakes_cci were identified as showing sufficient consistency between MERIS/OLCI and MODIS. Turbidity (or TSM) was not included in the inter-sensor consistency evaluation since these algorithms did not show adequate performance (see previous section 6.4.2.3), The lakes included in the CRDP v2.0 for MODIS to fill the gap between MERIS (April 2012) and OLCI (April 2016) are listed in Table 46.

Table 46 List of lakes provided with MODIS data to fill the gap between MERIS and OLCI

	Region name	Lake name	
1	GLBL00000001	Curonian	
2	GLBL00000012	IJsselmeer	

3	;	GLBL00000016	Markermeer
4	ļ	GLWD00000002	Superior
5	i	GLWD00000003	Victoria
6	i	GLWD0000006	Michigan
7	•	GLWD00000012	Erie
8	3	GLWD00000015	Ontario
9)	GLWD00000020	Titicaca
10	0	GLWD00000023	Athabasca
1:	1	GLWD00000026	Great Salt
12	2	GLWD00000037	Manitoba
13	3	GLWD00000040	Zaysan
14	4	GLWD00000041	Qinghai
15	5	GLWD00000042	Cabora Bassa
16	6	GLWD00000051	Van
17	7	GLWD0000053	Uvs
18	8	GLWD0000057	Cedar
19	9	GLWD0000058	Alakol
20	0	GLWD0000066	Tai
2:	1	GLWD00000067	Kivu
22	2	GLWD00000071	Tsimlyanskoye
23	3	GLWD00000077	Buhayrat ath Tharthar
24	4	GLWD00000079	Kakhovskoye
25	5	GLWD00000098	Argyle
26	6	GLWD00000120	Tengiz
27	7	GLWD00000121	Khyargas
28	8	GLWD00000131	Qapshaghay Bogeni
29	9	GLWD00000135	Sevan
30	0	GLWD00000146	Saint Clair
3:	1	GLWD00000147	Lesser Slave
32	2	GLWD00000179	Zhari Namco
33	3	GLWD00000181	Bosten
34	4	GLWD00000190	Novosibirskoye
35	5	GLWD00000194	Salton
36	6	GLWD00000205	Bay
37	7	GLWD00000239	Ulungar
38	8	GLWD00000241	Sarykamyshskoye

Figure 150 and Figure 151 show the inter-sensor comparison results for Chl-a, Turbidity and LWLR between MERIS/OLCI and MODIS in lakes Van (in Turkey) and Sevan (in Armenia). For Chl-a, MERIS and MODIS retrievals show similar variation trends in both lakes, while MODIS results are generally higher than MERIS. Higher consistency on Chl-a is found between OLCI and MODIS for both lakes, although slightly higher values are obtained with MODIS compared to OLCI in the low Chl-a range (<1 mg/m³) in Lake Van. As expected, the MODIS Turbidity retrievals do not show good consistency with MERIS and OLCI for either lake. As for LWLR, we find adequate consistency between MERIS/OLCI and MODIS for all of the seven LWLR bands for both lakes. In lake Van, high correlation relationships are found between MERIS/OLCI and MODIS for all seven LWLR bands with correlation coefficients (R) over 0.7 (Figure 150). In lake Sevan, high correlation relationships (R>0.8) are found between MERIS/OLCI and MODIS at short (blue) bands (Figure 151).

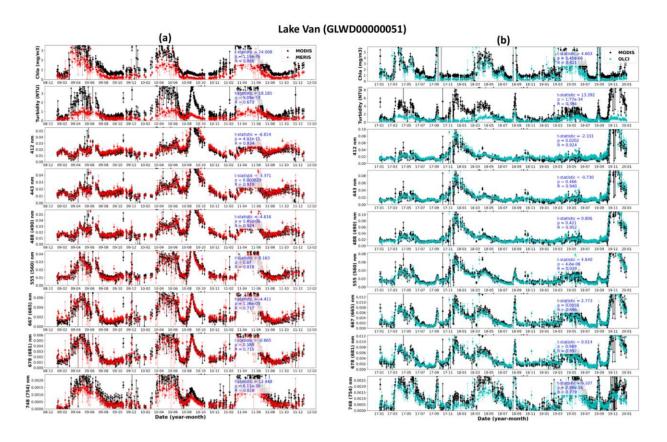


Figure 150 Inter-sensor comparison results of chlorophyll-a, turbidity and LWLR between (a) MERIS and MODIS (2009-2011), (b) OLCI and MODIS (2017-2019) in Lake Van. LWLR of seven shared bands with < 6 nm differences between MERIS/OLCI and MODIS are shown, with MODIS bands indicated in brackets where they differ from the MERIS/OLCI bands.

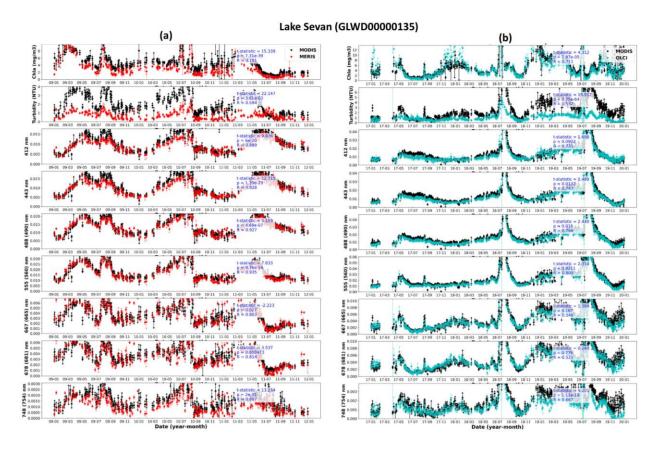


Figure 151 Same as Figure 150, but for Lake Sevan.

6.5. Conclusions and recommendations

For MERIS and OLCI, systematic negative bias remains visible in the latest R_{rs} validation results. This is attributed to the challenging conditions for atmospheric correction in optically complex inland waters, including non-zero reflectance in the near infra-red and adjacent land. This underestimation in R_{rs} is propagated to Chl-a and TSM, but corrected again by end-to-end algorithm tuning. This is an essential step because in situ matchup data of Chl-a and TSM are far more numerous than those including R_{rs} .

Validation of the Chl-a algorithms shows that the weighted-blending procedure dramatically improved the retrieval performance compared to that of the individual algorithms. For TSM, the agreement between the in situ and satellite retrieved TSM slightly improved after the blending stage compared to the individual algorithms, which show room for individual improvement using currently available data sets.

The validation presented in this report is only based on the MERIS matchups, because of the scarce in situ data coincident with OLCI. Analyses completed in the Copernicus Land Monitoring Service, which uses the same configuration of *Calimnos*, have shown that there is consistency between per-lake time series of Chl-a and Turbidity observed with MERIS (2002-2012) and OLCI (2016-present), which strongly suggests that the same algorithms may be applied to both sensors until further in situ reference data become available. However, the additional bands available on OLCI may well be able to improve on aspects of the retrieval of both LWLR and the derived biogeochemical products.

Validation of MODIS reflectance and water quality products show the capability of using MODIS products to fill the data gaps between MERIS and OLCI for specific lakes. POLYMER was evaluated against the NASA operational atmospheric correction algorithm for MODIS (l2gen), with both showing systematic negative bias in the $R_{\rm rs}$ for the bands evaluated, due to the challenges for accurately performing atmospheric correction over inland optical complex waters. POLYMER was found to produce almost double of valid matchup numbers with in situ measurements compared to l2gen, demonstrating its capability of handling optically complex waters better, and therefore adopted here. However, it should be noted that the validation data set does not adequately include all optical water conditions and there is risk unnecessary masking, loss or misrepresentation of LWLR with POLYMER, particularly in extremely turbid waters.

Validation of Chl-a and TSM algorithms shows that the per-OWT tuning and weighted-bending procedure dramatically improved the retrieval performance compared to that of the individual algorithms. The overall performance of Chl-a and TSM algorithms for MODIS was relatively poor compared to that of MERIS, which is reasonable considering that MODIS is more prioritized to oceanic waters.

By combining MODIS into CRDP v2, an uninterrupted global inland water quality data set can be generated. Currently, a total of 38 lakes are selected to be provided with the MODIS dataset following primary inter-sensor evaluations with MERIS/OLCI, which are believed to provide more reliable observations and suffering less from land adjacency effects. Further evaluation of product consistency over time with MERIS and OLCI for selected lakes is pending, before this work can be fully expanded to all lakes in CRDP v2.0

Finally, it should be noted that, thus far, algorithm calibration, validation and uncertainty characterisation have made use of all available (under license) data for each analysis. In future, data sets of sufficient density will need to be split into calibration and validation data sets, with the latter contributing to the characterisation of product uncertainty.

7. References

Binding, C., Jerome, J., Bukata, R., & Booty, W. (2010). Suspended particulate matter in Lake Erie derived from MODIS aquatic colour imagery. International Journal of Remote Sensing, 31, 5239-5255

Carrea, L., Embury, O. and Merchant, C. J. (2015) Datasets related to in-land water for limnology and remote sensing applications: distance-to-land, distance-to-water, water-body identifiers and lake-centre co-ordinates. Geoscience Data Journal, 2(2). pp. 83-97. ISSN 2049-6060 doi:10.1002/gdj3.32

Chen, S., Han, L., Chen, X., Li, D., Sun, L. and Li, Y. (2015). Estimating wide range Total Suspended Solids concentrations from MODIS 250-m imageries: An improved method. ISPRS Journal of Photogrammetry Remote Sensing, 9958-69.

Chen, Z., Hu, C. and Muller-Karger, F. (2007). Monitoring turbidity in Tampa Bay using MODIS/Aqua 250-m imagery. Remote Sensing of Environment, 109(2), 207-220.

Dall'Olmo, G., Gitelson, A.A., Rundquist, D.C., Leavitt, B., Barrow, T. and Holz, J.C. (2005). Assessing the potential of SeaWiFS and MODIS for estimating chlorophyll concentration in turbid productive waters using red and near-infrared bands. Remote Sensing of Environment, 96(2), 176-187.

Dogliotti, A.I., Ruddick, K., Nechad, B., Doxaran, D. and Knaeps, E. (2015). A single algorithm to retrieve turbidity from remotely-sensed data in all coastal and estuarin waters. Remote Sensing of Environment, 156157-168.

Doxaran, D., Froidefond, J.-M., Castaing, P. and Babin, M. (2009). Dynamics of the turbidity maximum zone in a macrotidal estuary (the Gironde, France): Observations from field and MODIS satellite data. Estuarine, Coastal Shelf Science, 81(3), 321-332.

El-Alem, A., Chokmani, K., Laurion, I. and El-Adlouni, S.E. (2012). Comparative analysis of four models to estimate chlorophyll-a concentration in case-2 waters using MODerate resolution imaging spectroradiometer (MODIS) imagery. Remote Sensing, 4(8), 2373-2400.

Embury, O., Merchant, C. J. and Corlett G.K. (2012) A reprocessing for climate of sea surface temperature from the along-track scanning radiometers: Initial validation, accounting for skin and diurnal variability effects. Remote Sensing of Environment, 116. pp. 62-78. ISSN 0034-4257 doi:10.1016/j.rse.2011.02.028

Gilerson, A.A., Gitelson, A.A., Zhou, J., Gurlin, D., Moses, W., Ioannou, I., & Ahmed, S.A. (2010). Algorithms for remote estimation of chlorophyll-a in coastal and inland waters using red and near infrared bands. Optics Express, 18, 24109-24125

Gitelson, A. (1992). The peak near 700 nm on radiance spectra of algae and water: relationships of its magnitude and position with chlorophyll concentration. International Journal of Remote Sensing, 13(17), 3367-3373.

Gitelson, A.A., Dall'Olmo, G., Moses, W., Rundquist, D.C., Barrow, T., Fisher, T.R., Gurlin, D. and Holz, J. (2008). A simple semi-analytical model for remote estimation of chlorophyll-a in turbid waters: Validation. Remote Sensing of Environment, 112(9), 3582-3593.

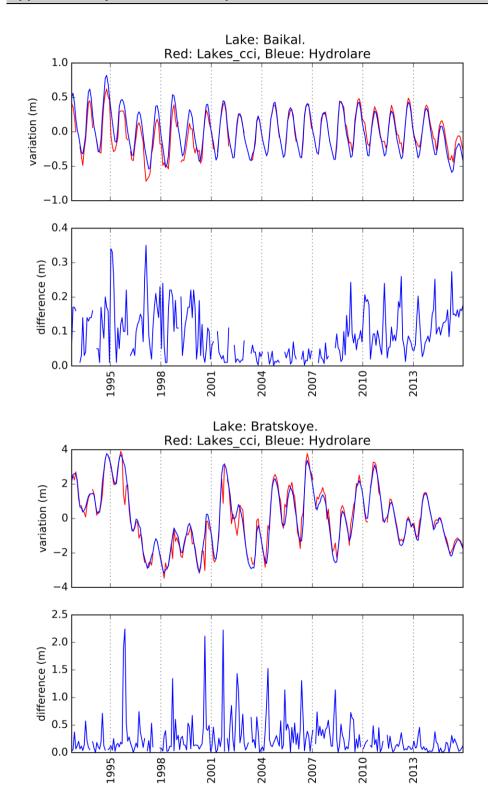
Gitelson, A.A., Schalles, J.F. and Hladik, C.M. (2007). Remote chlorophyll-a retrieval in turbid, productive estuaries: Chesapeake Bay case study. Remote Sensing of Environment, 109(4): 464-472.

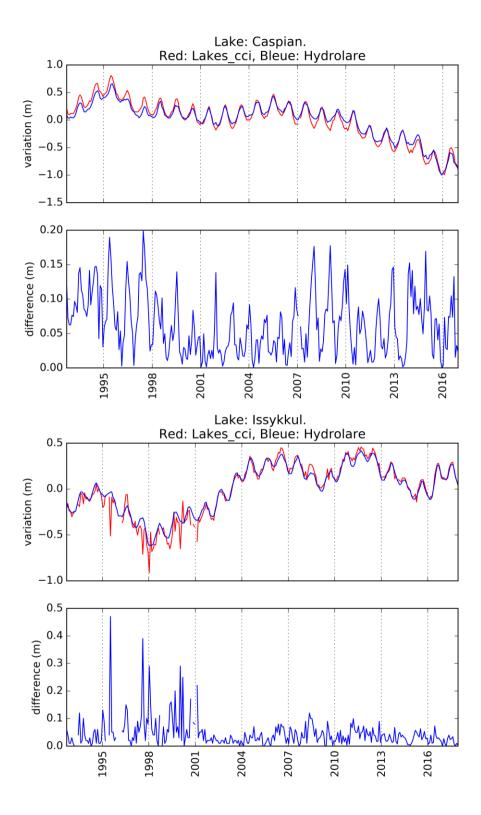
- Gons, H.J., Rijkeboer, M., & Ruddick, K.G. (2005). Effect of a waveband shift on chlorophyll retrieval from MERIS imagery of inland and coastal waters. Journal of Plankton Research, 27, 125-127
- Gurlin, D., Gitelson, A.A. and Moses, W.J. (2011). Remote estimation of chl-a concentration in turbid productive waters—Return to a simple two-band NIR-red model? Remote Sensing of Environment, 115(12), 3479-3490.
- Hu, C., Chen, Z., Clayton, T.D., Swarzenski, P., Brock, J.C. and Muller-Karger, F.E. (2004). Assessment of estuarine water-quality indicators using MODIS medium-resolution bands: Initial results from Tampa Bay, FL. Remote Sensing of Environment, 93(3), 423-441.
- Letelier, R.M. and Abbott, M.R. (1996). An analysis of chlorophyll fluorescence algorithms for the Moderate Resolution Imaging Spectrometer (MODIS). Remote Sensing of Environment, 58(2), 215-223.
- Maritorena, S., Siegel, D.A. and Peterson, A.R. (2002). Optimization of a semianalytical ocean color model for global-scale applications. Applied Optics, 41(15), 2705-2714.
- Miller, R.L. and McKee, B.A. (2004). Using MODIS Terra 250 m imagery to map concentrations of total suspended matter in coastal waters. Remote Sensing of Environment, 93(1-2), 259-266.
- Mishra, S., Mishra, D.R., & Lee, Z. (2013a). Bio-optical inversion in highly turbid and cyanobacteria-dominated waters. IEEE Transactions on Geoscience Remote Sensing, 52, 375-388
- Mishra, S., Mishra, D.R., Lee, Z., & Tucker, C.S. (2013b). Quantifying cyanobacterial phycocyanin concentration in turbid productive waters: A quasi-analytical approach. Remote Sensing of Environment, 133, 141-151
- MODIS Characterization Support Team., "MODIS Level 1B Product User's Guide- For Level 1B Version 6.2.2 (Terra) and Version 6.2.1 (Aqua)," (2017) https://mcst.gsfc.nasa.gov/sites/default/files/file attachments/M1054E PUG 2017 0901 V6.2.2 T erra V6.2.1 Aqua.pdf
- Nechad, B., Dogliotti, A., Ruddick, K., & Doxaran, D. (2016). Particulate backscattering and suspended matter concentration retrieval from remote-sensed turbidity in various coastal and riverine turbid waters. In, Proceedings of ESA living planet symposium, Prague (pp. 9-13)
- Nechad, B., Ruddick, K., & Park, Y.J.R.S.o.E. (2010). Calibration and validation of a generic multisensor algorithm for mapping of total suspended matter in turbid waters, 114, 854-866
- Neil, C., Spyrakos, E., Hunter, P.D., & Tyler, A.N. (2019). A global approach for chlorophyll-a retrieval across optically complex inland waters based on optical water types. Remote Sensing of Environment, 229, 159-178
- O'Reilly, J.E., Maritorena, S., Mitchell, B.G., Siegel, D.A., Carder, K.L., Garver, S.A., Kahru, M., & McClain, C. (1998). Ocean color chlorophyll algorithms for SeaWiFS. Journal of Geophysical Research: Oceans, 103, 24937-24953
- Ondrusek, M., Stengel, E., Kinkade, C.S., Vogel, R.L., Keegstra, P., Hunter, C. and Kim, C. (2012). The development of a new optical total suspended matter algorithm for the Chesapeake Bay. Remote Sensing of Environment, 119243-254.

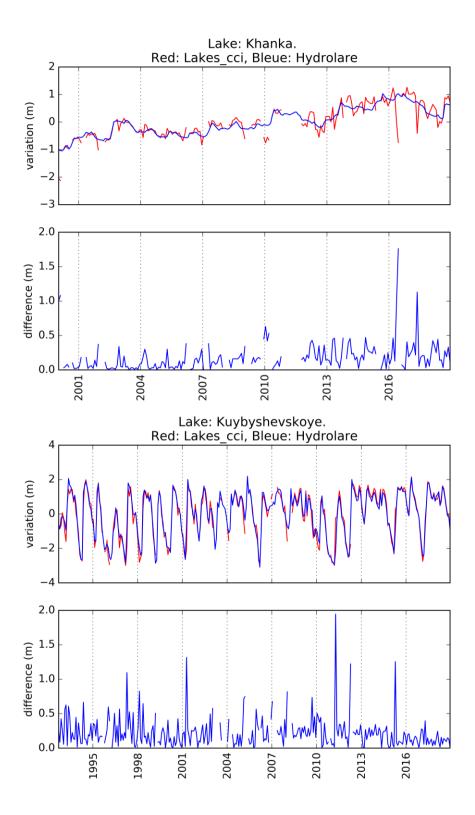
- Petus, C., Chust, G., Gohin, F., Doxaran, D., Froidefond, J.-M. and Sagarminaga, Y. (2010). Estimating turbidity and total suspended matter in the Adour River plume (South Bay of Biscay) using MODIS 250-m imagery. Continental Shelf Research, 30(5), 379-392.
- Riggs, G. A., Hall, D. K., & Román. M.O. (2019). MODIS Snow Products Collection 6.1 User Guide.
- Shi, K., Zhang, Y., Xu, H., Zhu, G., Qin, B., Huang, C., Liu, X., Zhou, Y. and Lv, H. (2015a). Long-Term Satellite Observations of Microcystin Concentrations in Lake Taihu during Cyanobacterial Bloom Periods. Environ Science Technology, http://www.ncbi.nlm.nih.gov/pubmed/25936388.
- Shi, K., Zhang, Y., Zhu, G., Liu, X., Zhou, Y., Xu, H., Qin, B., Liu, G. and Li, Y. (2015b). Long-term remote monitoring of total suspended matter concentration in Lake Taihu using 250 m MODIS-Aqua data. Remote Sensing of Environment, 16443-56.
- Sipelgas, L., Raudsepp, U. and Kõuts, T. (2006). Operational monitoring of suspended matter distribution using MODIS images and numerical modelling. Advances in Space Research 38(10), 2182-2188.
- Spyrakos, E., O'Donnell, R., Hunter, P.D., Miller, C., Scott, M., Simis, S.G., Neil, C., Barbosa, C.C., Binding, C.E., & Bradt, S. (2018). Optical types of inland and coastal waters. Limnology Oceanography, 63, 846-870
- Saunders, P.M. (1967) The temperature at the ocean-air interface. Journal of the Atmospheric Science, 24. pp. 269-273. doi:0.1175/1520-0469(1967)024<0269:TTATOA>2.0.CO;2
- Steinmetz, F., Deschamps, P.-Y., & Ramon, D. (2011). Atmospheric correction in presence of sun glint: application to MERIS. Optics Express, 19, 9783-9800
- Vantrepotte, V., Loisel, H., Mériaux, X., Neukermans, G., Dessailly, D., Jamet, C., Gensac, E., & Gardel, A. (2011). Seasonal and inter-annual (2002-2010) variability of the suspended particulate matter as retrieved from satellite ocean color sensor over the French Guiana coastal waters. Journal of Coastal Research, 1750-1754
- Vermote, E. F., Roger, J. C., & Ray, J. P. (2015). MODIS Surface Reflectance User's Guide Collection 6. Maryland: MODIS Land Surface Reflectance Science Computing Facility.
- Wang, M., Nim, C.J., Son, S. and Shi, W. (2012). Characterization of turbidity in Florida's Lake Okeechobee and Caloosahatchee and St. Lucie estuaries using MODIS-Aqua measurements. Water research, 46(16), 5410-5422.
- Wu, Y., Duguay, C.R. & Xu, L. (2021). Assessment of machine learning classifiers for global lake ice cover mapping from MODIS TOA reflectance data. Remote Sensing of Environment, 253, 112206, https://doi.org/10.1016/j.rse.2020.112206.
- Zhang, M., Tang, J., Dong, Q., Song, Q. and Ding, J. (2010). Retrieval of total suspended matter concentration in the Yellow and East China Seas from MODIS imagery. Remote Sensing of Environment, 114(2), 392-403.
- Zhang, Y., Shi, K., Liu, X., Zhou, Y., & Qin, B. (2014). Lake topography and wind waves determining seasonal-spatial dynamics of total suspended matter in turbid Lake Taihu, China: assessment using long-term high-resolution MERIS data. PLoS One, 9
- Zhao, H., Chen, Q., Walker, N.D., Zheng, Q. and MacIntyre, H.L. (2011). A study of sediment transport in a shallow estuary using MODIS imagery and particle tracking simulation. International Journal of Remote Sensing, 32(21), 6653-6671.
- Leonard J., 1974: Aperçu de la végétation de la partie Est du lac Tchad. ORSTOM

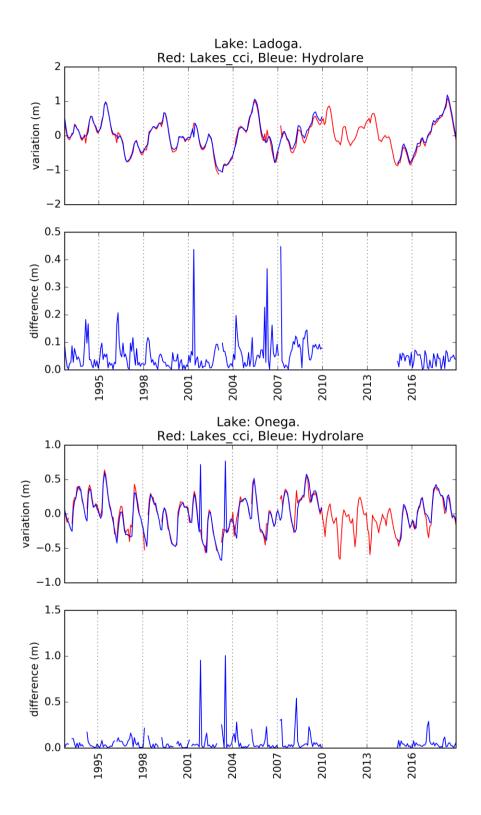
Iltis André and Jacques Lemoalle, 1983: the aquatic vegetation of lake Chad, in lake Chad, Ecology and productivity of a shallow tropical ecosystem. JP Carmouze, Durand JR and C Leveque edts. Junk Publishers.

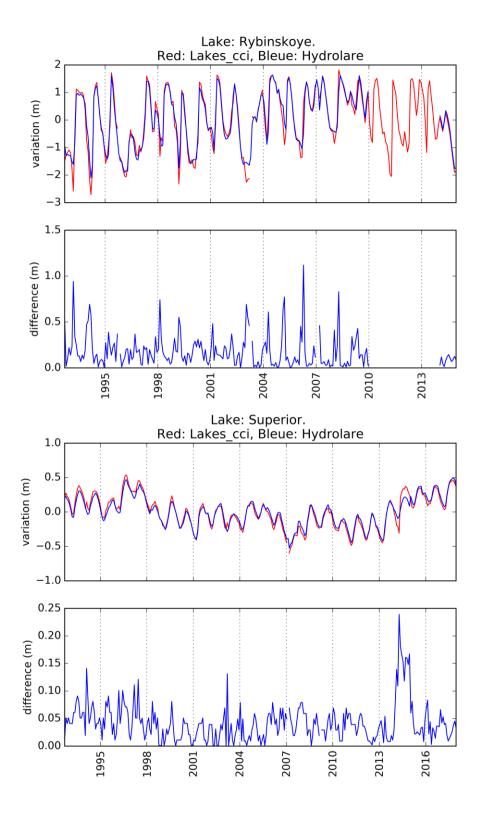
Appendix A. Hydrolare LWL Comparison











Appendix B. Water Office Canada Comparison

